# DEVELOPMENT OF A COMPUTER SIMULATION MODEL OF LARGEMOUTH BASS POPULATION DYNAMICS 

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# DEVELOPMENT OF A COMPUTER SIMULATION MODEL OF LARGEMOUTH BASS POPULATION DYNAMICS 

Thesis Approved:


Many aspects of the population dynamics of largemouth bass have been investigated in Lake Car1 Blackwell by the Oklahoma Cooperative Fishery Research Unit. The objective of this study was to incorporate these research findings into a computer simulation model of population dynamics. Funds were provided by Federal Aid to Fish and Wildife Restoration, Oklahoma D-J Project F-36-R, Job 3.

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## CHAPTER I

INTRODUCTION

The largemouth bass, Micropterus salmoides (Lacépède), is one of the most important sport fishes of warmwater lakes and reservoirs. Many anglers fish exclusively for largemouth bass, others fish for it only incidently, but most consider this species an outstanding sport fish. Horvath (1974) reported that about 24 percent of the fishing trips on reservoirs in the southeastern states are specifically for bass and another 18 percent for bass plus other species (Jenkins 1975). The largemouth bass is also an economically important species because these anglers spend a considerable amount of money on boats, motors, gasoline, bait, tackle, and licenses; therefore, waters which support healthy populations of largemouth bass become valuable natural resources.

Careful management of natural resources is necessary to ensure proper protection and preservation while allowing some type of utilization by society. In the case of largemouth bass, we would like to regulate the fisherman's harvest to ensure that the resource is not depleted. To do this we must understand how regulations and environmental factors influence population processes and ultimately yield. We must also understand the ecological role (or niche) of the largemouth bass in the fish community. In multispecies fisheries largemouth bass serve a dual function. They regulate the size of sunfish populations,
thereby allowing these fish to grow to a larger size, and they convert the biomass of many non-game fish (e.g., shad, Dorosoma spp.) to a more usable and aesthetic form.

The management of largemouth bass fisheries in large reservoirs (greater than 200 hectares) is often difficult because reservoir size may prohibit the fishery manager from collecting enough data to propose rational management strategies. Also, most of these reservoirs have unstable water levels which prevent the reservoir system fish population from attaining a stable (or steady) state. Management of largemouth bass fisheries in large reservoirs must thus rely on very few experimental studies on the response of the fish populations to the implementation of various management strategies. What is needed is a reliable method for predicting the consequences of a proposed management decision prior to implementation.

One approach that has been utilized in the management of other sport and commercial fisheries is the use of systems analysis, computer modeling and simulation techniques to develop a model to aid in management. The objective of this study is to develop a computer simulation model which will predict year-class strength, production and yield for the largemouth bass population of Lake Carl Blackwell. The long-range objective of this type of research is to develop a largemouth bass management model that will provide biologists with a useful tool for optimizing the yield from the fishery.

## LITERATURE REVIEW

Models in General

The term, model, can be defined generally as any physical or abstract representation of a real system. Models may be categorized as mental, verbal, graphical, physical, or mathematical. Each of us has a mental image of how a pair of guppies in an aquarium will increase in numbers, slowly at first and then more rapidly as more individuals become sexually mature, until the population finally stabilizes at a certain level. Our mental image, when put into words, becomes a verbal model which can be more vividly expressed by means of a graphical model (Figure 1). The aquarium which the guppies occupied was a physical representation of the real system (i.e., a tropical aquatic ecosystem) which the guppies normally inhabit. Physical models such as this one are useful in that many variables are controlled allowing us to study the effect of only a few. Mathematical models are the most rigorous type of models and permit us to say precisely how the components of our simplified system are related. The rate of change in the number of guppies in our aquarium at any instant is described by a mathematical model:

$$
\begin{equation*}
\frac{\mathrm{dN}}{\mathrm{dt}}=r\left[\frac{\mathrm{~K}-\mathrm{N}}{\mathrm{~K}}\right] \mathrm{N} \tag{2.1}
\end{equation*}
$$



Figure 1. Graphical model of the increase in a population of guppies over time.
where $N=$ number of guppies, $K=$ number of guppies at the stabilization level (asymptote), $r=$ intrinsic rate of population increase, and $t=$ time. Integrating (2.1) we get an equation for the S-shaped logistic curve for population growth in Figure 1:

$$
\begin{equation*}
N(t)=\frac{K}{1+N(0) e^{-r t}} \tag{2.2}
\end{equation*}
$$

This is a simple mathematical model which relates only two components of the system, i.e., rate of population change and population density.

More complex mathematical models consist of (i) system (or state) variables, (ii) transfer functions, (iii) forcing functions, and (iv) parameters (Walters 1971). System variables are sets of numbers used to represent the state of the system at a given time. One or more system variables are used to characterize a particular component of the system. Transfer functions are equations which represent flows or interactions between components, forcing functions are equations which represent inputs to the system, and parameters are constants of the mathematical equations. Depending on the description of the parameters and the form of the transfer functions and forcing functions, a model may be deterministic or stochastic. The deterministic model tells us that for given values of the independent variables we should expect the dependent variables to have a single corresponding value. The possibility of chance variation is ignored. Stochastic models attempt to include the effects of random variability so that for any given values of the independent variables we might expect the dependent variable to have a series of values, each with an associated probability.

All of the models referred to hereinafter are similar in that they are mathematical, but they differ in the level of hierarchial organiza-
tion of the system's components and the choice of system variables. The components and system variables used in the development of a particular model depend on the fishery under study, the amount and type of available data, and the questions the model is intended to answer.

Models may be evaluated in terms of their resolution, realism, precision, and generality (Holling 1966a). Resolution is a depth criterion related to the number of essential features in the real system that the model is intended to mimic. A model that includes only a few simple components is said to have low resolution and, conversely, if it includes many of the attributes of the system it is said to have high resolution. Realism refers to the degree to which the mathematical equations correspond to the biological processes which they describe. A model which predicts the growth rate of a fish simply on the basis of its age ignores the true components of the growth process, i.e., feeding energetics, and loses a degree of realism. The third criterion, precision, is concerned with the ability of the model to generate values for a component that compare with the values observed in the real system. Generality is a breadth criterion related to the ability of the model to work in a variety of real world systems. Holling (1966b) has shown that by dividing relevant components into basic (universal) and subsidiary (sporadic) components, generality becomes theoretically possible. Components shared by all examples are called basic in that they underlie all manifestations of the process. Those that are present in only some situations are called subsidiary.

## Classical Models of Fisheries

Mathematical models of fishery systems have traditionally been used for fish stock assessment and prediction of maximum sustained yield. Baranov (1918) was the first to develop a theoretical model of an exploited fish population. The dynamics of this model were governed by recruitment, growth, natural mortality and fishing mortality. The total biomass of usable stock (P), i.e., fishes large enough to be harvested, was increased by the recruitment of new individuals to the usable stock and by growth of individuals and decreased by both natural and fishing mortality. A general model can be constructed expressing the relative rate of change in biomass of the usable stock in these terms:

$$
\begin{equation*}
\frac{d P}{P d t}=R(P)+G(P)-M(P)-F(E)+e, \tag{2.3}
\end{equation*}
$$

where $R, G$, and $M=$ rates of recruitment, growth, and natural mortality, respectively, and are functions of the biomass of usable stock ( $P$ ) and its age composition (Beverton and Holt 1963). F = fishing mortality and is a function of fishing effort (E), and e is a variable rate of change in the biomass due to environmental factors. In the steady state, with population in equilibrium under average environmental condition, $\frac{d P}{d t}=0$ and $e=0$, so that

$$
\begin{equation*}
F(E)=R(P)+G(P)-M(P), \tag{2.4}
\end{equation*}
$$

and the equilibrium harvest, $\mathrm{Y}=\mathrm{F}(\mathrm{E}) \mathrm{P}$, will equal the additions due to recruitment and growth minus the loss due to natural mortality, i.e.,

$$
\begin{equation*}
Y=F(E) P=[R(P)+G(P)-M(P)] P . \tag{2.5}
\end{equation*}
$$

Ricker (1975) discussed various methods to compute equilibrium harvest and Paulik and Bayliff (1967) have developed a computer program for

Ricker's method.
The two general approaches most often used to predict the yield of exploited fish populations are (i) dynamic pool models and (ii) logistic models. Dynamic pool models are the most widely employed. In these models the elemental rates of recruitment, growth and natural mortality are estimated separately and combined into an appropriate form of the general model (2.5) assuming a steady state. These models, elaborated by Beverton and Holt (1957), are especially applicable to fisheries where one may regulate both fishing effort and minimum size of capture.

The other approach, which has been developed most completely by Schaefer (1954, 1957), was modified by Fox (1970) and Pella and Tomlinson (1969), and reviewed by Silliman (1971). It involves combining the rates of recruitment, growth and natural mortality into a single function of the biomass of usable stock ( P ). Models of this type, called logistic or surplus production models, are useful in that the only data needed are total catch, total effort and the instantaneous rate of fishing mortality. However, predictive reliability of this model is not very dependable due to the inherent assumptions (Watt 1956).

The basic weakness of these existing mathematical models is that they are deterministic and assume a steady-state fishery, i.e., one in which recruitment, growth and natural mortality are constant from year to year. This assumption may not be too unreasonable when dealing with a large marine fishery but in general, the smaller the fishery, the more chance there is that results predicted by a deterministic model will not match the actual results. To avoid this limitation

Watt (1956) proposed a model which would include the influence of environmental factors on recruitment, growth and natural mortality, and has applied this model to a sport fishery for smallmouth bass in South Bay of Lake Huron (Watt 1959). However, this type of model will work only where fishing intensities have covered a wide range of values, and a great amount and variety of population data are available.

Systems Analysis, Computer Modeling and Simulation Applied to Fisheries

The availability of electronic digital computers has enhanced the growth and development of new quantitative techniques, such as systems analysis, computer modeling and simulation. Considerable progress has also been made in the application of these techniques to ecology (Patten 1971, 1972, 1975a; Watt 1966, 1968) and fisheries science (Saila 1972).

Systems analysis involves determining which variables are most important in regulating the system, and incorporating these variables into a mathematical systems model. Computer implementation and the concurrent ease of bookkeeping and computation has allowed these models to become more complex and include more of the relevant variables than was previously possible.

Once the mathematical model has been formulated and programmed for the computer the behavior of the system can be simulated. Hence, computer simulation models have evolved. Simulation can also be used for determining parameter values by varying input values until simulated results agree with observed data. Sensitivity analysis involves simulation using variations in input variables and parameters to determine
the magnitude of input effect on system behavior. Validation of the model involves testing to see if the model adequately predicts observed system behavior. Another technique of systems analysis involves optimization of an objective function by manipulation of control variables (Farrell et al. 1975). Saila and Hess (1975) have applied optimization techniques to fisheries management using maximum biomass yield as the objective function and rate of fishing as the control variable for the Beverton-Holt and Schaefer models.

Paulik $(1969,1972)$ has reviewed the literature on computer simulation models in fisheries research, management and teaching and has predicted that the resource management agency of the future will maintain a hierarchy of simulation models to serve as basic planning tools for studying system response to natural and artificial change. Lackey (1975) also foresees a much closer involvement between modelers and decision-makers in natural resource management. Simulation in fisheries is commonly used to evaluate costs and benefits of management strategies and to learn basic system properties, especially ecological properties.

Two main tactical approaches to development of simulation models in fisheries and ecological systems can be categorized as the "experimental components" (Holling 1963, 1966a, 1966b) and the "compartmental system" (Patten 1971) approaches. The experimental components approach emphasizes a detailed analysis of ecological processes by breaking them down into simple subprocesses or experimental components. This approach would seem well suited for a model of population dynamics in which the processes of growth, mortality, reproduction, predation, and competition would be analyzed as subprocesses. The compartmental
system approach emphasizes the quantities of energy and materials in ecosystem compartments. Complex processes associated with populations making up the compartments are assumed to counter-balance one another resulting in simple behavior of the compartment as a whole. Models developed using the experimental components approach have tended to be realistic and precise and those using the compartmental system approach have tended to be general, but not realistic (Walters 1971).

## Current Simulation Models in Use

Simulation models for many of the important commercial fisheries have been developed using a detailed analysis of the population processes. Most of these have been developed for a specific fishery (Francis 1974; Jensen 1975; Jones and Hall 1973; Larkin and Hourston 1964; Larkin and McDonald 1968; Paulik and Greenough 1966) but a few models are available that are generally applicable (Silliman 1966, 1969; Walters 1969).

The development of simulation models for inland recreational fisheries has encountered many difficulties. First, there is relatively little data available for these fisheries when compared to commercial marine fisheries and second, the dynamic pool and logistic models are inadequate in describing multispecies fisheries in which a steady state cannot be safely assumed. Some progress has been initiated toward simulating multispecies centrarchid fisheries (Zuboy and Lackey 1975) and put-and-take trout fisheries (Hammond and Lackey 1976), providing a foundation on which to build more complex models. Another promising approach has been to model fish biomass dynamics by analyzing the ecological processes involved (Hackney and Minns 1974;

Kitchell et a1. 1974).
Recent authors (Dickie 1973; Lackey 1975; Patten 1969; Regier and Henderson 1973; Schaaf 1975) have emphasized the need for a more general modeling approach directed explicitly at the ecosystem level of organization. This approach stresses the importance of the interactive system aspect of fisheries and consequently efforts would be devoted to measurement of overall system properties and proposing generalizations which would enable us to simplify the systems that must be managed. These models usually analyze the flow of energy and/or biomass through several gross compartments of the fishery system (Riffenberg 1969; Patten 1969; Walters and Efford 1972).

Computer simulation models are also valuable tools in teaching natural resources management and evaluating management strategies since they allow the student and/or manager to make and test decisions on a simulated resource and analyze their consequences almost immediately (Clark and Lackey 1975; Li and Adams 1976; Titlow and Lackey 1973, 1974).

## CHAPTER III

## DESCRIPTION OF STUDY AREA

Lake Carl Blackwell (Figure 2) is a shallow, turbid reservoir located in north-central Oklahoma, about 12.8 kilometers west of Stillwater in Payne and Noble Counties. Dam construction on Stillwater Creek, a Works Progress Administration project, began in 1936 and was completed in 1938 with the primary purpose of providing erosion control although the lake has also been used for outdoor recreation, municipal water supply, and flood control. The reservoir and some of the surrounding land was leased to Oklahoma State University in 1948 and deeded to the University in 1954. From 1950 to 1974 , it also served as the sole water supply for municipal Stillwater but with the completion of nearby Lake McMurtry, it now serves as an alternate water supply (Shirley 1975).

The original spillway elevation was 288.37 meters above mean sea level (M.S.L.) but in 1948 the spillway was reconstructed and lowered to an elevation of 287.78 meters above M.S.L. At this elevation the surface area is 1400 hectares, volume is 67.8 million cubic meters, mean depth is 4.8 meters, and the shoreline development index (S.D.I.) is 6.8. The reservoir is situated in a relatively small watershed (approximately 14 times the surface area of the lake) in a region characterized by cyclic rainfall, and thus has been subject to water level fluctuations since its impoundment (Figure 3). In October 1972 the reservoir


Figure 2. Lake Carl Blackwell and its location relative to Oklahoma and surrounding states.


Figure 3, Nater level in meters above mean sea level for Lake Carl Blackwell from impoundment to May 1977.
reached the lowest recorded elevation of 281.75 meters above M.S.L., over 6 meters below spillway elevation. At this level, the surface area was only 491.7 hectares, volume was 11.0 million cubic meters, mean depth was 2.2 meters, and the S.D.I. was 3.5. Maximum depth occurs in the old stream channel near the dam and the shallowest depths occur at the west end.

The reservoir is contained within the Redbeds Plains physiographic region, characterized by fine red soils derived from Permian clays and shales. The rolling hills surrounding Lake Carl Blackwell are partially wooded, but pastures of native grasses prevail. Wind-generated wave action resulting from the high average wind velocities of the prevailing southwest winds along with the relatively low, unprotected shoreline, shallow depth, and east-west orientation of the reservoir allow almost continuous vertical and horizontal water circulation. Consequently, temperature and dissolved oxygen curves are generally orthograde and the water remains turbid. Thermal stratification occurs occasionally during the summer months with coincidence of high ambient air temperatures and decreased wind velocities. The turbidity seems to be a result of resuspension of shallow sediment by wave action in the western end of the lake, and movement of sediment to the eastern end by wind-driven currents (Norton 1968; Hysmith 1975). In 1968 and 1969, Hysmith (1975) measured turbidities ranging from 17.0 to $109.7 \mathrm{ppm} \mathrm{SiO}_{2}$ and averaging 42.5 ppm . He was, however, unable to show that primary productivity was limited by turbidity, although Leonard (1950) felt that turbidity rather than chemical conditions was the primary factor limiting primary productivity during the first 12 years of impoundment.

Almost the entire lake is devoid of submergent and emergent
aquatic macrophytes, apparently due to turbid water conditions, unstable bottom sediments, and fluctuating water levels. Potamogeton nodosus, American pondweed; Scirpus spp., bulrushes; and Typha spp., cat-tails, do occur occasionally under stable water levels in coves protected from prevailing winds. A periodic sequence of natural drawdowns, plant succession, and flooding is a recurring phenomenon of the lake, as was noted during the first 12 years following impoundment (Loomis 1951; de Gruchy 1952). Cyperus spp., sedges; Amannia coccinea, scarlet amannia; and Polygonum spp., smartweeds, are the predominant terrestrial macrophytes that follow the receding water line (de Gruchy 1952).

The following fish species are known to occur in Lake Carl Blackwe11, and are listed in order of decreasing relative abundance based on cove rotenone samples taken from 1966 to 1975.

## Scientific Name

1) Dorosoma cepedianum (LeSueur)
2) Lepomis macrochirus Rafinesque
3) Pomoxis annularis Rafinesque
4) Lepomis megalotis Cope
5) Aplodinotus grunniens Rafinesque
6) Lepomis humilis (Girard)
7) Lepomis cyanellus Rafinesque
8) Micropterus salmoides (Lacepede)
9) Ictalurus punctatus (Rafinesque)
10) Cyprinus carpio Linnaeus
11) Carpiodes carpio (Rafinesque)
12) Morone chrysops (Rafinesque)
13) Pylodictis olivaris (Rafinesque)

Common Name
Gizzard shad
B1uegill
White crappie
Longear sunfish
Freshwater drum
Orangespotted sunfish
Green sunfish
Largemouth bass
Channel catfish
Carp
River carpsucker
White bass
Flathead catfish
14) Pimephales spp.

Minnows
15) Ictalurus melas (Rafinesque)
16) Notropis lutrensis (Baird and Girard) Red shiner
17) Notemigonus chrysoleucas (Rafinesque) Golden shiner
18) Gambusia affinis (Baird and Girard) Mosquitofish
19) Lepomis microlophus (Günther) Redear sunfish

In addition to these species, a few black crappie, Pomoxis nigromaculatus (LeSueur), were collected in 1973 and 1974. Loomis (1951) reported black crappie to be the fourth most abundant fish species in the lake. Walleye fry, Stizostedion vitreum vitreum (Mitchil1), were stocked in 1969, 1970, and 1971, and northern pike, Esox lucius Linnaeus, in 1968, but there was no evidence of natural reproduction (Johnson 1974). The fish population is unusual for an Oklahoma reservoir in that the gars (Lepisosteus spp.) and the buffalofishes (Ictiobus spp.) are absent.

The fishery of Lake Carl Blackwell is concentrated on channel catfish, largemouth bass, white crappie, and white bass. Based on a creel survey conducted in $1969,61.8 \%$ of the anglers were fishing for channel catfish, and $13.8 \%, 11.1 \%$, and $10.9 \%$ were fishing for largemouth bass, white crappie, and white bass, respectively (Zweiacker 1972).

Lake Carl Blackwe11 was chosen for the study area since several Investigations have been made on aspects of the ecology of largemouth bass, including population dynamics of adults (Zweiacker 1972), growth, production, and mortality of young-of-the-year (Shirley 1975), growth in relation to water level (Zweiacker et al. 1973), and the relationships between weather and other environmental factors and year-class strength (Summerfelt 1975; Summerfelt and Shirley 1976).

## CHAPTER IV

## MODELING PROCEDURE

Population dynamics of largemouth bass in reservoir environments are very complex and may best be studied in terms of Holling's experimental components approach with emphasis on the processes of growth, mortality, reproduction and year-class formation. Each of these processes is influenced by the life history stage or age of the fish, many density dependent and density independent factors, and the season of the year. A population dynamics model must reflect these ecological processes if it is to be a realistic representation. Since the model will be intended for use in evaluating management strategies, considerable flexibility of input requirements is needed because the same amount and type of data will not be available for all reservoir bass fisheries.

The first step in developing a model of reservoir bass populations is to determine which components of the reservoir ecosystem are relevant to the analysis of these ecological processes. Initial analysis of each process involves construction of box-arrow diagrams to indicate paths of cause-and-effect relationships (Figures 4, 5, and 6). Development of these diagrams was the result of review of the literature on these topics: The purpose of these figures is to provide a reference point and a guide for modeling and to aid in conceptualization of the interrelationships.


Figure 4. Factors influencing the reproductive process and yearclass formation in largemouth bass.


Figure 5. Factors influencing the growth process of largemouth bass. Solid lines and broken lines indicate direct and indirect influences, respectively.


Figure 6. Factors influencing mortality of largemouth bass. Solid lines and broken lines indicate direct and indirect influences, respectively.

Selection and definition of system variables and parameters is the next step in the modeling procedure. The philosophy employed in this study was to begin with a simple model (Model I) which included only a few system variables and parameters and to expand and modify this model so that it would include more relevant components.

Construction of the mathematical model is the third and most rigorous task and involves specifying the form of the transfer functions, forcing functions and estimation of parameter values. The relationships between population size, growth, recruitment, and survival rates, and the relation of these factors with environmental factors must also be determined and quantified. Data collected from Lake Carl Blackwell were analyzed by simple linear and multiple linear regression techniques, described by Draper and Smith (1966), to arrive at the mathematical equations.

After the forms of the equations were specified, the mathematical model was adapted for computer simulation using FORTRAN IV programming language. For each model a computer program was written with considerable flexibility of input requirements to allow for manipulation of the simulated fish population by varying the input data. The FORTRAN language was chosen because it is generally available on most computer systems and most recently-trained fishery biologists have had some exposure to it. Programs were run on the IBM System 370/Model 158 digital computer at the Oklahoma State University Computer Center.

## CHAPTER V

MODEL I

## Model Description

Model I was age-structured, utilized age-specific fecundities and survival rates and was similar to the Leslie matrix algorithm (Leslie 1945) since fecundity, vulnerability to predators, and susceptability to angling change as a fish grows older, and since a new cohort is added to the population each year. The notation used is as follows: $N_{i}(t)=$ number of individuals of age $i$ at time $t, m_{i}=$ fecundity (number of eggs) per individual of age $i, S_{i}=$ probability that an individual of age $i$ will survive to age $i+1$. Fecundity per individual, $m_{i}$, would equal fecundity per female times 0.5 , assuming a $1: 1$ sex ratio. The basic time unit is a year which commences at the time eggs are laid (approximately 15 May for Lake Carl Blackwe11). The number of eggs produced is calculated by

$$
\begin{equation*}
N_{o}(t)=\sum_{i} N_{i}(t) m_{i} \tag{5.1}
\end{equation*}
$$

and a new age distribution is obtained by

$$
\begin{equation*}
N_{i}(t+1)=N_{i-1}(t) S_{i-1} \tag{5.2}
\end{equation*}
$$

for all $i=1,2,3, \ldots, k$, where $k=$ maximum age.

Reliable estimates of $S_{o}$, survival from egg stage to age $I$, are difficult to obtain for natural populations. For this reason, $\mathrm{S}_{\mathrm{o}}$ is estimated indirectly assuming an equilibrium population and using age-
specific fecundity and survival data. Vaughan and Saila (1976) derive this estimation procedure based on the Leslie matrix algorithm.

$$
\begin{equation*}
S_{o}=\frac{1}{\sum_{i=1}^{k-1}\left[m_{i+1}\left(\prod_{j=1}^{i} S_{j}\right)\right]} \tag{5.3}
\end{equation*}
$$

A program 1isting is given in Appendix A and a sample output is in Appendix $B$.

## Results and Discussion

Simulation runs were made using the average age-specific survival rates from Zweiacker et al. (1973) and the average age-specific fecundities from Kelley (1962) which appear in Table 1. Figure 7 illustrates the results of a simulation run starting with 1000 age I fish. The simulated population initially oscillated due to the timelag required for the fish to reach maturity and finally stabilized at about simulation year 18 .

Sensitivity analysis was performed to determine how the population size after a 20 -year simulation was affected by varying the input parameters. Net sensitivity of the population size to a $10 \%$ change in any given input parameter was computed according to the formula given by Francis (1974):

$$
\begin{equation*}
\mathrm{S}(\mathrm{x}, \mathrm{y}, \Delta \mathrm{x})=\frac{\mathrm{y}(\mathrm{x}+\Delta \mathrm{x})-\mathrm{y}(\mathrm{x})}{\mathrm{y}(\mathrm{x})} \tag{5.4}
\end{equation*}
$$

where $S(x, y, \Delta x)$ is the net sensitivity of $y$ to a change, $\Delta$, in $x$. The relative sensitivity was then obtained by dividing net sensitivity by the largest net sensitivity value.

Initial age structure and values of population parameters (Table 1)

Table 1. Initial age structure, age-specific fecundity and survival rates used in nominal simulation of Model I.

| Age <br> $(i)$ | Numbers <br> $\left(\mathrm{N}_{\mathrm{i}}\right)$ | Fecundity <br> $\left(\mathrm{m}_{\mathrm{i}}\right)$ | Survival <br> $\left(\mathrm{S}_{\mathrm{i}}\right)$ |
| :---: | :---: | :---: | :---: |
| 0 | - | - | 0.00015 |
| 1 | 787 | 0 | 0.676 |
| 2 | 528 | 0 | 0.616 |
| 3 | 322 | 9335 | 0.659 |
| 4 | 210 | 4350 | 0.560 |
| 5 | 116 | 5750 | 0.375 |
| 6 | 43 | 13610 | 0.197 |
| 7 | 8 | 13610 | 0.071 |
| 8 | 1 | 13610 | 0.000 |



Figure 7. Model I simulation of population changes starting with 1000 age I fish.
used in simulation resulted in a total population size of 2396 after a 20-year simulation. The effect of varying the survival rate of age 0 fish is shown in Figure 8. Results of the sensitivity analysis (Table 2) indicate that the total population size is most responsive to changes in survival rates of age 0 , age $I$, and age II fish, respectively. Therefore, based on this model, it is most important that we have accurate estimates of survival rates of these age groups in order to simulate population trends. Horst (1977) also found that population growth rate was most sensitive to changes in survivorship of younger ages from sensitivity analysis of the Leslie matrix model.

The next step in any modeling problem is to analyze the assumptions on which the initial model is based. In Model I, I assumed that the population operated in a deterministic fashion with constant agespecific survival rates and fecundities. Thus, simulated population trends beginning with 1000 age I fish (Figure 7) do not mimic the situation encountered in new reservoirs where in the first years of impoundment, large year-classes of bass are produced and the population exhibits a "boom and bust" phenomenon. Also the effects of density and environmental factors are ignored in Model I. Since this assumption is unrealistic, further developments of this model will involve varying the age-specific survival rates and/or fecundity based on density or environmental factors. Also there is evidence for differential mortality of older male bass (Bryant and Houser 1971; Hubert 1976) which tends to shift the sex ratio away from unity. In many reservoirs this shift may be negligible but inclusion of a parameter in the model to account for this variation would increase the model's flexibility.


Figure 8. Model I simulation results with nominal and adiusted values for the survival rate of ape 0 fish, $S_{0}$.

Table 2. Total population sizes (N) and resulting sensitivities obtained after a 20-year simulation of Model I with adjusted input parameters.

| Parameter | +10\% |  |  | -10\% |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | N | Net sensitivity | Relative sensitivity | N | Net sensitivity | Relative sensitivity |
| So | 3990 | +0.6653 | 1.000 | 1374 | -0.4265 | 1.000 |
| $\mathrm{S}_{1}$ | 3928 | +0.6394 | 0.961 | 1388 | -0.4207 | 0.986 |
| $\mathrm{S}_{2}$ | 3935 | +0.6423 | 0.965 | 1390 | -0.4199 | 0.984 |
| $\mathrm{S}_{3}$ | 2981 | +0.2442 | 0.367 | 1891 | -0.2108 | 0.494 |
| $\mathrm{S}_{4}$ | 2735 | +0.1415 | 0.213 | 2079 | -0.1323 | 0.310 |
| $\mathrm{S}_{5}$ | 2568 | +0.0718 | 0.108 | 2232 | -0.0684 | 0.160 |
| $S_{6}$ | 2421 | +0.0104 | 0.016 | 2366 | -0.0125 | 0.029 |
| $\mathrm{S}_{7}$ | 2397 | +0.0004 | 0.001 | 2396 | 0.0 | 0.0 |
| $\mathrm{m}_{3}$ | 3221 | +0.3443 | 0.518 | 1762 | -0.2646 | 0.620 |
| $\mathrm{m}_{4}$ | 2624 | +0.0952 | 0.143 | 2184 | -0.0885 | 0.207 |
| $\mathrm{m}_{5}$ | 2556 | +0.0668 | 0.100 | 2236 | -0.0668 | 0.156 |
| $\mathrm{m}_{6}$ | 2539 | +0.0597 | 0.090 | 2256 | -0.0584 | 0.137 |
| $\mathrm{m}_{7}$ | 2424 | +0.0117 | 0.018 | 2369 | -0.0113 | 0.026 |
| $\mathrm{m}_{8}$ | 2397 | +0.0004 | 0.001 | 2396 | 0.0 | 0.0 |

## CHAPTER VI

MODEL II

## Introduction

Sensitivity analysis of Model I showed that the population size was most sensitive to changes in the survival rate from egg stage to age I. This stage is also the one at which natural mortality of largemouth bass is the greatest. Summerfelt and Shirley (1975) found that in Lake Carl Blackwell, the 1973 year-class, a large year-class, suffered $95 \%$ mortality from the time of hatching (5 May) until 1 October of their first growing season, and $66 \%$ of that mortality had occurred during the first 40 days after hatching. The authors inferred that wave action was the major limiting factor during this period. Kramer and Smith (1962) also considered wind the single most important factor in year-class formation in Lake George, Minnesota. Summerfelt (1975) has also found that in Lake Car1 Blackwe11 year-class strength is determined by events occurring during the first few weeks of fish life; passage of frontal systems associated with strong winds and cooler temperatures apparently disrupt spawning and result in increased mortality of bass embryos and larvae. Conversely, spawning success was greatest during short intervals when weather was stable. Eipper (1975) has concluded that generally year-class strength fluctuation is the result of the very high mortality during the period between egg fertilization and the end of the first few weeks of life. He also concluded
that strong winds and the various indirect influences of low temperature probably are most responsible for mortality during this period.

In many reservoirs large year-classes of largemouth bass are produced in years of stable or rising water levels during spawning (Aggus and Elliot 1975; Bross 1969; Keith 1975; von Geldern 1971). It is postulated that increasing water levels favor the survival of young-of-the-year largemouth bass by the flooding of shoreline areas containing terrestrial vegetation, which increases cover for nest sites and for shelter from predation and releases nutrients into the littoral zone thereby promoting production of food for the young bass (Shirley 1975). Also, the increased depth of water over the nests decreases the effects of wind, wave action and temperature fluctuation (Kramer and Smith 1962).

The relation between environmental factors and year-class strength of largemouth bass in Lake Carl Blackwell was studied by Summerfelt and Shirley (1976) by correlating these factors with the estimated ecological density of 11 consecutive year-classes (1965-1975). Cove poisonings with rotenone were used to make late summer estimates of numerical density of young-of-the-year (YOY) bass and these estimates were adjusted to a constant date (13 August) using an estimated daily instantaneous mortality rate of 0.0015 . Year-class strength was estimated in this way for $1966,1967,1968,1971,1973,1974$, and 1975. The cove rotenone samples probably reflect the ecological density (number per unit of acceptable habitat) of Odum (1971:163-166) assuming that YOY bass are largely limited to the littoral zone. The 1965 and 1970 estimates were back-calculated from estimates of number of age I bass in the 1966 and 1971 cove poisoning collections using the daily
instantaneous mortality rate of 0.0015 . The 1972 year-class was estimated during the fall of 1972 by the mark-recapture technique developed by Lewis et al. (1963) in which bass were collected by shoreline electrofishing, marked and released for recapture in subsequent trips around the lake. This estimate was then divided by the area of water less than 2 meters deep to make it more comparable with the cove rotenone estimates. The 1969 year-class was estimated by comparing the electrofishing catch rate of that year-class with that of the 1968 year-class which had been estimated by cove poisoning. Catch rates were taken from Zweiacker (1972) and the density of the 1969 year-class was calculated by multiplying the ratio of the electrofishing catch rates. Summerfelt and Shirley (1976) discussed the comparability of the 1969 and 1972 year-class estimates with those estimated by cove poisoning.

Using their estimates of ecological density of YOY bass, Summerfelt and Shirley (1976) correlated these values with a series of biotic and abiotic environmental parameters including: water level, change in water level, pH, methyl orange alkalinity, hardness, turbidity, wind velocity, and number of spawners. Correlations were made using monthly maximum, minimum and mean and seasonal mean for each parameter except number of spawners and water levels. Correlations were also made between YOY bass density and the water level on the 1st and 15 th of each month (January - August), monthly change in water level, change in water level since the end of the previous growing season (water level on the lst and 15 th of each month minus water level on 1 October the previous fall) and the estimated number of spawners in those years when reliable mark-recapture estimates of adult bass were made.

The results of their study showed that year-class strength was positively correlated with water level, change in water level and turbidity, negatively correlated with hardness, alkalinity and pH , and uncorrelated with wind, air and water temperature, and size of the spawning population. They concluded that the fluctuations in yearclass size were due to the water level and its effect upon food and cover for YOY bass. Other significant correlations were attributed to the effects of changing water levels on the physical and chemical composition of the water.

Mode1 II represents an attempt to include the effects of environmental factors on reproduction and year-class formation within the framework of Model I. An additional cove rotenone collection was made in August 1976 resulting in 12 consecutive estimates of ecological density of YOY bass in Lake Carl Blackwell (Figure 9).

## Mode1 Description

Mean water level ${ }^{1}$ during May, water level fluctuation from 1 October of the previous fall to 15 May, and density of YOY bass (Table 3) were analyzed by regression (Draper and Smith 1966) to determine how useful these variables were in predicting year-class strength. These variables were chosen because they were most significantly correlated with year-class strength and thus probably the most meaningful.

The relationship between water level fluctuation and year-class strength and results of linear regression are illustrated in Figure 10. The correlation coefficient of 0.8779 was highly significant ( $\mathrm{P}=0.0002$ )

[^0]

Figure 9. Estimated ecological density of young-of-the-vear largemouth bass on 13 August in Lake Carl Blackwell, Oklahoma (1965-1976).

Table 3. Estimated ecological density of young-of-the-year largemouth bass, water level during spawning, and water level fluctuation in Lake Car1 Blackwel1, Oklahoma (1965-1976).

| Year class | $\begin{gathered} \text { Estimated } \\ \text { density } \\ \text { (no/ha) } \end{gathered}$ | Water level during spawning ${ }^{b}$ (m., M.S.L.) | Water level fluctuation (m.) |
| :---: | :---: | :---: | :---: |
| 1965 | 54.6 | 285.62 | -0.286 |
| 1966 | 24.8 | 284.66 | -0.674 |
| 1967 | 95.2 | 283.77 | -0.518 |
| 1968 | 87.8 | 284.39 | 0.600 |
| 1969 | 141.9 | 284.84 | 0.869 |
| 1970 | 7.4 | 284.81 | 0.104 |
| 1971 | 5.5 | 283.41 | -0.472 |
| 1972 | 0.13 | 282.57 | -0.613 |
| 1973 | 447.4 | 285.53 | 5.432 |
| 1974 | 200.5 | 287.81 | 1.122 |
| 1975 | 266.4 | 287.92 | 0.277 |
| 1976 | 88.9 | 286.99 | -0.390 |
| Mean: | 118.4 | 285.19 | 0.4587 |
| Standard deviation: | 132.0 | 1.68 | 1.5526 |

adjusted to 13 August (1965-1975 data from Summerfelt and Shirley 1976, unpub1. manuscript).
$\mathrm{b}_{\text {Mean }}$ water level during May.
${ }^{C}$ Fluctuation in water level from 1 October of previous growing season to 15 May.


Figure 10. Relationship between year-class strength as indexed by ecological density of young-of-the-year largemouth hass on 13 August and water level fluctuation.
and water level fluctuation accounted for $77.07 \%\left(R^{2}=0.7707\right)$ of the observed variation in density. The residuals in this analysis were greatest for 1970,1971 , and 1972 , when the water level was at an extreme low, and 1975, when water level was at or near spillway all year. These data indicate the importance of actual water level in addition to water level fluctuation in year-class formation.

The relationship between water level during spawning and yearclass strength and results of linear regression are illustrated in Figure 11. Although this correlation ( $\mathrm{r}=0.5337$ ) was not significant $(P=0.0739)$, analysis without the 1973 data yielded a highly significant $(P=0.0044)$ correlation coefficient of 0.7828 . Even though the water level during the 1973 spawning season was more than 2 meters below spillway, the water level was rising rapidly which resulted in very successful largemouth bass reproduction and YOY survival and growth. Thus, it appears that there is an important interaction between water level during spawning and water level fluctuation.

Results of the multiple regression using these two variables as predictor variables is summarized in Table 4. The equation for predicting YOY bass density on 13 August (Y) is

$$
\begin{equation*}
Y=-7601.3833+62.5356\left(X_{1}\right)+26.9689\left(X_{2}\right) \tag{6.1}
\end{equation*}
$$

where $X_{1}$ = water level fluctuation from 1 October of previous fall to 15 May (meters), and $X_{2}=$ mean water level during May (meters, M.S.L.). This relationship is highly significant since the calculated $\mathrm{F}=$ 33.6198 for regression has an associated probability of a greater Fvalue of 0.0002 . Furthermore, these two variables account for $88.20 \%$ $\left(\mathrm{R}^{2}=0.8820\right)$ of the observed variation in density. This value is a


Figure 11. Relationship between year-class strength as indexed by ecological density of young-of-the-year largemouth bass on 13 August and water level during May.

Table 4. Analysis of variance table for multiple regression analysis of dependent variable - young-of-the-year density - and independent variables - water level fluctuation ( $\mathrm{X}_{1}$ ) and mean water level during May ( $\mathrm{X}_{2}$ ).

| Source | d.f. | S.S. | M.S. | $F$ | $P$ |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Corrected total | 11 | 191669.1068 |  |  |  |
| Regression | 2 | 169042.7681 | 84521.3841 | 33.6198 | 0.0002 |
| R $\left(\mathrm{b}_{1} \mid \mathrm{b}_{0}\right)$ | 1 | 147714.9289 | 147714.9289 | 58.7560 | 0.0001 |
| R $\left(\mathrm{b}_{2} \mid \mathrm{b}_{0}, \mathrm{~b}_{1}\right)$ | 1 | 21327.8392 | 21327.8392 | 8.4835 | 0.0172 |
| Residual | 9 | 22626.3387 | 2514.0376 |  |  |

substantial increase in the $R^{2}$ observed for the regressions with either water level fluctuation alone $\left(R^{2}=0.7707\right)$ or water level during May alone $\left(R^{2}=0.2848\right)$. Also, the addition to the model of the second variable, water level during spawning, was significant as evidenced by the sequential F-test (Draper and Smith 1966:71-72) (F=8.4835; 1, 9; $\mathrm{P}=0.0172$ ). Water level fluctuation is over twice as important as water level during May in predicting YOY bass density because the ratio of standardized regression coefficients was 2.317 (0.7959/0.3435).

In order to include this relationship in the population dynamics mode1, it was necessary to relate survival from egg stage to age $I$ to these two variables. Survival rates were estimated for years when reliable population estimates were available for the spawning population in the spring and the number of yearlings the following spring. Population estimates were adjusted to 15 May (the approximate midpoint of the spawning period) by assuming a constant exponential mortality and using the average age-specific survival rates from Zweiacker et al. (1973). Using a logarithmic transformation of fecundity and length data from Kelley (1962) and Coomer (1976) a linear regression equation was derived. Age-specific fecundities were then estimated based on the mean lengths presented in Zweiacker et al. (1973) for age groups of largemouth bass from Lake Carl Blackwell (Table 5). Egg potential was estimated by the equation

$$
\sum_{i=3}^{8} N_{i} m_{i}(0.5)
$$

where $N_{i}=$ number of fish per age $i$, and $m_{i}=$ number of eggs per female of age i. This equation assumes a 1:1 sex ratio and that females mature at age III. Zweiacker et a1. (1973) noted that in Lake Carl Blackwe11,

Table 5. Mean total length and estimated agespecific fecundity of largemouth bass in Lake Car1 Blackwell.

| Age | Mean <br> total length <br> $(\mathrm{mm})$ | Number of eggs <br> per female |
| :---: | :---: | :---: |
| III | 369 | 18487 |
| IV | 425 | 28917 |
| V | 462 | 37665 |
| VI | 485 | 43929 |
| VII | 504 | 49613 |
| VIII | 531 | 58527 |

a few age II bass spawned but most do not spawn until age III. The resulting estimates of number of fish per age group in the spring, egg potential, and annual instantaneous mortality rates ( $Z_{o}$ ) from egg to age I are presented in Table 6. Annual instantaneous mortality rate (Z) is related to yearly survival rate (S) by

$$
\begin{equation*}
s=e^{-Z} \tag{6.3}
\end{equation*}
$$

where $\mathrm{e}=$ the base of the natural logarithm.
Results of the multiple regression analysis using water level fluctuation from 1 October of the previous fall to 15 May (meters) $\left(\mathrm{X}_{1}\right)$ and mean water level during May (meters, M.S.L.) ( $\mathrm{X}_{2}$ ) to predict the annual instantaneous mortality rate ( $Z_{o}$ ) from egg to age I are summarized in Table 7. The equation for predicting $Z_{o}$ is

$$
\begin{equation*}
Z_{o}=230.8063-0.9689\left(X_{1}\right)-0.7757\left(x_{2}\right) . \tag{6.4}
\end{equation*}
$$

This regression equation is significant ( $\mathrm{F}=13.1073$; 2, 4; $\mathrm{P}=0.0193$ ) and accounts for $86.76 \%\left(R^{2}=0.8676\right)$ of the observed variation in $Z_{0}$. Figure 12 shows how well the observed and predicted values coincide. Regression coefficients were converted for use with water level data recorded in feet rather than meters since lake levels for Lake Carl Blackwe11 and most reservoirs are recorded in feet. This equation was then incorporated into the framework of Model I. If water level data are not available, the computer program will use the average survival rate of age group 0 as computed in Model I. A program listing of Model II is given in Appendix $C$ and a sample output is in Appendix D. Model II is essentially the same as Model I except for the equation to predict mortality from egg to age $I$, parameters to account for the percentage of each age group that are mature and female, and use of

Table 6. Estimated number of fish per age group in the spring, egg potential, and annual instantaneous mortality rate ( $Z_{0}$ ) from egg to age $I$ for largemouth bass in Lake Carl Blackwell, Oklahoma.

| Age | Number of fish per age group |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $1968{ }^{\text {a }}$ | $1969{ }^{\text {a }}$ | $1970{ }^{\text {b }}$ | $1971{ }^{\text {b }}$ | $1972{ }^{\text {c }}$ | $1973{ }^{\text {c }}$ | $1974{ }^{\text {c }}$ | $1975{ }^{\text {c }}$ |
| I | 1151 | 357 | $178{ }^{\text {C }}$ | $306{ }^{\text {c }}$ | 322 | $32^{\text {d }}$ | $78741^{\text {d }}$ | $12640{ }^{\text {e }}$ |
| II | 305 | 766 | 241 | $120^{\text {c }}$ | 207 | 217 | $22^{\text {d }}$ | - |
| III | 192 | 138 | 472 | 148 | 74 | 127 | 134 | 14 |
| IV | 175 | 269 | 91 | . 311 | 62 | 49 | 84 | 88 |
| v | 206 | 142 | 151 | 51 | 53 | 34 | 27 | 47 |
| VI | 78 | 70 | 53 | 57 | 27 | 20 | 13 | 10 |
| VII | 24 | 15 | 14 | 10 | 11 | 5 | 4 | 3 |
| VIII | - | - | 1 | 1 | - | 1 | - | - |
| $\begin{aligned} & \text { Egg } \\ & \text { Potential: } \end{aligned}$ | 10493072 | 9748767 | 10063037 | 8354395 | 3444482 | 3115281 | 3346385 | 2580950 |
| $z_{0}:$ | 10.28849 | 10.91087 | 10.40079 | 10.16375 | 11.58655 | 3.67791 | 5.57877 |  |

$\mathbf{a}_{\text {Zweiacker (1972: 54) }}$
${ }^{\mathrm{b}}$ From Spring 1969 estimates
From Shirley's (unpubl. data) eatimate of 6 October 1972
${ }^{d}$ Shirley (1975: $32 \& 39$ )
$\mathrm{e}_{\text {From Summerfelt }}$ and Shirley (1975: 34) estimate of 14 October 1974

Table 7. Analysis of variance table for multiple regression analysis of dependent variable - annual instantaneous mortality rate ( $Z_{0}$ ) from egg to age $I$ - and independent variables water level ${ }^{\mathrm{o}}$ fluctuation ( $\mathrm{X}_{1}$ ) and mean water level during May ( $\mathrm{X}_{2}$ ).

| Source | d.f. | S.S. | M.S. | F | P |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Corrected total | 6 | 55.3258 |  |  |  |
| Regression | 2 | 48.0014 | 24.0007 | 13.1073 | 0.0193 |
| R $\left(\mathrm{b}_{1} \mid \mathrm{b}_{0}\right)$ | 1 | 40.0728 | 40.0728 | 21.8846 | 0.0095 |
| R $\left(\mathrm{b}_{2} \mid \mathrm{b}_{0}, \mathrm{~b}_{1}\right)$ | 1 | 7.9286 | 7.9286 | 4.3300 | 0.1059 |
| Residual | 4 | 7.3244 | 1.8311 |  |  |



Figure 12. Observed and predicted annual instantaneous mortality rates ( $Z_{0}$ ) from egg to age I for largemouth bass in Lake Carl Blackwe11 (1968-1974).
fecundity per female rather than fecundity per individual.

Results and Discussion

Simulation of Model II was made using initial age structure for spring 1968 (Table 6), age-specific survival rates from Zweiacker et al. (1973), age-specific fecundity from Table 5, and water level data for Lake Carl Blackwell from 1968 to 1977. The simulated predictions of year-class strength as indicated by the number of age I recruits is compared with the observed number of age I recruits in Figure 13. In terms of precision, Model II would rate very highly because of the close agreement between observed and simulated values. However, the model cannot be validated with data that was used for its derivation. An effort should be made in the future to collect data from the Lake Carl Blackwell bass population to validate the model, but this is beyond the scope of the present project.

Also we must take note of the confidence limits on the population estimates used before we condemn or praise the model. For example, Shirley's (1975) Schnabel estimate of the number of age I recruits in the 1973 year-class was 79,098 with $95 \%$ confidence limits of 51,718 and 135,825 . Discrepancies between simulated and observed number of age I recruits could be attributed to errors in the population estimates or errors in the model.

Model II should prove to be of value in largemouth bass fishery management by enabling the fishery biologists to quickly and easily predict year-class strength for any given year and hence the future population size and structure. With this information at hand the fishery managers can make better decisions on stocking recommendations


Figure 13. Observed estimates and Model II simulation of number of age $I$ recruits in year-classes 1968 through 1977 in Lake Carl Blackwe11.
and creel limits.
It is unlikely, though, that the parameters derived in this study for the relationship between year-class strength and water level fluctuation and water level during spawning will be exactly the same for all reservoirs. Therefore, it is necessary that research be done on other reservoirs to evaluate the generality of this relationship and to determine the appropriate parameter values for these reservoirs. The program has been written so that different parameter values for this relationship can be used by changing only one program statement in SUBROUTINE YOYSRV. This subroutine could also be easily adapted to use other equations to predict survival from egg to age $I$.

MODEL III

## Introduction

Model III is an extension of the previous models to allow the prediction of production and yield. Production is the total elaboration of fish tissue during any time interval, and yield is that portion of production that is used by man. Estimates of production and yield of largemouth bass populations are extremely useful to fishery managers and ecologists since the largemouth bass is an important game fish and also one of the top carnivores of aquatic ecosystems.

Production and sustainable yield of a fish stock should, according to the logistic model (Schaefer and Beverton 1963), be at a maximum when biomass is at one-half of carrying capacity. Traditionally, maximum sustainable yield of fish stocks has been the objective of fisheries management but more recently the concept of optimum sustainable yield has become the accepted philosophy (Larkin 1977; Nielsen 1976; Roedel 1975). Roedel (1975) defined optimum sustainable yield as

[^1]
## Mode1 Description

Model III is similar to the previous models in that it is agestructured and its basic time unit is a year. Survival from egg to age $I$ is calculated as in Model II. Other state variables are computed as follows with FORTRAN variable names given in parentheses when different from those used here.

$$
\begin{align*}
N_{i+1}(t+1)= & \text { number of fish in age group } i+1(i=0,1,2, \ldots k) \text { at } \\
& \text { time } t+1 \\
= & N_{i}(t) e^{-Z_{i}(t)} \tag{7.1}
\end{align*}
$$

where

$$
\begin{aligned}
& \mathrm{k}=\text { maximum age (INPUT), } \\
& \mathrm{e}=\text { the base of the natural logarithm, } \\
& Z_{i}(t)=\text { instantaneous annual total mortality rate on age group } \\
& \text { i during time period } t, t+1 \\
& =F_{i}(t)+M_{i}(t), \\
& F_{i}(t)=\text { instantaneous annual fishing mortality rate on age } \\
& \text { group } i \text { during time period } t, t+1 \text { (INPUT) } \\
& =q_{i}(t) f(t), \\
& q_{i}(t)=\text { catchability coefficient (vulnerability) of age group } \\
& i \text { during time period } t, t+1 \text {, } \\
& \mathrm{f}(\mathrm{t})=\text { fishing effort during time period } \mathrm{t}, \mathrm{t}+1 \text {, } \\
& \text { and } M_{i}(t)=\text { instantaneous annual natural mortality rate on age }
\end{aligned}
$$

where

$$
\begin{align*}
\mathrm{m}_{i}(\mathrm{t})= & \text { number of eggs produced per female of age group } i \text { at } \\
& \text { time } t \\
= & (\text { FECND }) \\
= & a \overline{1}_{i}(t)^{b}, \tag{7.5}
\end{align*}
$$

$\overline{1}_{i}(t)=$ average total length (mm) of individuals of age group $i$ at time t
$=(\mathrm{AVGTL})$, a = constant in fecundity estimation equation (INPUT)
$=(\mathrm{AFEC})$,
b = exponent in fecundity estimation equation (INPUT)
$=(\mathrm{BFEC})$,
$P F_{i}=$ proportion of age group $i$ that is female (INPUT)
$=$ (FEMALE),
and $\mathrm{PM}_{i}=$ proportion of age group $i$ that is mature (INPUT)
= (MATURE).
$\mathrm{B}_{\mathrm{o}}(\mathrm{t})=$ biomass of eggs produced at time t (kg)
$=$ (EGGB)
$=$ EGGW N $N_{o}(t) 0.001$
where
EGGW = individual egg weight
$=0.0012$ grams (based on estimated specific gravity of 1.47 and mean egg diameter of 1.16 mm ),
$\begin{aligned} \overline{\mathrm{I}}_{i+1}(\mathrm{t}+1)= & \text { average total length (mm) of individuals of age group } i+1 \\ & \text { at time } t+1 \\ = & \text { (AVGTL) } \\ = & \overline{1}_{i}(t) e^{G_{i}}(\mathrm{t})\end{aligned}$
where
$G_{i}(t)=$ instantaneous annual rate of growth in total length for age group $i$ during time period $t, t+1$ (INPUT)
$=(G T L)$
$=\log _{e}\left[\overline{1}_{i+1}(t+1) / \overline{1}_{i}(t)\right]$.
$\bar{w}_{i+1}(t+1)=$ average weight $(g)$ of individuals of age group $i+1$ at time $\mathrm{t}+1$
$=($ AVGW $)$
$=\bar{w}_{i}(t) e^{b G_{i}(t)}$
where
$b=$ exponent in the length weight relationship: $w=a 1^{b}$ (INPUT)
$=$ (BWTLEN), and
$b G_{i}(t)=$ instantaneous rate of growth in weight of age group $i$
during time period $t, \mathrm{t}+1$
$=(\mathrm{GW})$.
$B_{i}(t)=$ biomass $(k g)$ of age group $i$ at time $t$
$=N_{i}(t) \bar{w}_{i}(t) 0.001$.
$\bar{N}_{i}(t)=$ average number of fish of age group $i$ during time period
$t, t+1$
$=($ AVGN $)$
$=\int_{t}^{t+1} N_{i}(t) e^{-Z}{ }_{i}^{(t)} d t$
$=\frac{N_{i}(t)\left(1-e^{-Z_{i}(t)}\right)}{Z_{i}(t)}$.
$\bar{B}_{i}(t)=$ average biomass (kg) of age group i during time period $t, t+1$
$=(\mathrm{AVGB})$

$$
\begin{align*}
& =\int_{t}^{t+1_{B_{i}}(t)} e^{\left[b G_{i}(t)-Z_{i}(t)\right] t} d t  \tag{7.13}\\
& =\frac{B_{i}(t)\left[1-e^{-\left[z_{i}(t)-b G_{i}(t)\right]}\right]}{Z_{i}(t)-b G_{i}(t)} \text { when } b G<Z \tag{7.14}
\end{align*}
$$

$=B_{i}(t)$ when $b G=Z$
$=\frac{B_{i}(t)\left[e^{\left[b G_{i}(t)-Z_{i}(t)\right]}-1\right]}{b G_{i}(t)-Z_{i}(t)}$ when $b G>Z$
$C_{i}(t)=$ number of age group $i$ harvested during time period $t, t+1$

$$
\begin{equation*}
=F_{i}(t) \bar{N}_{i}(t) \tag{7.17}
\end{equation*}
$$

$Y_{i}(t)=$ weight (kg) of age group $i$ harvested during time period $t$, t+1
$=F_{i}(t) \bar{B}_{i}(t)$.
$\mathrm{GP}_{\mathrm{i}}(\mathrm{t})=$ gross production (kg) of age group $i$ during time period t , $\mathrm{t}+1$
$=b G_{i}(t) \bar{B}_{i}(t)$.
$N P_{i}(t)=$ net production (kg) of age group $i$ during time period $t, t+1$
$=\left[b G_{i}(t)-Z_{i}(t)\right] \bar{B}_{i}(t)$.
Numbers, biomass, production and yield are then summed over ages 1 to $k$ to give the level of these state variables for the entire stock for each year of simulation. The instantaneous rates of growth and natural and fishing mortality, G, M, and F, respectively, are expressed here as time-varying coefficients but in most cases they will be constant for each simulation run. Program statements could be added to the computer program to make growth and mortality a function of population number, biomass or environmental factors. Output from Model III
consists of number at start of year, mean number during year, mean total length, mean weight per fish, biomass at start of year, mean biomass during year, yield in weight and numbers, and gross and net production for each age group. In addition number at start of year, mean number during year, biomass at start of year, mean biomass during year; yield in weight and numbers and gross and net production for the entire stock is given. A computer program listing, sample output, and sample input data for Mode1 III are included in Appendices E, F, and G, respectively. Derivations of the parameters to be used as input data are described in succeeding sections.

Fecundity

Although numerous studies on fecundity have been made, there are few investigations where sample size allows quantification of the relationship between fecundity and age or size. Part of the problem is that there is typically great variability in the number of eggs in fish of the same length, weight, and age because environmental factors, such as food supply, influence the amount of energy channeled into gonadal development. To avoid this problem most authors studying the fecundity of various fish species have plotted fecundity and length data as a scatter diagram and have concluded that the relationship is of the form

$$
\begin{equation*}
\mathrm{m}=\mathrm{al} \mathrm{~b}^{\mathrm{b}} \tag{7.21}
\end{equation*}
$$

where $m=$ fecundity, $1=$ fish length, and $a$ and $b$ are $a$ constant and an exponent derived from the data respectively (Bagenal 1967). This curve can be transformed to a straight line by a logarithmic transformation:

$$
\begin{equation*}
\log m=\log a+b \log 1 \tag{7.22}
\end{equation*}
$$

and the logarithmic values analyzed by simple linear regression analysis.

Fecundity estimates and lengths are presented (Figure 14) for largemouth bass from a small infertile lake in northern Michigan (Clady 1970), large reservoirs in Tennessee (Coomer 1976) and Arkansas (O1msted 1974), and a stream in Maine (Kelley 1962). These data were fitted to a line of the form in 7.22 by linear regression and the results of the analyses presented in Table 8.

Regressions for each author's data separately and the combined data were highly significant (Table 8) but some of the coefficients were different. Analysis of covariance (Snedecor and Cochran 1967:432436) was employed to compare the regression lines. Since this analysis assumes homogeneity of variance, Bartlett's test (Snedecor and Cochran 1967:296-298) was applied to compare the residual mean squares from the four sets of data. The chi-square value, corrected for unequal sample size, was 8.86 ( 3 d.f.) which has an associated probability of a greater chi-square of 0.0312 (Table 9). This probability makes the assumption of equal variances invalid. One reason for the unequal variances could be the different length ranges sampled by the authors (Table 8). Variability in fecundity tends to increase with an increase in fish size (Bagenal 1967). There was a significant positive correlation ( $\mathrm{r}=0.9531$; $\mathrm{P}=0.0369$ ) between the residual mean squares (Tab1e 9) and mean total length of bass sampled (Table 8). Data from Clady (1970) included slow growing bass from a narrow length range ( $254-368 \mathrm{~mm}$ ), with the largest of these being 10 years old. The fish may not display a typical length-fecundity relationship because of the relatively low variance in Clady's data.


Figure 14. Scatter diagram of fecundity and total length for largemouth bass collected from a lake in Michigan (Clady 1970), reservoirs in Tennessee (Caamer 1976) and Arkansas (O1msted 1974), and a stream in Maîne (Kelley 1962).

Table 8. Results of regressions of log-transformed values of length and fecundity of largemouth bass from Michigan (Clady 1970), Tennessee (Coomer 1976), Maine (Kelley 1962), and Arkansas (Olmsted 1974).

| Author | No. of fish | Total length (mm) |  | Fecundity |  | a | b | $\mathrm{F}^{\text {a }}$ | $\mathrm{P}^{\text {b }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\overline{\mathrm{X}}$ | range | $\overline{\mathrm{X}}$ | range |  |  |  |  |
| Clady 1970 | 26 | 306.6 | (254-368) | 18728.5 | (7511-28536) | 3.6608 | 1.4860 | 8.66 | 0.0071 |
| Coomer 1976 | 20 | 359.6 | (218-461) | 17917.6 | (2137-46128) | $5.342 \times 10^{5}$ | 3.3149 | 47.55 | 0.0001 |
| Kelley 1962 | 20 | 404.0 | (295-503) | 31564.8 | (5549-81582) | $3.642 \times 10^{-4}$ | 3.0162 | 19.19 | 0.0004 |
| O1msted 1974 | 16 | 334.9 | (252-523) | 10462.9 | (2942-30709) | $2.042 \times 10^{-3}$ | 2.6276 | 51.81 | 0.0001 |
| Combined | 82 | 348.8 | (218-523) | 20048.7 | (2137-81582) | $9.622 \times 10^{-3}$ | 2.4558 | 70.93 | 0.0001 |

[^2]Table 9. Comparison of residual mean squares of log-transformed fecundity data for largemouth bass from Michigan (Clady 1970), Tennessee (Coomer 1976), Maine (Kelley 1962), and Arkansas (Olmsted 1974) by Bartlett's test for unequal sample size.

| Author | $\begin{gathered} \mathrm{d} . \mathrm{f} \\ \mathrm{f}_{\mathrm{i}} \end{gathered}$ | $\begin{aligned} & \text { S.S. } \\ & \mathrm{f}_{\mathrm{i}} \mathrm{~s}_{\mathrm{i}} \end{aligned}$ | $\begin{gathered} \text { M.S. } \\ \mathrm{s}_{\mathrm{i}}^{2} \end{gathered}$ | $\log \mathrm{s}_{\mathrm{i}}^{2}$ | $\mathrm{f}_{\mathrm{i}} \log \mathrm{s}_{\mathrm{i}}^{2}$ | $1 / \mathrm{f}_{\mathrm{i}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Clady 1970 | 24 | 0.313157 | 0.013048 | -1.884456 | -45.226945 | 0.041667 |
| Coomer 1976 | 18 | 0.414454 | 0.023025 | -1.637800 | -29.480406 | 0.055555 |
| Kelley 1962 | 18 | 0.841928 | 0.046774 | -1.329995 | -23.939919 | 0.055555 |
| O1msted 1974 | 14 | 0.266395 | 0.019028 | -1.720601 | -24.088414 | 0.071429 |
| Totals: | 74 | 1.835934 | 0.101875 |  | -122.735684 | 0.224207 |
| - $\mathrm{a}=4$ |  |  |  |  |  |  |
| $M=(2.3026)\left[\left(\Sigma f_{i}\right) \log \left(\Sigma f_{i} s_{1}^{2} / \Sigma f_{i}\right)-\Sigma f_{i} \log s_{i}^{2}\right]=9.067546$ |  |  |  |  |  |  |
| $C=1+\frac{1}{3(a-1)}\left[\sum^{\prime} \frac{1}{f_{i}}-\frac{1}{\sum f_{i}}\right]=1.023410$ |  |  |  |  |  |  |
| $\chi^{2}=M / C=8.8601$ with 3 d.f. |  |  |  |  |  |  |
| $\mathrm{P}=0.0312$ |  |  |  |  |  |  |

Bartlett's test was applied to compare the residual mean squares of the data from Coomer (1976), Kelley (1962) and Olmsted (1974), omitting that from Clady (1970). The analysis resulted in a nonsignificant chi-square value of 3.79 ( 2 d.f.) and thus equal variances. Covariance analysis (Table 10) indicated that the regression lines for these three data sets were paralle1 ( $\mathrm{F}=0.16$; 2 and $50 \mathrm{~d} . \mathrm{f} . ; \mathrm{P}=0.8521$ ). The F -test for adjusted means was significant ( $\mathrm{F}=7.11 ; 2$ and 52 d.f.; $\mathrm{P}=0.0019$ ) indicating that if the mean logarithm of fecundity for each data set was adjusted to the same logarithm of total length the results would be significantly different. This difference was due primarily to the lower adjusted mean from Olmsted's data since the F-test for adjusted means from covariance analysis of Kelley's data and Coomer's data was non-significant ( $\mathrm{F}=1.07$; 1 and 37 d.f.; $\mathrm{P}=0.3072$ ).

The analysis reported above removes the variation due to techniques, types of study areas and/or geographic location. Therefore the parameters derived in Table 10 should be fairly representative of the length-fecundity relationship for largemouth bass.

## Growth Rates and Length-Weight Relationships

Model III requires parameters for the length-weight relationship and age-specific growth rates. There appeared to be an important, and possibly predictable, trend in growth rates of largemouth bass in Lake Carl Blackwell from 1962 through 1967 (Zweiacker et al. 1973). Therefore, 2384 largemouth bass collected from Fall 1972 through Spring 1977 plus an additional 64 bass collected in the spring of 1967 were weighed, measured and scale samples taken. Scale impressions were made on plastic slides, examined at 41.5 magnification with a 16 mm micro-

Table 10. Analysis of covariance and comparison of regression lines for log-transformed length-fecundity relationship for largemouth bass from Tennessee (Coomer 1976), Maine (Kelley 1962) and Arkansas (O1msted 1974).

| Author | d.f. | $\Sigma x^{2}$ | Exy | Deviations from regression |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\Sigma y^{2}$ | d.f. | s.s.a | M.S. |
| Coomer 1976 | 19 | 0.099641 | 0.330303 | 1.509386 | 18 | 0.414454 | 0.023025 |
| Kelley 1962 | 19 | 0.098648 | 0.297543 | 1.739380 | 18 | 0.841928 | 0.046774 |
| 01msted 1974 | 15 | 0.142793 | 0.375202 | 1.252266 | 14 | 0.266388 | 0.019028 |
|  |  |  |  |  | 50 | 1.522770 | 0.088827 |
| Pooled, W | 53 | 0.341082 | 1.003048 | 4.501032 | 52 | 1.551286 | 0.029832 |
|  |  | Difference between slopes |  |  | 2 | 0.028516 | 0.014258 |
| Between, B | 2 | 0.069150 | 0.380606 | 2.141083 |  |  |  |
| $w+B$ | 55 | 0.410232 | 1.383654 | 6.642115 | 54 | 1.975247 |  |
|  | Difference between adjusted means |  |  |  | 2 | 0.423961 | 0.211980 |
| Comparison of slopes: $\mathrm{F}=0.014258 / 0.088827=0.16051$ |  |  |  |  |  |  |  |
| 2 and 50 d.f. $P=0.8521$ N.S. |  |  |  |  |  |  |  |
| Comparison of adjusted means: $\mathrm{F}=0.211980 / 0.029832=7.106$ |  |  |  |  |  |  |  |
| 2 and $52 \mathrm{~d} . \mathrm{f} . \quad \mathrm{P}=0.0019$ |  |  |  |  |  |  |  |
| Fecundity $=0.00045091$ Length ${ }^{2.941}$ |  |  |  |  |  |  |  |

$$
\mathrm{a}_{\text {S.S. }}=\Sigma \mathrm{y}^{2}-\left[(\Sigma \mathrm{xy})^{2} / \Sigma \mathrm{x}^{2}\right]
$$


#### Abstract

tessar lens and lengths from scale focus to each annulus measured. Linear and curvilinear (5th degree polynomial) regressions were calculated for the total length-scale radius relationship and used to backcalculate the total length of bass at the time of formation of each annulus. These growth rate data were combined with that of Zweiacker et al. (1973) and analyzed by correlation and regression techniques to determine the relationships with environmental factors. Parameters for the length-weight relationships, $w=a 1^{b}$, were derived by linear regression using a logarithmic transformation of the data, $\log \mathrm{w}=\log$ $a+b \log 1$.


The majority (8 of 12 ) of the total length-scale radius relationships used for back calculating length-at-annulus for various collection periods were linear (Table 11) and all regressions were highly significant $(\mathrm{P}<0.005)$. Back-calculated lengths at annulus and growth increments for largemouth bass from Lake Carl Blackwell are presented for years 1959 to 1976 in Table 12. In general, growth was above the Oklahoma average (Houser and Bross 1963) as also noted by Zweiacker et al. (1973). However, growth patterns since 1974 are unusual in that increments for bass in the second year of life are well below average for Lake Carl Blackwe11. Growth increments for age 3 bass from the 1973 and 1974 year-classes were below average as were increments for age 4 bass from the 1971 and 1973 year-classes.

Correlation analysis was performed on growth increment data
(Table 12) and average annual water level (AAWL), mean temperature from May through October (TEMP), mean annual turbidity (TURB), density of young-of-the-year bass on 13 August (YOYD), standing crop of all bass (LMBSC), density of gizzard shad (GSD), and standing crop of gizzard

Table 11. Total length-scale radius (X41.5) relationships used for back-calculating length at annulus for largemouth bass from Lake Carl Blackwell.

| Time of collection | Total length-scale <br> radius (X41.5) relationships | $\mathrm{R}^{2}$ |
| :---: | :---: | :---: |
| Spring 1977 | $\mathrm{Y}=31.2586+1.6600 \mathrm{X}$ | 0.9331 |
| Fall 1976 | $\mathrm{Y}=13.6770+1.7152 \mathrm{X}$ | 0.9033 |
| Fal1 1975 | $\mathrm{Y}=21.1284+1.6959 \mathrm{X}$ |  |
|  | $-\left(5.905 \times 10^{-4}\right) \mathrm{x}^{2}$ | 0.9429 |
| Fall 1974 | $Y=44.9098+1.4382 \mathrm{X}$ |  |
|  | $-\left(3.977 \times 10^{-11}\right) \mathrm{x}^{5}$ | 0.7320 |
| August 1974 | $\mathrm{Y}=64.3382+1.2016 \mathrm{X}$ | 0.3767 |
| Spring 1974 | $\mathrm{Y}=31.5970+1.4797 \mathrm{X}$ |  |
|  | $+\left(2.1777 \times 10^{-6}\right) \mathrm{x}^{3}$ |  |
|  | $-\left(2.1090 \times 10^{-11}\right) \mathrm{x}^{5}$ | 0.9264 |
| Fall 1973 | $\mathrm{Y}=35.2742+1.5032 \mathrm{X}$ |  |
|  | $-\left(5.2330 \times 10^{-12}\right) \mathrm{x}^{5}$ | 0.7714 |
| Spring 1973 | $\mathrm{Y}=33.0287+1.5485 \mathrm{X}$ | 0.9253 |
| Fall 1972 | $\mathrm{Y}=23.1977+1.5733 \mathrm{X}$ | 0.8264 |
| 1971 from |  |  |
| Zweiacker |  |  |
| 1968 \& 1969 |  |  |
| Zweiacker <br> et al. (1973) | $Y=37.67+1.29 \mathrm{X}$ | 0.9409 |
| Spring 1967 | $\mathrm{Y}=19.1374+1.7615 \mathrm{X}$ | 0.9123 |

Table 12. Mean back-calculated total lengths (mm) and annual growth increments at end of each year of life for largemouth bass in Lake Carl Blackwell, 19591976 (sample sizes in parentheses).

| Year Class | I |  | II |  | Total lengths (T.L.) and growth increments (Inc.)IIIIV |  |  |  |  |  |  |  |  |  | vi |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | T.L. | Inc. | T.L. | Inc. | т.L. | Inc. | т.L. | Inc. | T.L. | Inc. | T.L. | Inc. | T.L. | Inc. | T.L. | Inc. |
| 1976 | ${ }_{(21)}^{150.3}$ | 150.3 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1975 | $\begin{aligned} & 150.4 \\ & (189) \end{aligned}$ | 150.4 | $\begin{aligned} & 239.3 \\ & (53) \end{aligned}$ | 79.5 |  |  |  |  |  |  |  |  |  |  |  |  |
| 1974 | $\begin{aligned} & 139.8 \\ & (388) \end{aligned}$ | 139.8 | $\begin{gathered} 229.8 \\ (185) \end{gathered}$ | 83.6 | $\begin{aligned} & 283.2 \\ & (96) \end{aligned}$ | 43.2 |  |  |  |  |  |  |  |  |  |  |
| 1973 | $\begin{gathered} 177.0 \\ (1338) \end{gathered}$ | 177.0 | $\begin{aligned} & 231.3 \\ & (200) \end{aligned}$ | 82.4 | $\begin{aligned} & 295.2 \\ & (99) \end{aligned}$ | 41.3 | $\begin{aligned} & 343.1 \\ & (59) \end{aligned}$ | 38.8 |  |  |  |  |  |  |  |  |
| $1972{ }^{\text {a }}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| 1971 | 157.9 <br> (196) | 157.9 | $\begin{aligned} & 278.6 \\ & (99) \end{aligned}$ | 113.1 | $\begin{aligned} & 359.8 \\ & (44) \end{aligned}$ | 72.7 | $\begin{aligned} & 394.3 \\ & (34) \end{aligned}$ | 37.1 | $\begin{aligned} & 429.7 \\ & (30) \end{aligned}$ | 39.1 | $\begin{aligned} & 467.8 \\ & (23) \end{aligned}$ | 24.8 |  |  |  |  |
| 1970 | $\begin{gathered} 132.7 \\ (99) \end{gathered}$ | 132.7 | ${ }_{(90)}^{272.6}$ | 138.5 | $\begin{aligned} & 350.8 \\ & (39) \end{aligned}$ | 88.0 | $\begin{aligned} & 429.7 \\ & (15) \end{aligned}$ | 66.3 | $\begin{aligned} & 487.0 \\ & (9) \end{aligned}$ | 46.9 | $\begin{aligned} & 504.6 \\ & (9) \end{aligned}$ | 17.6 | $529.5$ <br> (8) | 12.7 |  |  |
| 1969 | $\begin{aligned} & 134.3 \\ & (92) \end{aligned}$ | 134.3 | $\begin{aligned} & 257.5 \\ & (92) \end{aligned}$ | 123.2 | $\begin{aligned} & 345.7 \\ & (76) \end{aligned}$ | 85.3 | ${ }_{(36)}^{402.6}$ | 56.8 | $\begin{aligned} & 449.5 \\ & (7) \end{aligned}$ | 35.2 | $\underset{(2)}{513.0}$ | 18.4 | $\underset{(1)}{569.1}$ | 19.9 | $\underset{(1)}{577.4}$ | 8.3 |
| 1968 | $\begin{aligned} & 128.8 \\ & (165) \end{aligned}$ | 128.8 | $\underset{(63)}{251.1}$ | 128.9 | $\begin{aligned} & 341.9 \\ & (63) \end{aligned}$ | 88.1 | $\begin{aligned} & 398.8 \\ & (51) \end{aligned}$ | 51.7 | $\begin{aligned} & 437.1 \\ & (24) \end{aligned}$ | 44.5 | $\begin{aligned} & 450.0 \\ & (4) \end{aligned}$ | 26.5 |  |  |  |  |
| 1967 | $\begin{aligned} & 129.1 \\ & (521) \end{aligned}$ | 129.1 | $\begin{aligned} & 266.5 \\ & (254) \end{aligned}$ | 136.4 | $\begin{aligned} & 336.6 \\ & (35) \end{aligned}$ | 87.8 | $\begin{aligned} & 391.9 \\ & (35)^{2} \end{aligned}$ | 55.2 | $\begin{aligned} & 436.4 \\ & (33) \end{aligned}$ | 39.2 | $\begin{aligned} & 474.9 \\ & (13) \end{aligned}$ | 30.1 | $\begin{aligned} & 501.8 \\ & (5) \end{aligned}$ | 16.5 |  |  |
| 1966 | $\begin{aligned} & 146.7 \\ & (140) \end{aligned}$ | 146.7 | $\begin{gathered} 284.1 \\ (128) \end{gathered}$ | 136.3 | $\begin{aligned} & 360.6 \\ & (60) \end{aligned}$ | 75.2 | 380.7 <br> (8) | 67.8 | $\begin{aligned} & 420.4 \\ & (8) \end{aligned}$ | 39.7 | ${ }_{(8)}^{442.5}$ | 22.0 |  |  |  |  |
| 1965 | $\begin{aligned} & 151.2 \\ & (177) \end{aligned}$ | 151.2 | $\begin{gathered} 288.0 \\ (177) \end{gathered}$ | 137.7 | $\begin{aligned} & 373.4 \\ & (168) \end{aligned}$ | 85.0 | $\begin{aligned} & 424.0 \\ & (104) \end{aligned}$ | 50.9 | $437.0$ <br> (3) | 55.8 | $\begin{aligned} & 465.4 \\ & (3) \end{aligned}$ | 28.4 | $\underset{(2)}{485.0}$ | 33.8 |  |  |
| 1964 | $\begin{gathered} 158.2 \\ (158) \end{gathered}$ | 158.2 | $\begin{gathered} 283.4 \\ (158) \end{gathered}$ | 126.2 | $\begin{aligned} & 365.1 \\ & (158) \end{aligned}$ | 81.7 | $\begin{aligned} & 428.1 \\ & (150) \end{aligned}$ | 58.9 | $\begin{aligned} & 460.0 \\ & (96) \end{aligned}$ | 35.8 | ${ }_{(2)}^{486.2}$ | 22.1 | ${ }_{(2)}^{515.9}$ | 29.7 | $\begin{aligned} & 534.5 \\ & (1) \end{aligned}$ | 15.7 |
| 1963 | $\begin{aligned} & 150.7 \\ & (98) \end{aligned}$ | 150.7 | $\underset{(98)}{278.0}$ | 127.3 | $\begin{aligned} & 364.9 \\ & (98) \end{aligned}$ | 86.9 | ${ }_{(98)}^{419.6}$ | 54.6 | $\begin{aligned} & 465.0 \\ & (82) \end{aligned}$ | 41.0 | $\begin{aligned} & 484.0 \\ & (47)^{\circ} \end{aligned}$ | 20.0 | $497.0$ <br> (1) | 53.0 |  |  |
| 1962 | ${ }_{(26)}^{161.8}$ | 161.8 | $\underset{(26)}{290.2}$ | 127.7 | ${ }_{(26)}^{372.5}$ | 82.2 | $\begin{aligned} & 427.0 \\ & (26) \end{aligned}$ | 54.5 | $\begin{aligned} & 461.4 \\ & (26) \end{aligned}$ | 34.5 | ${ }_{(20)}^{490.0}$ | 24.0 | $\begin{aligned} & 508.0 \\ & (10) \end{aligned}$ | 21.0 |  |  |
| 1961 | ${ }_{(5)}^{127.3}$ | 127.3 | ${ }_{(5)}^{269.1}$ | 141.8 | ${ }_{(5)}^{350.9}$ | 81:9 | $\begin{aligned} & 411.3 \\ & (5) \end{aligned}$ | 60.3 | $\begin{aligned} & 464.8 \\ & (5) \end{aligned}$ | 53.6 | $\begin{aligned} & 498.2 \\ & (5) \end{aligned}$ | 33.4 | $\begin{aligned} & 481.0 \\ & (2) \end{aligned}$ | 26.0 |  |  |
| 1959 | $\underset{(1)}{138.9}$ | 138.9 | $\underset{(1)}{218.2}$ | 79.3 | $290.4$ <br> (1) | 72.2 | ${ }_{(1)}^{387.3}$ | 96.9 | ${ }_{(1)}^{422.5}$ | 35.2 | ${ }_{(1)}^{461.3}$ | 38.8 | $\underset{(1)}{512.4}$ | 51.1 | $\underset{(1)}{537.0}$ | 24.7 |
| Number | 3614 |  | 1629 |  | 968 |  | 622 |  | 324 |  | 137 |  | 32 |  | 3 |  |
| Weighted Means | 154.9 | 154.9 | 263.5 | 118.0 | 346.1 . | 75.2 | 409.4 | 53.0 | 453.8 | 39.2 | 479.6 | 23.3 | 511.5 | 21.8 | 549.6 | 16.2 |
| Means | 145.9 | 145.9 | 262.5 | 117.5 | 342.2 | 76.5 | 403.0 | 57.7 | 447.6 | 41.7 | 478.2 | 25.5 | 511.1 | 29.3 | 549.6 | 16.2 |

shad (GSSC) from Table 13. Results of this analysis appear in Table 14 with the first year growth increment denoted by GIO1, second year GI12, and so on. For the 1972 year-class a first year growth increment of 143.0 mm (the mean length on 6 October 1972) was used since there were not enough native bass collected the following spring to allow calculation of actual growth of that year-class (Shirley 1975). The only complete data sets for the 17 years are for average annual water level (AAWL) and the first year growth increment (GIO1). The correlation between these two variables is not significant ( $r=0.1530 ; \mathrm{P}=0.5578$ ) when all years are considered but the correlation for years 1962 through 1967 is significant ( $r=0.8456 ; \mathrm{P}=0.0339$ ). Zweiacker et al. (1973) found that in years 1962 through 1967 first year growth of largemouth bass in Lake Carl Blackwell was positively correlated with the average annual water level, whereas growth increments in the second, third and fourth years of life were negatively correlated with water level. These correlations were probably due to the influence that water level has on food availability. Low lake levels have a negative effect on the littoral zone invertebrates upon which age 0 bass feed but they also make forage fishes more vulnerable to predation by age $I$ and older bass. From 1962 to 1967 there was a continuing decline in the lake water level but during some of the other years the water level was rising (Figure 3). Therefore, all years were classified as rising-water or falling-water years based on the difference in mean month1y water levels in January and December of each year. Based on this classification, years 1959, 1961, 1968, 1969, 1973, and 1974 were rising-water years and the remaining 11 were falling-water years. Correlation between GIO1 and AAWL for the falling-water years was improved ( $\mathrm{r}=$

Table 13. Biological and physical parameters used in correlation analysis of growth increment data for largemouth bass in Lake Carl Blackwell, Oklahoma.

| Year | AAWL ${ }^{\text {a }}$ | TEMP ${ }^{\text {b }}$ | TURB ${ }^{\text {c }}$ | YOYD ${ }^{\text {d }}$ | LMBSC ${ }^{\text {e }}$ | GSD ${ }^{\text {f }}$ | $\operatorname{GSSC}^{\text {g }}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1976 | 286.6 |  |  | 88.9 | 5.87 | 2595.4 | 88.29 |
| 1975 | 287.9 |  |  | 266.4 | 16.16 | 5713.9 | 321.59 |
| 1974 | 287.8 | 23.18 | 28.72 | 200.5 | 24.54 | 7053.5 | 143.21 |
| 1973 | 285.4 | 23.23 | 67.76 | 447.4 | 6.49 | 15865.9 | 69.06 |
| 1972 | 282.9 | 23.52 | 46.09 | 0.13 |  |  |  |
| 1971 | 283.5 | 23.32 | 42.58 | 5.5 | 3.17 | 2981.8 | 54.90 |
| 1970 | 284.4 | 23.42 | 35.23 | 7.4 |  |  |  |
| 1969 | 284.0 | 23.61 | 47.34 | 141.9 |  |  |  |
| 1968 | 284.1 | 22.90 | 40.82 | 87.8 | 3.16 | 1422.6 | 39.11 |
| 1967 | 283.9 | 22.37 | 30.28 | 95.2 | 28.86 | 3387.3 | 66.87 |
| 1966 | 284.6 | 22.82 | 22.20 | 24.8 | 12.34 | 47.5 | 1.03 |
| 1965 | 285.6 | 24.13 | 27.46 | 54.6 |  |  |  |
| 1964 | 286.1 |  | 21.38 |  |  |  |  |
| 1963 | 287.0 |  | 20.50 |  |  |  |  |
| 1962 | 297.6 |  | 24.90 |  |  |  |  |
| 1961 | 287.8 |  | 30.65 |  |  |  |  |
| 1959 | 287.4 |  | 55.57 |  |  |  |  |
| Means | 285.68 | 23.25 | 36.10 | 118.38 | 12.57 | 4883.48 | 98.01 |
| Std. Dev. | 1.7173 | 0.4828 | 13.7425 | 132.0018 | 9.8586 | 4953.6658 | 99.0427 |
| $\begin{gathered} a_{\text {AAWL }} \\ b_{\text {TEMP }} \end{gathered}$ | $=$ averag $=$ mean $t$ | annual | water 1 | vel (m., M | S.L.)'. October | (C). |  |

Table 13. (Continued).
$c_{\text {TURB }}=$ mean annual turbidity (JTU).
${ }^{\mathrm{d}}$ YOYD $=$ density of young-of-the-year bass on 13 August (no./ha).
${ }^{\mathrm{e}} \mathrm{LMBSC}=$ standing crop of largemouth bass (kg/ha).
$\mathrm{f}_{\text {GSD }}=$ density of gizzard shad (no./ha).
$\mathrm{g}_{\text {GSSC }}=$ standing crop of gizzard shad (kg/ha).

Table 14. Correlation coefficients ( $r$ ) and probabilities for a greater value of $r$ (in parentheses) for growth increments (GI) and biological and physical parameters. Variable names defined in text and Table 13.

|  | GI01 | GI12 | GI23 | GI34 | GI45 | GI56 | GI67 | GI78 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| AAWL | 0.1530 | -0.5892 | -0.6591 | 0.0123 | -0.0132 | -0.3061 | -0.5205 | -0.7656 |
|  | $(0.5578)$ | $(0.0208)$ | $(0.0104)$ | $(0.9683)$ | $(0.9676)$ | $(0.3332)$ | $(0.1509)$ | $(0.4448)$ |
| TEMP | 0.2836 | -0.2247 | 0.3921 | 0.0986 | 0.6544 | 0.4419 | 0.2668 | - |
|  | $(0.4271)$ | $(0.5610)$ | $(0.2966)$ | $(0.7865)$ | $(0.0401)$ | $(0.2010)$ | $(0.5630)$ |  |
| TURB | 0.1633 | 0.0050 | -0.1441 | -0.0561 | -0.0060 | -0.2614 | -0.4525 | - |
|  | $(0.5610)$ | $(0.9871)$ | $(0.6549)$ | $(0.8625)$ | $(0.9859)$ | $(0.4656)$ | $(0.3080)$ |  |
| YOYD | 0.5456 | -0.6213 | -0.4313 | 0.2514 | -0.1507 | -0.3637 | -0.4104 | 0.1189 |
|  | $(0.0665)$ | $(0.0413)$ | $(0.1854)$ | $(0.4558)$ | $(0.6582)$ | $(0.2452)$ | $(0.2726)$ | $(0.9241)$ |
| LMBSC | -0.4603 | -0.1834 | 0.0940 | -0.0832 | 0.6850 | -0.1590 | 0.1256 | 0.8456 |
|  | $(0.2511)$ | $(0.6939)$ | $(0.8412)$ | $(0.8593)$ | $(0.0895)$ | $(0.7068)$ | $(0.7884)$ | $(0.3585)$ |
| GSD | 0.6924 | -0.6849 | -0.1009 | 0.4057 | 0.0396 | -0.1103 | -0.4801 | 0.9982 |
|  | $(0.0570)$ | $(0.0895)$ | $(0.8296)$ | $(0.3665)$ | $(0.9329)$ | $(0.7949)$ | $(0.2756)$ | $(0.0270)$ |
| GSSC | 0.0661 | -0.7009 | -0.7371 | -0.5582 | 0.3786 | -0.6621 | -0.4184 | -0.5886 |
|  | $(0.8765)$ | $(0.0793)$ | $(0.0588)$ | $(0.1928)$ | $(0.4023)$ | $(0.0736)$ | $(0.3502)$ | $(0.5994)$ |
|  |  |  |  |  |  |  |  |  |

$0.0891 ; \mathrm{P}=0.8667$ ).
There were significant negative correlations between AAWL and the second and third year growth increments ( $r=-0.5892 ; \mathrm{P}=0.0208$ ) and ( $\mathrm{r}=$ $-0.6591 ; \mathrm{P}=0.0104$ ) respectively. Growth increments from the fourth through the eighth year of life were not significantly correlated with AAWL.

Mean temperature from May through October (TEMP) was not significantly correlated with growth increments for any age groups except age 4 fish ( $\mathrm{r}=0.6544 ; \mathrm{P}=0.0401$ ). This correlation was most likely spurious. Mean annual turbidity (TURB) was not significantly correlated with growth increments for any age groups.

Correlation between first year growth increments (GIO1) and young-of-the-year bass density (YOYD) was significant only at the $6.65 \%$ level ( $\mathrm{r}=0.5456$; $\mathrm{P}=0.0665$ ). This correlation probably indicates that conditions which are favorable for survival of young-of-the-year bass are also favorable for growth. The significant correlation between second year growth increments (GI12) and YOYD ( $\mathrm{r}=-0.6213$; $\mathrm{P}=0.0413$ ) was probably due to the correlation between YOYD and AAWL ( $\mathrm{r}=0.5432$; $\mathrm{P}=0.0680$ ), since AAWL was also correlated with GI12.

Standing crop of largemouth bass (LMBSC) was not significantly correlated with growth increments for any age groups. Gizzard shad density (GSD) was not significantly correlated at the $5.0 \%$ level with growth increments for any age groups but age 7 ( $\mathrm{r}=0.9982$; $\mathrm{P}=0.0270$ ), but this correlation involved only 3 pairs of data. Positive correlation between GI01 and GSD was significant, however, at the $5.7 \%$ level. Gizzard shad density (GSD) was also significantly correlated with YOYD ( $\mathrm{r}=0.9365$; $\mathrm{P}=0.0006$ ), thereby confounding interpretation of these
correlations. Gizzard shad standing crop (GSSC) was not correlated with growth increments for any age groups at the $5.0 \%$ level of significance. Correlations for the second ( $r=-0.7009$; $P=0.0793$ ), third ( $r=$ $-0.7371 ; P=0.0588)$, and sixth $(r=-0.6621 ; P=0.0736)$ years of life did approach the $5.0 \%$ level of significance.

Lack of consistent significant correlations prevents any accurate predictions of growth increments based on these environmental factors. Multiple regression analysis was attempted but was unsuccessful because of the lack of a complete data set. It is questionable whether this type of analysis would be effective due to the interrelationships between turbidity, temperature, water level, water fluctuation, bass density, gizzard shad density, and the large variance associated with several of these variables.

Instantaneous rates of growth in length are required input for simulation of Model III. Instantaneous population growth rates ( $G_{x}$ ) were computed from mean lengths of the surviving fish of successive ages (Table 15) and the individual growth rates (G) were computed from back-calculated lengths of individual fish (Table 16). Both types of growth rates were computed in order to detect any possible sizeselective mortality or sampling bias (Ricker 1969). Ricker recommends that the best estimate of growth of individual fish (G) comes from the back-calculated lengths at the last two annuli on the scales since the estimate obtained from earlier annuli may not be representative of all fish that were alive at that time if there was any size-selective mortality. In this study the individual growth rates were not estimated in this way but differences still occurred. The population growth rates were slightly higher than individual growth rates in 38 of the 78

Table 15. Annual instantaneous population growth rates (G) for largemouth bass in Lake Carl Blackwell for vear-classes 1959 1975.

| Year <br> class | $1-2$ | - $2-3$ | $3-4$ | Interval $4-5$ | $5-6$ | 6-7 | 7-8 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 0.464420 |  |  |  |  |  |  |
| 1.974 | 0.496996 | 0.208944 |  |  |  |  |  |
| 1973 | 0.267566 | 0.243938 | 0.150369 |  |  |  |  |
| $1972^{\text {a }}$ |  |  |  |  |  |  |  |
| 1971 | 0.567815 | 0.255771 | 0.091564 | 0.085975 | 0.084954 |  |  |
| 1970 | 0.719915 | 0.252210 | 0.202871 | 0.125177 | 0.035502 | 0.048167 |  |
| 1969 | 0.650944 | 0.294552 | 0.152372 | 0.110192 | 0.132140 | 0.103780 | 0.014479 |
| 1968 | 0.667590 | 0.308667 | 0.153942 | 0.09170 ? | 0.029086 |  |  |
| 1967 | 0.724787 | 0.233521 | 0.152111 | 0.107552 | 0.084545 | 0.055097 |  |
| 1966 | 0.660937 | 0.238443 | 0.054242 | 0.099195 | 0.051234 |  |  |
| 1965 | 0.644357 | 0.259690 | 0.127083 | 0.030200 | 0.062964 | 0.041252 |  |
| 1964 | 0.582999 | 0.253312 | 0.159186 | 0.071870 | 0.055394 | 0.059293 | 0.035419 |
| 1963 | 0.612330 | 0.272002 | 0.139678 | 0.104166 | 0.040048 | 0.026505 |  |
| 1962 | 0.584209 | 0.249667 | 0.136547 | 0.077481 | 0.060140 | 0.036076 |  |
| 1961 | 0.748536 | 0.265418 | 0.158822 | 0.122284 | 0.069394 | -0.035134 |  |
| 1959 | 0.451658 | 0.285847 | 0.287940 | 0.086990 | 0.087859 | 0.105057 | 0.046892 |
| Means | 0.589671 | 0.258713 | 0.151287 | 0.092732 | 0.066105 | 0.048899 | 0.032263 |
| Std. Dev. | 0.127193 | 0.025782 | 0.054303 | 0.025727 | 0.028456 | 0.041980 | 0.016435 |

[^3]Table 16. Mean annual instantaneous individual growth rates (G) for largemouth bass in Lake Carl Blackwell, 1959-1976 (sample sizes in parentheses).

| Year <br> class | Age Interval |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | $\begin{aligned} & 0.403795 \\ & (53) \end{aligned}$ |  |  |  |  |  |  |
| 1974 | $\begin{aligned} & 0.452234 \\ & (185) \end{aligned}$ | $\begin{aligned} & 0.165514 \\ & (96) \end{aligned}$ |  |  |  |  |  |
| 1973 | $\begin{aligned} & 0.440441 \\ & (200) \end{aligned}$ | $\begin{aligned} & 0.150713 \\ & (99) \end{aligned}$ | $\begin{aligned} & 0.120008 \\ & (59) \end{aligned}$ |  |  |  |  |
| $1972^{\text {a }}$ |  |  |  |  |  |  |  |
| 1971 | $\begin{aligned} & 0.520806 \\ & (99) \end{aligned}$ | $\begin{aligned} & 0.225718 \\ & (44) \end{aligned}$ | $\begin{aligned} & 0.098816 \\ & (34) \end{aligned}$ | $\begin{aligned} & 0.095403 \\ & (30) \end{aligned}$ | $\begin{aligned} & 0.054471 \\ & (23) \end{aligned}$ |  |  |
| 1970 | $\begin{aligned} & 0.709420 \\ & (90) \end{aligned}$ | $\begin{aligned} & 0.288823 \\ & (39) \end{aligned}$ | $\begin{aligned} & 0.167583 \\ & (15) \end{aligned}$ | $\begin{aligned} & 0.101262 \\ & (9) \end{aligned}$ | $0.035502$ <br> (9) | $0.024277$ <br> (8) |  |
| 1969 | $\begin{aligned} & 0.650944 \\ & (92) \end{aligned}$ | $\begin{aligned} & 0.283352 \\ & (76) \end{aligned}$ | $\begin{aligned} & 0.152083 \\ & (36) \end{aligned}$ | $\begin{aligned} & 0.081546 \\ & \text { (7). } \end{aligned}$ | $0.036526$ <br> (2) | $0.035593$ <br> (1) | $0.014479$ <br> (1) |
| 1968 | $\begin{aligned} & 0.720192 \\ & (63) \end{aligned}$ | $\begin{aligned} & 0.297972 \\ & (63) \end{aligned}$ | $\begin{aligned} & 0.138847 \\ & (51) \end{aligned}$ | $\begin{aligned} & 0.107371 \\ & (24) \end{aligned}$ | $\begin{aligned} & 0.060694 \\ & (4) \end{aligned}$ |  |  |
| 1967 | $\begin{aligned} & 0.717071 \\ & (254) \end{aligned}$ | $\begin{aligned} & 0.302246 \\ & (35) \end{aligned}$ | $\begin{aligned} & 0.151814 \\ & (35) \end{aligned}$ | $\begin{aligned} & 0.094119 \\ & (33) \end{aligned}$ | $\begin{aligned} & 0.065480 \\ & (13) \end{aligned}$ | $\begin{aligned} & 0.033434 \\ & (5) \end{aligned}$ |  |
| 1966 | $\begin{aligned} & 0.653466 \\ & (128) \end{aligned}$ | $\begin{aligned} & 0.233878 \\ & (60) \end{aligned}$ | $\begin{aligned} & 0.196128 \\ & (8) \end{aligned}$ | $\begin{aligned} & 0.099195 \\ & (8) \end{aligned}$ | $\begin{aligned} & 0.050996 \\ & (8) \end{aligned}$ |  |  |
| 1965 | $\begin{aligned} & 0.650327 \\ & (177) \end{aligned}$ | $\begin{aligned} & 0.258302 \\ & (168) \end{aligned}$ | $\begin{aligned} & 0.127887 \\ & (104) \end{aligned}$ | $\begin{aligned} & 0.136609 \\ & \text { (3) } \end{aligned}$ | $0.062964$ <br> (3) | $\begin{aligned} & 0.072238 \\ & (2) \end{aligned}$ |  |
| 1964 | $\begin{aligned} & 0.589340 \\ & (158) \end{aligned}$ | $\begin{aligned} & 0.253312 \\ & (158) \end{aligned}$ | $\begin{aligned} & 0.148018 \\ & (150) \end{aligned}$ | $\begin{aligned} & 0.081021 \\ & (96) \end{aligned}$ | $\begin{aligned} & 0.046520 \\ & \text { (2) } \end{aligned}$ | $\begin{aligned} & 0.059293 \\ & \text { (2) } \end{aligned}$ | $\begin{aligned} & 0.029813 \\ & \text { (1) } \end{aligned}$ |
| 1963 | $\begin{aligned} & 0.612330 \\ & (98) \end{aligned}$ | $\begin{aligned} & 0.272002 \\ & (98) \end{aligned}$ | $\begin{aligned} & 0.139404 \\ & (98) \end{aligned}$ | $\begin{aligned} & 0.092304 \\ & (82) \end{aligned}$ | $\begin{aligned} & 0.042200 \\ & (47) \end{aligned}$ | $0.112765$ <br> (1) |  |
| 1962 | $\begin{aligned} & 0.579892 \\ & (26) \end{aligned}$ | $\begin{aligned} & 0.249322 \\ & (26) \end{aligned}$ | $\begin{aligned} & 0.136547 \\ & (26) \end{aligned}$ | $\begin{aligned} & 0.077716 \\ & (26) \end{aligned}$ | $\begin{aligned} & 0.0502 .20 \\ & (20) \end{aligned}$ | $\begin{aligned} & 0.042217 \\ & (10) \end{aligned}$ |  |
| 1961 | $\begin{aligned} & 0.748536 \\ & (5) \end{aligned}$ | $\begin{aligned} & 0.265790 \\ & (5) \end{aligned}$ | $\begin{aligned} & 0.158537 \\ & (5) \end{aligned}$ | $\begin{aligned} & 0.122527 \\ & (5) \end{aligned}$ | $\begin{aligned} & 0.069394 \\ & (5) \end{aligned}$ | $\begin{aligned} & 0.055570 \\ & \text { (2) } \end{aligned}$ |  |
| 1959 | $\begin{aligned} & 0.451658 \\ & (1) . \end{aligned}$ | $\begin{aligned} & 0.285847 \\ & \text { (1) } \end{aligned}$ | $\begin{aligned} & 0.287940 \\ & \text { (1) } \end{aligned}$ | $\begin{aligned} & 0.086990 \\ & \text { (1) } \end{aligned}$ | $\begin{aligned} & 0.087859 \\ & \text { (1) } \end{aligned}$ | $\begin{aligned} & 0.105057 \\ & \text { (1) } \end{aligned}$ | $\begin{aligned} & 0.047088 \\ & \text { (1) } \end{aligned}$ |
| Number | 1629 | 968 | 622 | 324 | 137 | 32 | 3 |
| Weighted Means | 0.593678 | 0.242867 | 0.138566 | 0.090425 | 0.050015 | 0.044099 | 0.030460 |
| Means | 0.593363 | 0.252342 | 0.155662 | 0.098005 | 0.055236 | 0.060049 | 0.030460 |

[^4]estimates. ' Of the remaining 40 estimates, 20 were slightly less than individual growth rates and 20 were equal to individual growth rates. In most cases the differences were probably not significant except for the 1973 year-class during the interval 1 - 2 (Tables 15 and 16). During this interval the population growth rate was 0.267566 as opposed to an individual growth rate of 0.440441 . This would probably indicate a size-selective mortality on the larger fish within this age group since these fish were approaching a size where they would begin to be exploited by anglers. Model III does not, however, accomodate sizeselective mortality within age classes. Since size-selective mortality does not appear to occur consistently in the Lake Carl Blackwe11 bass population, the omission of this factor should not cause any major errors, but users should be cautious because the use of incorrect growth rates can cause rather large errors in estimates of production (Ricker 1969).

The weighted mean lengths (Table 12) were used to derive a von Bertalanffy curve (Figure 15 and 16) to describe the general growth pattern of largemouth bass in Lake Carl Blackwell using Beverton's method (Ricker 1975:225). Age-specific annual instantaneous growth rates to be used for Model III simulation were then computed from the fitted von Bertalanffy equation and are presented in Table 17.

Parameters and correlation coefficients (r) for length-weight relationships computed for collections of largemouth bass from Lake Carl Blackwell from 1967 to 1977 were very similar (Table 18). The parameters for the Fall 1975 collection were chosen to use in the simulation of Model III since these were derived from one of the larger collections with a fair representation of most size groups.


Figure 15. (A) Walford plot of length (mm) at age t+l against length at age $t$ and (B) $\log _{e}\left(L_{\infty}-l_{t}\right)$ plotted against age using $\mathrm{L}_{\infty}=621.4 \mathrm{~mm}$ for largemouth bass from Lake Carl Blackwell, Oklahoma.


Figure 16. von Bertalanffy curve fitted to data for largemouth bass from Iake Carl Blackwell, Oklahoma, Open circles are weighted mean lengths from Table 12.

Table 17. Annual instantaneous rates of growth in total length
(G) computed from the fitted von Bertalanffy equation in Figure 16.

| Age <br> interval | $G$ |
| :---: | :---: |
| $1-2$ | 0.50927 |
| $2-3$ | 0.27063 |
| $3-4$ | 0.16947 |
| $4-5$ | 0.11457 |
| $5-6$ | 0.08094 |
| $6-7$ | 0.05879 |
| $7-8$ | 0.04351 |
| $8-9$ | 0.03262 |

Table 18. Parameters and correlation coefficients (r) for length-weight relationships, log w = log $a+b \log 1$, computed for collections of largemouth bass from Lake Carl Blackwell, Oklahoma (1967-1977).

| Collection <br> period | No. of <br> fish | a | b | r |
| :--- | :---: | :---: | :---: | :---: |
| Spring 1977 | 367 | $5.095 \times 10^{-7}$ | 3.5639 | 0.9901 |
| Fa11 1976 | 375 | $5.561 \times 10^{-6}$ | 3.1406 | 0.9858 |
| Fall 1975 | 780 | $4.531 \times 10^{-6}$ | 3.1633 | 0.9884 |
| Fa11 1974 | 178 | $5.457 \times 10^{-6}$ | 3.1420 | 0.9862 |
| August 1974 | 79 | $1.884 \times 10^{-6}$ | 3.3432 | 0.9907 |
| Spring 1974 | 987 | $4.027 \times 10^{-6}$ | 3.2012 | 0.9811 |
| Fa11 1973 | 961 | $3.712 \times 10^{-6}$ | 3.2369 | 0.9860 |
| Spring 1973 | 59 | $1.717 \times 10^{-6}$ | 3.3643 | 0.9919 |
| Fal1 1972 | 255 | $4.380 \times 10^{-7}$ | 3.6078 | 0.9862 |
| Spring 1967 | 55 | $2.699 \times 10^{-6}$ | 3.2848 | 0.9944 |

## Mortality Rates

Model III requires instantaneous annual rates of fishing mortality (F) and natural mortality (M) for each age group. Table 19 lists total (A), fishing (u), and natural (v) mortality rates that have been reported for largemouth bass from various lakes and reservoirs. These rates can be converted to instantaneous rates by the following relationships (Ricker 1975):

$$
\begin{align*}
& \mathrm{Z}=-\log _{\mathrm{e}}(1-\mathrm{A})  \tag{7.23}\\
& \mathrm{F}=\frac{\mathrm{uZ}}{\mathrm{~A}}  \tag{7.24}\\
& \mathrm{M}=\frac{\mathrm{vZ}}{\mathrm{~A}} \tag{7.25}
\end{align*}
$$

The assumption that mortality rates are constant after recruitment has been made for convenfence in analyzing largemouth bass populations (Anderson 1974a, 1974b). Bennett (1969) presents data that shows that high natural mortality among small bass is usually followed by a period of low mortality until after the fish reaches age 7 or 8 , at which time high mortality resumes. Age-specific mortality rates from Zweiacker (1972) and Clady (1970) also follow this general pattern.

Nominal simulation runs of Model III were made assuming $60 \%$ mortality ( $30 \%$ fishing, $30 \%$ natural) on all age groups except age 0 . The corresponding instantaneous rates of fishing and natural mortality would be 0.458145 , resulting in a total instantaneous mortality rate (Z) of 0.916291 . Age-specific rates were also used for simulation of the largemouth bass population of Lake Car1 Blackwell. Average agespecific mortality rates (A) from Zweiacker et al. (1973) were converted to instantaneous rates (Z) by equation 7.23. Instantaneous

Table 19. Reported annual rates of total (A), fishing (u), and natural (v) mortality for largemouth bass.

| A | u | v | Location | Source |
| :---: | :---: | :---: | :---: | :---: |
| 0.56 | 0.35 | 0.21 | Ridge L, IL | Bennett et al. (1969) |
| - | 0.416 | - | Watauga Res., TN | Chance (1955) |
| - | 0.412 | - | S. Holston Res., TN | Chance (1955) |
| - | - | 0.23 | Cub L., MI | Clady (1970) |
| - | - | 0.44 | Cub L., MI | Clady (1970) |
| - | 0.39 | - | Center Hill Res., TN | Coomer (1976) |
| 0.70 | 0.35 | 0.35 | Sugarloaf L., MI | Cooper and Latta (1954) |
| 0.42 | 0.22 | 0.20 | Whitmore L., MI | Cooper and Schafer (1954) |
| 0.37 | - | - | Beaver L., AR | Houser and Rainwater (1975) |
| 0.43 | - | - | Beaver L., AR | Houser and Rainwater (1975) |
| 0.74 | - | - | Beaver L., AR | Houser and Rainwater (1975) |
| 0.44 | - | - | Beaver L., AR | Houser and Rainwater (1975) |
| 0.34 | - | - | Beaver L., AR | Houser and Rainwater (1975) |
| 0.47 | - | - | Bull Shoals L., AR\&MO | Houser and Rainwater (1975) |
| 0.74 | - | - | Bull Shoals L., AR\&MO | Houser and Rainwater (1975) |
| 0.65 | - | - | Bull Shoals L., AR¢̣MO | Houser and Rainwater (1975) |
| 0.31 | - | - | Bull Shoals L., AR\&MO | Houser and Rainwater (1975) |
| - | 0.296 | - | Spavinaw L., OK | Jackson (1966) |
| - | 0.322 | - | L. Eucha, OK | Jackson (1966) |
| 0.56 | 0.20 | 0.36 | Clear L., CA | Kimsey (1957) |
| 0.68 | 0.20 | 0.48 | Sutherland Res., CA | LaFaunce et al. (1964) |
| 0.78 | 0.40 | 0.38 | Sutherland Res., CA | LáFaunce et al. (1964) |
| 0.73 | 0.47 | 0.26 | Sutherland Res., CA | LaFaunce et al. (1964) |

Table 19. (Continued).

| A | v | v | Location | Source |
| :--- | :--- | :--- | :--- | :--- |
| 0.55 | 0.35 | 0.20 | Sutherland Res., CA | LaFaunce et al. (1964) |
| 0.83 | 0.48 | 0.35 | Sutherland Res., CA | LaFaunce et al. (1964) |
| 0.62 | 0.15 | 0.47 | Gladstone L., MN | Maloney et al. (1962) |
| 0.24 | 0.12 | 0.12 | Browns L., WI | Mraz and Threinen (1957) |
| 0.89 | 0.40 | 0.49 | Folsom L., CA | Rawstron (1967) |
| 0.92 | 0.36 | 0.56 | Merle Collins Res., <br> CA | Rawstron and Hashagen (1972) |
| 0.71 | 0.45 | 0.26 | Merle Collins Res., <br> CA | Rawstron and Hashagen (1972) |
| 0.86 | 0.62 | 0.24 | Merle Collins Res., <br> CA | Rawstron and Hashagen (1972) |
| 0.76 | 0.65 | 0.11 | Merle Collins Res., <br> CA | Rawstron and Hashagen (1972) |
| 0.86 | 0.65 | 0.21 | Merle Collins Res., <br> CA | Rawstron and Hashagen (1972) |
| 0.47 | 0.012 | 0.458 | L. Carl Blackwe11, OK Zweiacker (1972) |  |

age-specific fishing mortality rates (F) were obtained using 7.24 and the average $u$ and $A$ for the two years presented in Table 19 for Lake Carl Blackwell (Zweiacker 1972). Natural mortality rates (M) were then obtained by subtraction. The resulting age-specific rates appear in Table 20.

## Results and Discussion

Simulation of the largemouth bass population of Lake Carl Blackwell with Model III resulted in estimates of year-class strength that were essentially the same as those for Model II. This result was expected since young-of-the-year survival was computed in the same manner. Thus Model III offers the same utility as Model II in quickly and easily predicting year-class strength. Predictions of production, yield and catch (Table 21) reflect the increase due to a large 1973 year-class that was produced during a year of rising water level. Production of largemouth bass (age I and older) in Lake Car1 Blackwell in 1968 was estimated at 415.49 kg and yield at 6.83 kg by Zweiacker (1972). The difference between these estimates and Model III predictions was due to the use of slightly different population estimates, and different growth and mortality rates. Mode1 III predictions of young-of-the-year production in 1973 (Table 22) compare favorably with the estimate made by Shirley (1975). Shirley's estimate was 5603.01 kg ( $5.77 / \mathrm{ha}$ ) and a prediction by Model III was $4329.56 \mathrm{~kg}(4.46 / \mathrm{ha})$.

It is unfortunate that we do not have more estimates of production and yield to compare with the Model III estimates. In general, I believe that Model III predictions are fairly accurate.

Table 20. Annual instantaneous rates of fishing (F), natural (M), and total (Z) mortality calculated for age groups of largemouth bass from Lake Carl Blackwell, Oklahoma.

| Age | F | M | $Z$ |
| ---: | :---: | :---: | :---: |
| I | 0.01080 | 0.38076 | 0.39156 |
| II | 0.01338 | 0.47169 | 0.48507 |
| III | 0.01152 | 0.40611 | 0.41763 |
| IV | 0.01600 | 0.56382 | 0.57982 |
| V | 0.02706 | 0.95376 | 0.98082 |
| VI | 0.04481 | 1.57970 | 1.62451 |
| VII | 0.07297 | 2.57211 | 2.64508 |
| VIII | 0.07297 | 2.57211 | $2.64508^{a}$ |

a Assumed to be the same as age VII.

Table 21. Gross production, yield and catch of largemouth bass (age I and older) in Lake Carl Blackwell, 1968-1977, as predicted by Model III.

| Year | Area ha | Production |  | Yield |  | Catch |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | kg | kg/ha | kg | $\mathrm{kg} / \mathrm{ha}$ | no | no/ha |
| 1968 | 746.27 | 359.57 | 0.4818 | 13.37 | 0.0179 | 23 | 0.0308 |
| 1969 | 853.11 | 326.48 | 0.3827 | 11.13 | 0.0130 | 19 | 0.0223 |
| 1970 | 815.88 | 307.60 | 0.3770 | 9.25 | 0.0113 | 17 | 0.0208 |
| 1971 | 619.60 | 260.36 | 0.4202 | 8.94 | 0.0144 | 14 | 0.0226 |
| 1972 | 523.68 | 179.68 | 0.3431 | 8.34 | 0.0159 | 10 | 0.0191 |
| 1973 | 970.88 | 106.14 | 0.1093 | 6.55 | 0.0067 | 7 | 0.0072 |
| 1974 | 1373.55 | 8122.86 | 5.9138 | 58.59 | 0.0427 | 606 | 0.4412 |
| 1975 | 1384.07 | 9472.67 | 6.8441 | 146.20 | 0.1056 | 520 | 0.3757 |
| 1976 | 1203.17 | 7795.99 | 6.4795 | 163.74 | 0.1361 | 302 | 0.2410 |
| 1977 | - | 8445.51 | - | 236.79 | - | 493 | - |

Table 22. Predictions of gross and net production of young-of-the-year largemouth bass in Lake Carl Blackwell, 19681977, based on Model III.

| Year | Mean biomass |  | Gross production |  | Net production |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | kg | kg/ha | kg | kg/ha | kg | $\mathrm{kg} / \mathrm{ha}$ |
| 1968 | 12.85 | 0.0172 | 133.16 | 0.1784 | 7.06 | 0.0095 |
| 1969 | 13.33 | 0.0156 | 138.11 | 0.1619 | 15.41 | 0.0181 |
| 1970 | 9.50 | 0.0116 | 98.41 | 0.1206 | 3.76 | 0.0046 |
| 1971 | 4.18 | 0.0067 | 43.27 | 0.0698 | -5.22 | -0.0084 |
| 1972 | 3.02 | 0.0058 | 31.28 | 0.0597 | -6.15 | -0.0117 |
| 1973 | 417.79 | 0.4303 | 4329.56 | 4.4594 | 2555.42 | 2.6321 |
| 1974 | 38.88 | 0.0283 | 402.91 | 0.2933 | 144.12 | 0.1049 |
| 1975 | 12.19 | 0.0088 | 126.33 | 0.0913 | 36.31 | 0.0262 |
| 1976 | 518.71 | 0.4311 | 5375.39 | 4.4677 | 835.57 | 0.6945 |
| -1977 | 415.53 | - | 4306.19 | - | 428.65 | - |

## Sensitivity Analysis

Seventy-five 10-year simulations of Model III were run with various input parameters and initial state variables adjusted by a $10 \%$ increase or decrease to evaluate the sensitivity of Model III output to these changes. Nominal simulation of Model III, which served as a control, was run assuming $60 \%$ mortality ( $30 \%$ fishing, $30 \%$ natural), a stable age structure starting with 2000 age I bass, and constant survival of young-of-the-year. The data deck for nominal simulation is listed in Appendix G. Catch (numbers), yield (kg) and gross production (kg) of the stock (ages 1 through k) were summed over the 10-year period and the sensitivity of these outputs to variation in any given input was calculated by equation 5.4 Results of sensitivity analysis of catch, yield and gross production are presented in Tables 23, 24, and 25 , respectively. Sensitivity to variations in maturity of age II and III was also evaluated (Table 26).

A11 three output responses were most sensitive to variations in $Z_{0}$, instantaneous mortality from egg to age $I$, and $B F E C$, the exponent of the length-fecundity relationship. Sensitivity to $Z_{o}$ corroborates the results of sensitivity analysis of Model I. Accurate estimates of survival from egg to age I are necessary to predict not only population trends but also catch, yield and production. It is also important that we have aqcurate estimates of the exponent for the length-fecundity relationship although this may not be as critical as it appears from the sensitivity analysis. A change in the exponent accompanied by an appropriate change in the constant in this relationship may still give reasonable estimates of fecundity. In the sensitivity analysis the

Table 23. Sensitivity of cumulative catch (numbers) to variations in initial population size and input parameters for a 10-year simulation of Model III.

| Adjusted parameter | +10\% |  | -10\% |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Catch | Sensitivity | Catch | Sensitivity |
| Nominal | 9706.5 | - | 9706.5 | - |
| $\mathrm{N}_{1}$ | 9966.0 | 0.0267 | 9447.0 | -0.0267 |
| $\mathrm{N}_{2}^{1}$ | 9941.0 | 0.0245 | 9472.1 | -0.0241 |
| $\mathrm{N}_{3}$ | 9940.2 | 0.0241 | 9472.7 | -0.0241 |
| $\mathrm{N}_{4}$ | 9841.4 | 0.0139 | 9571.2 | -0.0139 |
| $\mathrm{N}_{5}^{4}$ | 9770.8 | 0.0066 | 9641.5 | -0.0067 |
| ${ }^{\mathrm{N}} 6$ | 9734.3 | 0.0029 | 9678.1 | -0.0029 |
| $\mathrm{N}_{7}$ | 9721.1 | 0.0015 | 9691.3 | -0.0016 |
| - $\mathrm{N}_{8}$ | 9710.2 | 0.0004 | 9702.3 | -0.0004 |
| $\mathrm{G}_{1}^{8}$ | 10785.8 | 0.1112 | 8840.8 | -0.0892 |
| $\mathrm{G}_{2}^{1}$ | 10418.4 | 0.0733 | 9080.7 | -0.0645 |
| $\mathrm{G}_{3}$ | 9968.1 | 0.0270 | 9460.4 | -0.0254 |
| $\mathrm{G}_{4}^{3}$ | 9799.4 | 0.0096 | 9616.4 | -0.0093 |
| $\mathrm{G}_{5}^{4}$ | 9737.5 | 0.0032 | 9675.6 | -0.0032 |
| $\mathrm{G}_{6}^{5}$ | 9715.7 | 0.0009 | 9696.8 | -0.0010 |
| $\mathrm{G}_{7}^{6}$ | 9708.6 | 0.0002 | 9703.9 | -0.0003 |
| G 8 | 9706.5 | 0.0000 | 9706.2 | 0.0000 |
| $\mathrm{F}^{8}$ | 9721.7 | 0.0016 | 9673.7 | -0.0034 |
| $\mathrm{F}_{2}^{1}$ | 9457.5 | -0.0256 | 9968.6 | 0.0270 |
| $\mathrm{F}_{3}$ | 9529.8 | -0.0182 | 9892.3 | 0.0191 |
| $\mathrm{F}_{4}$ | 9605.4 | -0.0104 | 9812.1 | 0.0109 |
| $\mathrm{F}_{5}$ | 9656.4 | -0.0052 | 9758.4 | 0.0053 |
| $\mathrm{F}_{6}^{5}$ | 9685.3 | -0.0022 | 9728.1 | 0.0022 |
| $\mathrm{F}_{7}^{6}$ | 9699.6 | -0.0007 | 9713.0 | 0.0007 |
| ${ }^{\mathrm{F}} 8$ | 9707.0 | 0.0000 | 9705.4 | -0.0001 |
| $M_{1}^{8}$ | 9172.4 | -0.0550 | 10286.6 | 0.0598 |
| $\mathrm{M}_{2}^{1}$ | 9236.1 | -0.0485 | 10215.9 | 0.0525 |
| $\mathrm{M}_{3}^{2}$ | 9438.8 | -0.0276 | 9990.1 | 0.0292 |
| $\mathrm{M}_{4}^{3}$ | 9568.4 | -0.0142 | 9851.1 | 0.0149 |
| $\mathrm{M}_{5}^{4}$ | 9641.5 | -0.0067 | 9773.8 | 0.0069 |
| M 6 | 9679.3 | -0.0028 | 9734.3 | 0.0029 |
| $\mathrm{M}_{7}^{6}$ | 9697.3 | -0.0009 | 9715.6 | 0.0009 |
| $\mathrm{M}_{8}$ | 9705.9 | -0.0001 | 9706.5 | 0.0000 |
| $\mathrm{Z}_{0}^{8}$ | 4259.1 | -0.5612 | 34154.7 | 2.5187 |
| AFEC | 11052.5 | 0.1387 | 8473.5 | -0.1270 |
| BFEC | 284395.6 | 28.2995 | 2445.0 | -0.7481 |
| AWTLEN | 9706.3 | 0.0000 | 9706.2 | 0.0000 |
| BWTLEN | 9706.3 | 0.0000 | 9706.2 | 0.0000 |

Table 24. Sensitivity of cumulative yield (kg) to variations in initial population size and input parameters for a 10 -year simulation of Model III.

| Adjusted parameter | +10\% |  | -10\% |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Yield | Sensitivity | Yield | Sensitivity |
| Nominal | 2402.7 | - | 2402.7 | - |
| $\mathrm{N}_{1}$ | 2459.2 | 0.0235 | 2346.1 | -0.0236 |
| $\mathrm{N}_{2}$ | 2463.8 | 0.0254 | 2341.4 | -0.0255 |
| $\mathrm{N}_{3}^{2}$ | 2462.2 | 0.0248 | 2343.1 | -0.0248 |
| $\mathrm{N}^{3}$ | 2437.6 | 0.0145 | 2367.7 | -0.0146 |
| $\mathrm{N}_{5}^{4}$ | 2419.5 | 0.0070 | 2385.8 | -0.0070 |
| ${ }_{N}^{\mathrm{N}} 6$ | 2409.9 2406.5 | 0.0030 0.0016 | 2395.4 2398.4 | -0.0030 |
| N <br> N | 2406.5 | 0.0016 0.0004 | 2398.4 2401.7 | -0.0016 -0.0004 |
| $\mathrm{F}_{1} 8$ | 2920.2 | 0.0004 0.2154 | 2011.3 | -0.0004 -0.1629 |
| $\mathrm{G}_{2}$ | 2656.9 | 0.1058 | 2183.1 | -0.0914 |
| $\mathrm{G}_{3}^{2}$ | 2497.4 | 0.0394 | 2314.1 | -0.0369 |
| $\mathrm{G}_{4}$ | 2436.9 | 0.0142 | 2369.6 | -0.0138 |
| $\mathrm{G}_{5}^{4}$ | 2414.4 | 0.0049 | 2391.2 | -0.0048 |
| $\mathrm{G}_{6}^{5}$ | 2406.3 | 0.0015 | 2399.0 | -0.0015 |
| $\mathrm{G}_{7}^{6}$ | 2403.7 | 0.0004 | 2401.7 | -0.0004 |
| $\mathrm{F}_{8}$ | 2402.7 | 0.0000 | 2402.5 | -0.0001 |
| $\mathrm{F}_{1}$ | 2322.6 | -0.0333 | 2487.6 | 0.0353 |
| $\mathrm{F}_{2}$ | 2339.4 | -0.0263 | 2468.9 | 0.0276 |
| $\mathrm{F}_{3}$ | 2376.3 | -0.0110 | 2429.7 | 0.0112 |
| $\mathrm{F}_{4}$ | 2393.7 | -0.0037 | 2411.9 | 0.0038 |
| $\mathrm{F}_{5}^{4}$ | 2400.5 | -0.0009 | 2404.9 | 0.0009 |
| $\mathrm{F}_{6}$ | 2403.0 | 0.0001 | 2402.3 | -0.0002 |
| $\mathrm{F}_{7}^{6}$ | 2403.8 | 0.0004 | 2401.4 | -0.0005 |
| $\mathrm{F}_{8}$ | 2404.3 | 0.0007 | 2400.9 | -0.0007 |
| M ${ }_{1}$ | 2276.8 | -0.0524 | 2539.3 | 0.0568 |
| $\mathrm{M}_{2}^{1}$ | 2276.6 | -0.0525 | 2539.5 | 0.0569 |
| $\mathrm{M}_{3}^{2}$ | 2323.7 | -0.0329 | 2486.4 | 0.0348 |
| $\mathrm{M}_{4}^{3}$ | 2359.6 | -0.0179 | 2447.9 | 0.0188 |
| $\mathrm{M}_{5}^{4}$ | 2381.6 | -0.0088 | 2424.6 | 0.0091 |
| $\mathrm{m}_{6}^{5}$ | 2393.6 | -0.0038 | 2412.1 | 0.0039 |
| $\mathrm{M}_{7}^{6}$ | 2399.4 | -0.0014 | 2406.1 | 0.0014 |
| M | 2402.2 | -0.0002 | 2403.1 | 0.0002 |
| $\mathrm{z}^{8}$ | 1300.1 | -0.4589 | 6829.4 | 1.8424 |
| AFEC | 2663.0 | 0.1083 | 2160.8 | -0.1007 |
| BFEC | 45960.7 | 18.1288 | 902.3 | -0.6245 |
| AWTLEN | 2402.6 | 0.0000 | 2402.6 | 0.0000 |
| BWTLEN | 2921.6 | 0.2160 | 1997.7 | -0.1686 |

Table 25. Sensitivity of cumulative gross production (kg) to variations in initial population size and input parameters for a $10-y e a r$ simulation of Model III.

| Adjusted parameter | +10\% |  | -10\% |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Gross production | Sensitivity | Gross production | Sensitivity |
| Nominal | 4032.9 | - | 4032.9 | - |
| $\mathrm{N}_{1}$ | 4136.6 | 0.0257 | 3929.2 | -0.0257 |
| $\mathrm{N}_{2}$ | 4135.0 | 0.0253 | 3930.8 | -0.0253 |
| $\mathrm{N}_{3}$ | 4130.9 | 0.0243 | 3934.9 | -0.0243 |
| $\mathrm{N}_{4}$ | 4088.7 | 0.0138 | 3977.1 | -0.0138 |
| $\mathrm{N}_{5}$ | 4059.2 | 0.0065 | 4006.6 | -0.0065 |
| ${ }_{N}^{\mathrm{N}} 6$ | 4044.2 | 0.0028 | 4021.6 | -0.0028 |
| $\mathrm{N}_{7}^{6}$ | 4038.8 | 0.0015 | 4026.9 | -0.0015 |
| ${ }^{\mathrm{N}} 8$ | 4034.5 | 0.0004 | 4031.0 | -0.0005 |
| ${ }^{\text {G }}$ | 5195.9 | 0.2884 | 3177.3 | -0.2122 |
| $\mathrm{G}_{2}$ | 4584.1 | 0.1367 | 3563.3 | -0.1164 |
| $\mathrm{G}_{3}$ | 4236.9 | 0.0506 | 3843.0 | -0.0471 |
| ${ }_{4}^{4}$ | 4106.6 | 0.0183 | 3962.0 | -0.0176 |
| $\mathrm{G}_{5}$ | 4058.3 | 0.0063 | 4008.1 | -0.0061 |
| G 6 | 4041.1 | 0.0020 | 4024.8 | -0.0020 |
| ${ }_{G}^{6}$ | 4035.3 | 0.0006 | 4030.5 | -0.0006 |
| ${ }^{\text {G }}$ | 4033.4 | 0.0001 | 4032.1 | -0.0002 |
| $\mathrm{F}_{1}$ | 3801.5 | -0.0574 | 4284.7 | 0.0624 |
| $\mathrm{F}_{2}$ | 3825.8 | -0.0514 | 4257.6 | 0.0557 |
| $\mathrm{F}_{3}$ | 3916.5 | -0.0289 | 4156.3 | 0.0306 |
| $\mathrm{F}_{4}$ | 3974.5 | -0.0145 | 4094.2 | 0.0152 |
| $\mathrm{F}_{5}^{4}$ | 4006.2 | -0.0066 | 4060.8 | 0.0069 |
| $\mathrm{F}_{6}$ | 4022.0 | -0.0027 | 4044.3 | 0.0028 |
| $\mathrm{F}_{7}^{6}$ | 4029.3 | -0.0009 | 4036.6 | 0.0009 |
| $\mathrm{F}_{8}$ | 4032.8 | 0.0000 | 4032.7 | 0.0000 |
| $\mathrm{M}_{1}^{8}$ | 3801.5 | -0.0574 | 4284.7 | 0.0624 |
| $\mathrm{M}_{2}^{1}$ | 3825.8 | -0.0514 | 4257.6 | 0.0557 |
| $\mathrm{M}_{3}$ | 3916.5 | -0.0289 | 4156.3 | 0.0306 |
| $\mathrm{M}_{4}$ | 3974.5 | -0.0145 | 4094.2 | 0.0152 |
| $\mathrm{M}_{5}^{4}$ | 4006.2 | -0.0066 | 4060.8 | 0.0069 |
| $\mathrm{M}_{6}^{5}$ | 4022.0 | -0.0027 | 4044.3 | 0.0028 |
| $\mathrm{M}_{7}^{6}$ | 4029.3 | -0.0009 | 4036.6 | 0.0009 |
| M | 4032.8 | 0.0000 | 4032.7 | 0.0000 |
| Z | 1900.7 | -0.5287 | 13278.8 | 2.2926 |
| AFEC | 4552.5 | 0.1288 | 3554.9 | -0.1185 |
| BFEC | 103749.9 | 24.7259 | 1173.2 | -0.7091 |
| AWTLEN | 4032.6 | -0.0001 | 4032.6 | -0.0001 |
| BWTLEN | 5213.8 | 0.2928 | 3114.0 | -0.2278 |

Table 26. Sensitivity of cumulative catch (numbers), yield (kg) and gross production (kg) to variation in maturity for a l0-year simulation of Model III.

| Adjusted <br> maturity | Catch | Sensitivity | Yield | Sensitivity | Gross <br> production | Sensitivity |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Nominal | 9706.5 | - | 2402.7 | - | 4032.9 | - |
| $20 \%$ Age II | 11020.9 | 0.1354 | 2646.8 | 0.1058 | 4540.3 | 0.1258 |
| $40 \%$ Age II | 12559.0 | 0.2939 | 2949.6 | 0.2276 | 5131.2 | 0.2723 |
| $60 \%$ Age II | 14359.1 | 0.4793 | 3287.1 | 0.3681 | 5819.5 | 0.4430 |
| $80 \%$ Age II | 16464.3 | 0.6962 | 3675.5 | 0.5297 | 6620.5 | 0.6416 |
| $0 \%$ Age II |  |  |  |  |  |  |
| and |  |  |  |  |  |  |
| $20 \%$ Age III | 6146.8 | -0.3667 | 1690.6 | -0.2964 | 2644.8 | -0.3442 |
| $40 \%$ Age III | 6895.8 | -0.2896 | 1844.0 | -0.2325 | 2939.0 | -0.2712 |
| $60 \%$ Age III. | 7733.5 | -0.2033 | 2013.0 | -0.1622 | 3266.5 | -0.1900 |
| $80 \%$ Age III | 8667.5 | -0.1070 | 2198.8 | -0.0849 | 3630.2 | -0.0998 |

exponents were adjusted without any change in the constant and appeared to yield unreasonable fecundity estimates. To test this hypothesis, four 10-year simulations were run using the parameters derived for each author's fecundity data individually (Table 8). None of the sensitivity values exceeded 0.01 and thus there was little difference in the predictions of catch, yield and production based on the different parameter values for the length-fecundity relationship (Table 27).

In general, the three output responses were most sensitive to variations in growth rates, fishing and natural mortality rates, and initial population sizes of the younger age groups. Catch was more sensitive to variations in growth rates than to variations in mortality rates or initial population size, presumably because of increased fecundity. Yield and production were much more sensitive to variations in growth rates and BWTLEN, the exponent in the length-weight relationship, than to variations in mortality rates or initial population size. Sensitivity to variation in maturity of age II and III (Table 26) indicates the importance of accurate estimates of maturity in predicting catch, yield and production. Further research is needed to determine the variability of these parameters that Model III is most sensitive to and to understand the mechanisms controlling this variation within largemouth bass populations.

In addition to evaluating the robustness of the model, sensitivity analysis can aid in management of the system by determining the response of the system to changes in those parameters amenable to management. Of the parameters tested, fishing mortality is the only one over which we have some control. Based on the sensitivity analysis,

Table 27. Sensitivity (S) of cumulative catch (numbers), yield (kg) and gross production (kg) to the use of different parameters for the length-fecundity relationship derived for data from Clady (1970), Coomer (1976), Kelley (1962) and Olmsted (1974).

| Author | Catch | S | Yield | S | Gross <br> production | S |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Nominal | 9706.5 | - | 2402.7 | - | 4032.9 |  |
| Clady (1970) | 9803.3 | 0.0100 | 2419.1 | 0.0068 | 4069.0 | 0.0090 |
| Coomer (1976) | 9695.0 | -0.0012 | 2401.0 | -0.0007 | 4028.8 | -0.0010 |
| Kelley (1962) | 9703.8 | -0.0003 | 2402.2 | -0.0002 | 4031.9 | -0.0002 |
| O1msted (1974) | 9720.5 | 0.0014 | 2404.9 | 0.0009 | 4038.1 | 0.0013 |

generally a reduction in the fishing mortality should result in an increase in catch, yield and production. These outputs are more responsive to changes in the fishing mortality rate of the younger age groups. For example, a $10 \%$ decrease in the fishing mortality rate of age group I resulted in a slight decrease (0.34\%) in the cumulative catch and increases in cumulative yield and production of 3.53 and $6.24 \%$, respectively. A $10 \%$ decrease in the fishing mortality rate of age group II resulted in increases in cumulative catch, yield and production of $2.70,2.76$ and $5.57 \%$, respectively.

## Management Applications

Model III was developed for the purpose of evaluating alternative management strategies. In this section the utility of Model III is demonstrated by evaluating the effects of implementation of a 14-inch minimum length limit.

Using the average growth rates used in nominal simulation of Model III and assuming a standard deviation ( $\sigma$ ) of 25 (a reasonable value based on data for largemouth bass from Lake Carl Blackwell), a 14-inch ( 355.6 mm ) minimum length limit would effectively eliminate all fishing mortality on age groups I and II. At the start of the simulation year the mean length of age group III is 346 mm ( 13.6 in ) and therefore includes some members under the limit and some over. The mean length of this age group during the year as determined by integration of the exponential growth function would be 377.04 mm . Based on a normal curve with $\sigma=25$, an average of $80.44 \%$ of the age group would be above the minimum size limit during the year. The fishing mortality rate (F) for this age group would be ( 0.8044 ) $\mathrm{x}(0.45814)=$
0.36853. It is assumed that under-sized fish that are caught and released do not suffer any additional mortality.

A 10 -year simulation of the implementation of a 14 -inch minimum length limit was run using the same parameters used in nominal simulation (Appendix G) except for altered fishing mortality rates for age groups I, II, and III. The implementation of this regulation resulted in a $37.6 \%$ decrease in the number of fish removed over the 10 years when compared with the nominal simulation. However, the cumulative yield in weight and gross production increased 89.9 and $340.1 \%$, respectively, over the nominal simulation. The numbers of under-sized fish that were caught and released during the 10 -year simulation were 22181, 11101, and 919, for age groups I, II, and III, respectively. Total weights of these fish were $1975.0,3241.6$, and 527.7 kg , for age groups I, II, and III, respectively.

This example has been simplified by leaving all other parameters constant to show the effect of the length limit. In a more realistic application, the fishery biologist would input population parameters for largemouth bass populations under his jurisdiction, water level and water level fluctuation data for predicting year-class strength, and could program relationships between density, water level and growth. Several simulation trials could then be made with and without various minimum length limits to determine which length limit would produce the optimum yield. In the same manner, Model III can be used to evaluate different management schemes for water level manipulation, or supplemental stocking of fingerling bass.

A possible deficiency of Model III is the lack of compensatory mechanisms for population control. At the low levels of biomass of
largemouth bass in Lake Carl Blackwe11 these compensatory mechanisms may be inoperable. However, at extremely high levels of biomass, I would hypothesize that the population is constrained by decreased growth rates or fecundity or by increased natural or fishing mortality. These relationships need to be quantified and programmed into Model III so that it will become even more generally applicable.

## CHAPTER VIII

## SUMMARY AND CONCLUSIONS

Management of largemouth bass fisheries in large reservoirs is complicated by reservoir size, multispecies interactions and a fluctuating environment. The objective of this research was to develop a computer simulation model of the largemouth bass population of Lake Carl Blackwell which would predict year-class strength, production and yield and serve as a tool for management of largemouth bass fisheries in large reservoirs.

Model I was developed to simulate population trends based on an equilibrium (stable) population. Mode1 I is an age-structured deterministic model with numbers as the only state vector, and is similar to the Leslie matrix model (Leslie 1945). Constant age-specific fecundities and survival rates are required input. Young-of-the-year survival is estimated indirectly assuming an equilibrium population and using age-specific fecundity and survival data. Sensitivity analysis of this model indicates that density of bass is most sensitive to variations in survival from egg to age $I$.

Since Model I output was most sensitive to variations in survival from egg to age $I$, data on year-class strength of largemouth bass in Lake Carl Blackwell was analyzed by simple linear and multiple linear regression to develop a predictive equation to incorporate into Model II. Multiple regression equations with water level during spawning and
water level fluctuation since the end of the previous growing season as predictor variables resolved $88.2 \%$ of the observed variation in yearclass strength and $86.76 \%$ of the variation in mortality rate from egg to age I of largemouth bass in Lake Carl Blackwe11. Model II predictions of number of age I recruits agree closely with population estimates from Lake Carl Blackwell. This model should prove to be of value in largemouth bass fishery management by enabling fishery biologists to quickly and easily predict year-class strength for any given year and hence, future population size and structure.

Mode1 III is an extension of the previous models to allow prediction of production and yield. Survival from egg to age I is calculated as in Model II. Instantaneous rates of growth, fishing and natural mortality by age group, and exponents and constants in exponential length-fecundity and length-weight relationships are required input as well as proportion of each age group that are mature and female. Output from the computer simulation, presented by age group, consists of number at start of year, mean number during year, mean total length, mean weight per fish, biomass at start of year, mean biomass during year, yield in weight and numbers, and gross and net production.

Parameters to be used in Model III are derived for the lengthfecundity relationship using data from a small lake in northern Michigan (Clady 1970), large reservoirs in Tennessee (Coomer 1976) and Arkansas (Olmsted 1974), and a stream in Maine (Kelley 1962). The resulting predictive equation was: Fecundity $=0.00045091$ Length where length is in millimeters.

Growth increment data was compiled for largemouth bass from Lake Car1 Blackwell for 1959 through 1976. Correlation analysis was per-
formed with several physical and biological parameters. There were significant negative correlations between average annual water levels and the second and third year growth increments. Lack of consistent significant correlations, however, prevented the incorporation of these findings into the computer simulation model.

A von Bertalanffy equation was fitted to data for largemouth bass from Lake Car1 Blackwell and annual instantaneous rates of growth computed from the fitted equation.

Model III predictions of year-class strength also agree closely with population estimates. Predictions of production and yield compare favorably with estimates by Zweiacker (1972) and Shirley (1975). Sensitivity analysis of Model III indicates that production, yield and catch (numbers) are most sensitive to variation in mortality rate from egg to age I. Catch was more sensitive to variations in growth rate than to variations in mortality rates or initial population size of age I and older bass, presumably because of increased fecundity. Yield and production were much more sensitive to variations in growth rates and the exponent in the length-weight relationship than to variations in mortality rates or initial population size of age $I$ and older bass.

The management potential of Model III is demonstrated by simulating the population and fishery with a 14-inch minimum length limit. Model III should also be useful for evaluating different management schemes for water level manipulation or supplemental stockings of fingerling bass.

This research was intended to be a beginning rather than an end of an attempt to develop a methodology for predicting the consequences of proposed management strategies prior to implementation. I recommend
that three areas need to be investigated if we are to continue to build on our predictive capabilities.
(1) The relationships between density, growth and fishing and natural mortality:
(2) The dynamics of prey populations in reservoirs and their relationships with the predator stocks. Simultaneous simulations of predator and prey populations in reservoirs would be extremely useful in the management and understanding of these ecological systems.
(3) Testing the validity of Model III predictions of year-class strength, production and yield of largemouth bass populations in other reservoirs.

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APPENDICES

## APPENDIX A

MODEL I COMPUTER PROGRAM LISTING



APPENDIX B

MODEL I SAMPLE OUTPUT



APPENDIX C

MODEL II COMPUTER PROGRAM LISTING


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AND WAIER LEVEI.
GR.JWING SEASUN.
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IHE SAME IIME: DATA IS ARRANGLU WV CARDS AS FJLIJWS:
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-ARO 2 TITLE ALPHANUMEKIC COLS 1-80
CARD 3 FECNO REAL ARZANGEJ I V G-COLUMN FIELUS
QIGHT-JUSTIFIED 12 PUNCHE) WITH JECIMALS
CARI) 4 MATURE REAL ARKAVGED IN S-COLJIAN FIELDS
PIJVCHEO WITH JEこIMAL
ZARD 5 PERFEM REAL ARKANGED IN $S$-COLUMV FIELDS
CARD 6 REAL ARRANGED IN G-CULUMN FIELUS
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こARD 7 V REAL ARRANGEJIN S-COLUMN FIELOS
EARO 8 IYEAR INTEGER CDLS $1-4$
FLUCT REAL CIJLS G-Il IF UNAVAILABLE ENTER BLANKS
AND AVERAGE YOY SURVIVAL WILL BE USED
WATLEV REAL GOLS 13-18 IF JNAVAILABLE ENTER BLAVKS
AVO AVERADE YJY SURVIVAL WILL BE USED
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PROGRAM WILL LIST IUMBER OF FISH IN EACH AG: GROJP.
HESCRIPIINV OF VARIABLLS..
AVJSO - AVERAGI SURVIVAL RATE UF AGE O
AVGZO - AVERAJF INSTANTAVIOUS MJKTALITY RATF OF AGE O
FICNO - ARRAY CUNTAINING LSTIMATE JF NUMBER JF EGGS
PRODUCED PER.FEMALE OF EALH AGE GRUJP
FLUEI - WAIER LEVEL FLUCTUATIIN FROM ENU OF PREVIOUS GRIJWING
SEASUV TU SPAWNIN(; SEASJN JF IYEAR (FEET)
1 - SJBSCRJPT DENJTIVG AGF GRJUP
IN - I/O UNIT NUMBER FOR GARD READER
IYFAR - LABEL FOR YFAR UH SIMULATION
K - NUMRER OF AÜF GRUUPS TO BE CARKIEO IN SIMULATION
NOT COUNTING AGE $O$ MAXIMUM IS 13
LSIYKV - ARRAY CINTAININ: NUMBER OF FISH AT START OF LAST YEAR IN
EAC.H AGE GRJUP
LF - I/J UNIT NUMBER FUR LINE PRINTER
MAIJRt - ARRAY CGNTAINING PEZEENI THAT ARE MATURE BY AGE GROUP
MITRE - INTE(GER READ I RUM DAIA LARI) 5 IJ DEVJTE END OF JATA SET
N - ARRAY CONTAINING ESTIMAIE i) NUMBFR UF FISH
IV EACH AGE GRIJUP
NUYR - NUMBFK OF YLARS THE SIMJLATION WILL ZJV
MA XIMUM IS Y9
NTIJT - TUTAL NUMSER IF FISH IN AUE GROUPS 1 TU K
PtRFEM - ARRAY CONTAINING PERCENT THAT ARE FEYAIE bY AGE GRDJP
$S$ - ARRAY CONTAINING AGE-SPICIFIC SURVIVAL KATES
SJKV - SURVIVAL (JF AGF O BASED JV ENVIRUVMENIAL CONDITIONS FOR
SIMULATIUN YEAR
ifilf: - iItle fur simulaidun ylak
THIGGG - THTAL NUMBFR JF FGGS PROJJCED AT STARI OF SIMULATION YFAR
WAILIV - MEAN WATIR IEVEL DURING SPAWNING (FEFT ABOVF MSL)
, IHBUJIINES.
! $1 ; \cdot$

Y.JYSRV
fはJILS

```
0 0 0 1
n00:
0003
O004
0005
OOnt
0001
0008
0004
0010
0 0 1 1
0 0 1 2
0 0 1 3
0014
0015
0 0 1 6
\begin{tabular}{|c|c|}
\hline & \begin{tabular}{l}
C \\

\end{tabular} \\
\hline \multirow[t]{3}{*}{0001} & Subroutine egg \\
\hline & \begin{tabular}{l}
C \\
 C
\end{tabular} \\
\hline & C
C this supkuutine calcllates total egg proulcticn fer simulaticn year \\
\hline co02 & CCCMON / YCY / K, N(13), NTCT, S(13), (YEAR, FECAC(13), TCTEGG, s iAATURE(13), PERFEM(13) \\
\hline cce 3 & real n, mature \\
\hline C004 & TOPFGG \(=0.0\) \\
\hline C005 & Di 99 I \(=1, \mathrm{~K}\) \\
\hline cceb & TOTHGG = TCTFGG + (FECND(I)*N(I)*MATURE(I)*PERFEM(I)) \\
\hline noc 7 & 99 continue \\
\hline recy & return \\
\hline \multirow[t]{3}{*}{cece.} & ENO \\
\hline & C \\
\hline &  \\
\hline \multirow[t]{5}{*}{0001} & SUAROUTINE EUUILS (Avgzc) \\
\hline & \\
\hline &  \\
\hline & c. this subruutine calculates survival rate for age o based on \\
\hline & C THE METHCD CF VALGHAN ANE SAILA 1976 trans. am. fish. Soc. \\
\hline coob & CGMMCiN / YCY / K, N(13), NTOT, S(13), [YEAR, FECND(13), TCTEGG, \\
\hline & MATURE(13), PERFEM(13) \\
\hline Ccce 3 & REAL N, NTCI, MATURE \\
\hline COC4 & SUM \(=0.0\) \\
\hline CCO & FRGD \(=1.0\) \\
\hline \multirow[t]{2}{*}{coob} & \(1=\mathrm{K}-1\) \\
\hline & \({ }_{\text {C }}^{\text {C }}\) CNTER UC LCCP ti' lalculate sum of fecundity times cumulative survival \\
\hline \(\operatorname{coc} 7\) & Du \(10 \mathrm{I}=1, \mathrm{~L}\) \\
\hline CCCB & PRCD \(=\) PROC * S(I) \\
\hline cces &  \\
\hline 0010 & 10 CONTINUE \\
\hline coll & AvGzo = -(ALCG(1.0/SUM) ) \\
\hline COI2 & RETURN \\
\hline \multirow[t]{4}{*}{0013} & END \\
\hline &  \\
\hline & \(c\) \\
\hline &  \\
\hline \multirow[t]{6}{*}{C001} & SUBRLUT IAE YUYSRV (FLUCT, WATLEV, SURV, AVGZO) \\
\hline &  \\
\hline & C \\
\hline & C THIS SUBRCUTINE CALCULATES YUY SURVIVAL BASED ON \\
\hline & C ENVIRUNMENTAL CONDItIoAS during iyear \\
\hline & C \\
\hline cooz & If (hatlev .eq. 0.0\() \mathbf{z}=\) avezo \\
\hline \multirow[t]{3}{*}{cons} & IFIWATLEV .NE 0.0\() \mathrm{Z}=\) AVG20 + 221.86241513 \\
\hline & - 0.295316C2 * FLLC \\
\hline & \$ - 0.23642939 *ATLEV \\
\hline 2004 & SURV \(=\) 「. XP (-Z) \\
\hline \(\operatorname{cocs}\) & RETURA \\
\hline \multirow[t]{2}{*}{(CC6} & end \\
\hline & C C********************************************************************************* \\
\hline
\end{tabular}

APPENDIX D

MODEL II• SAMPLE OUTPUT


APPENDIX E

MODEL III COMPUTER PROGRAM LISTING

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DESCKIPTION OF VARIABLES..
A = VECTOR OF TOTAL ANNUAL MURTALITY RATES BY AGE GROUP
AFEC = PARAMETER 'A' IN FECUNDITY FSTIMATION EQUATION:
FECUNOITY = A* (LENGTH**B)
AVGP = VECTOR OF MEAN BIOMASS DURING YEAR BY AGE GROUP (KG)
AVGEO = MEAN BIOMASS OF AGE O DURING YEAR
AVGN = VECTGR OF AVERAGE NUMBER OF FISH DURING YEAR BY AGE GROUP
AVGSO = EQUILIBRIUM SURVIVAL RATE FUR AGE O
AVGTL = VECTOR OF AVERAGE TOTAL LENGTH AT START OF YEAR BY
AGE GROUP (MM)
AVGZO = EQUILIBRIUM INSTANTANEOUS MORTALITY RATE FOR AGE O
AVGW = VECTOR OF AVERAGE INIVIOUAL HEIGHTS AT START OF YEAR BY
AGE GROUP (GRAMS)
AWTLEN = PARAMETER 'A' IN WEIGHT-LENGTH PREDICTIUN EQUATIUN:
WEIGHT = A*{LENGTH**B)
B = VECTOR OF BIOMASS AT START OF YEAR BY AGE GROUP (KG)
BFEC = PARAMETER 'B' IN FECUNDITY ESTIMATION EQUATION:
FECUNDITY = A*(LENGTH**B)
BWTLFN = PARAMETER 'B' IN WEIGHT-LENGTH PREDICTION EQUATION:
WEIGHT = A* (LENGTH**B)
C = VECTOR OF CATCH IN NUMBERS BY AGE GROUP
EGGW = WEIGHT OF A SINGLE EGG (GRAMS)
EGG3 = TOTAL BIOMASS OF EGGS PRODUCED
F = VECTOR OF INSTANTANEOUS FISHING MORTALITY KATES BY AGE
GROUP
FECND = VECTOR OF NUMBER OF EGGS PER FEMALE BY AGE GROUP
FEMALE = VECTOK OF PROPORTION UF FISH THAT ARE FEMALE BY AGE GROUP
FLUCT = WATER LEVEL FLUGTUATIUN FRUM END OF PREVIUUS GROWING
SEASON TO SHAWNING SEASON UF SIMYR (FEET)
GP = VECIOR OF GROSS PRODUCTIUN DURING YEAR BY AGE GROUP (KG)
GTL = VECTOR OF INSTANTANEUUS RATF OF GROWTH IN LENGTH BY AGE
GROUP
GW = VECTOR OF INSTANTANEOUS RATE OF GROWTH IN WEIGHT BY AGE
GROUP
I = SUBSCRIPT DENCTING AGE GROUP
IN = I/O UNIT NUMBER FOR CARD READER
K = NUMBER OF AGE GROUPS TO BE CARRIEO IN SIMULATICN NOT
COUNTING AGE O. MAXIMUM IS 13.
LP = I/O UNIT NUMBER FOR LINE PRINTER
LSTYRL = VECTOR OF AVERAGE TOTAL LENGTH BY AGE GROUP FOR START
OF PREVIOUS YEAR
LSTYRN = VECTOR OF NUMBER OF FISH BY AGE GROUP FOR. START OF
PREVIOUS YEAR
LSTYRW = VECTOR OF AVERAGE INDIVIUUAL. WEIGHTS BY AGE GROUP FGR
START OF PREVIUUS YEAR
M = VECTOR OF INSTANTANECUS IVATURAL MORTALITY RATES BY AGE
GROUP
MATURE = VECTOR OF PROPORTICN UF INDIVIDUALS THAT ARE SFXIJALLY
MATURE BY AGF GROUP
MOH:- = INTEGER READ FROM LAST DATA CAPU TO DENOTE END OF EATA SET
OK NEW DATA SET TO PROCESS
N = VECTOR OF NUMBER OF FISH AJ START OF YEAR EY AGE GFGUP
NOYR = NUMBER OF YEARS THE SIMULATION WILL RUN. NAXIMUM IS 99.
NP = VECTUR OF NET INCREASE IN BIJMASS BY AGE GROUP DURING YEAR
(KG)
SIMYH = LABEL FOR YEAR OF SIMULATION
TITLE = TITLE FOR SIYULATIEN RUN

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TLG = VECTIR IIF AVERAGE TOTAL LENGTHS AT ENO UF YEAN OY AGE GRDUP * TUTAV: = AVERAG! BIOMASS JF STJCK DUKING YEAR (KU)
rijavn = average number in stuck during year
PIITE \(=\) GIUMASS DF STOCK AT START OF YEAR (KG)
TUTC = TUTAL LATCH IN NUMBERS DURING YEAK
TIJTEGG = TUTAL NUMBER CF FOGS PRODUCEO AT START BF YFAR
rITip \(=\) GRUSS PMALULTION OF STOCK UURING YEAR (KG)
TLTV = IOTAL VUMBER IN STOCK AT START UF YEAR
TLTIVP = NET INCREASE IN HICMASS OF STUCK DURINÓ YEAD (KG)
TuTY = TOTAL CATCH BY wEIGHT DUH ING YEAK (KG)
WATLEV = NHAN WATER LEVEL DUKING SPAWIIVG (HEET ABIVS MSL)
= VECTUK UF AVIRAGE IVIJVIJUAL WFIGHTS AT ENO GF YEAR
\(\begin{aligned} & \text { BY AGE GRUUP } \\ Y & \text { VECTOK OF CATCH HY WFICHT JURING YEAR SY AGE iHUUSP (KG) }\end{aligned}\)
YI:YOW = INSTANTANEUUS RATE UF GRUWTH IN WEIGHT OF AGE 0 FRUM EGG jtage Tu áge 1
Y:ZANP = NET INCREASE IN BICMASS MOUNG-OF-THE-TEAL FROM EGL STAGE PO AGE 1 (KG)
YOYGP = GHUSS PRODUCTION OF YOUNG-UF-THE-YEAK FKGM EGG STAGE TG AGE 1 (KG)
YUYSRV = SURVIVAL RATE FRR AUE O FRCM EGG STAGE IS AGE \(x\)
\(\angle O \quad=\) INSTANTANEOUS MORTALITY QATE FOR AGE \(O\) FRUM EGG STAGE TI AGE 1 \(\angle\) = vectuik uf instantaineuus tural mortality rate by age group
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SuかRUUILIFS..

```
    sumyup
    YUYSUB

START

CEMMON / YOY / AVGSO, AVGLO, FECND(13), FEMALE(13), FLUCT, K, 1 MATURE(13), WATLEV, YOYSRV, Z(13), ZO, IFLAG
DIMFVSISN A(L3), \(A V G B(13), ~ A V O ̈ V(13), ~ A V G T L(13), ~ A V G W(13), ~ B(13), ~\) \(1 \quad\) C(13), F(13), GP(13), GTL(13), GW(13), LSTYFiL(13), (STYKN(13), LSTYRW(13), Y(13), N(13), ivP(13), Y(13). TIrLf(20), TL2(13), W2(13)
REAL LSTYRL, LSTYKN, LSTYRW, M, MATURE, N, NP INTLGEK SIMYR
DATA IN, LP /5,6/

1 REAS IIN. 1010 K, NOYR, TITLE
Li.1.) FIKMAI (212.1. \(\because(044)\)
\(i\)
\(C\)
\(C\)
H!AUlw's fCR OUTPUT
औ11E (Lf, 1020) TITLE, NIBYR, K

1 - I.ARGEATJUTH GASS POPULATIUN, PRGUUCTIUN. AND YIELD',


'SIMUIFIICN WILL RUN ',I2,' YEARS AND CARRY ', I2,
- ACE \(\therefore\)...UPS'd
```

C RIAU POPULATIUN PARAMETERS

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```

                READ (IN,IU30) AVGTL, AVGW, N. GTL, ". M, FEMALE, MATURE
    (03) FORMAT (13F6.0)
        READ (IN, 1040) AFEC, HFF.C. AWTLEN, BUTLEN
    1040 FORMAT (4F10.0)
    C
    C READ LNVIRONMF VIAL CONUITIUNS FUH SIMULATION YEAR
        2. READ (IN, LOSO) SIMYR, FLUCT, WATLEV
    1050 FORMAT (14,1X,FU.0,1X,FU.0)
    c
C CHECK FCR END JF SIMULATIU'N
C
It (SIMYR .F\. U) GO TO yO
C
COMPUTE INSTANTANEOUS RATE CF GRCWTH IN WEIGHT
DO 10 I=1,K
GW(1) = 8WTLEN*GTLI1)
10 CuntINUE

1. c,DAPUTF INSTANTA:NEIUS IUTAL MURTALITY RATES. AVERAGE NUMBER PER AGE
griuu aND BIOMASS per age griJup
C
0(:20 I =1,k
L(I) = F(I)+M(1)
A(1) = 1-EXP(-Z(I))
AVGN(I) = N(I)*A(I)/Z|I)
H(I) = N(I)*AVGW(1)*0.001
2O LCNTINUE
C
Glmpute recundity
OU 30 I =1,K
FECNO(I) = AFFC*(AVITIL(I)**BFEC)
30 LCNTINUE
C
C COMPUIF THTAL FGG PRODUCTILN
MWになに=0.0
1)0 40 1 = 1,k
1HT:SG= IJ!:OG + (N(I)*FECIND(I)*MATURE(I)*FEMALF(I))
*O CiNTINUF
C COMPUTE GROUS AVD 'NET PRUDUCTICN OF YOUNG-OF-THE-YEAR
FGGN=0.0012
EGGB = EGG**TOTEGG*0.001
CALL YOYSUB
YOYGw = AL OG(AVIjW(1)/EGGW)
c
C CCMPUTE AVERAGE gIOMASS IF AT,E O DURING yEAR
If (ruYgW - LO) 41,42,43
41 AVG\&O = EGGG*(1-EXP(-(\angleO-YOYGW)))/(ZO-YOYGW)
(G) I!) }4
4) AV(%)
```

\begin{tabular}{|c|c|c|}
\hline \multicolumn{3}{|l|}{0079 WRITE (LP, 1085) (FECND(I), I=1,K)} \\
\hline CC80 & 1085 & FORMAT (1HO, \({ }^{\text {(FECUNDITY'.11X,13F8.1) }}\) \\
\hline 0081 & 1085 & WRITE (LP, 1086 ) (MATURE(I), I=1,K) \\
\hline 0082 & \multirow[t]{2}{*}{1086} & FORMAT (1HO, 'MATURITY', 12X, 13F8.5) \\
\hline 0083 & & WRITE (LP,1087) (FEMALE(I), \(1=1, \mathrm{~K}\) ) \\
\hline CC 84 & \multirow[t]{2}{*}{1087} &  \\
\hline 0085 & & WRITE (LP, 1088) AFEC, BFEC, AWTLEN, UWTLEN \\
\hline CC 86 & \multirow[t]{2}{*}{1088} &  \\
\hline & & F10.8,5X, 'BWTLEN \(=1,{ }^{\text {P }} 10.81\) \\
\hline C08 7 & \multirow[t]{3}{*}{\[
1089
\]} & WRITE (LP, 1089) AVGZO, AVGSO, YOYSRV, FLUCT, WATLEV \\
\hline CO88 & & FORMAT (1HO, 'AVGZO \(=\) ',F11.8.5X, 'AVGSO \(=1, F 11.9 .5 \mathrm{C}\), 'YOYSKV \(=\) ', \\
\hline & &  \\
\hline &  & te headings and simulaticn results \\
\hline CC8y & \multicolumn{2}{|l|}{WRITE (LP,1090)} \\
\hline \multirow[t]{8}{*}{co90} & \multirow[t]{8}{*}{1090} &  \\
\hline & &  \\
\hline & &  \\
\hline & & 'DURING', 3x, 'TOTAL',4X,'PER', 4X, 'AT START', 3x, 'DURING', 5x, \\
\hline & &  \\
\hline & & 'YEAR', 4 X , 'LENGTh', 3 X , 'FISH', 3x, 'OF YEAR', 5 X , 'YEAR', 5 X , \\
\hline & &  \\
\hline & &  \\
\hline \multirow[t]{3}{*}{\(\operatorname{cog} 1\)} & \multirow[t]{3}{*}{} & WRITE (LP,ILOO) TOTEGG, EGGW, EGGb, AVGBO, YOYGP, YCYNP, \\
\hline & & (I,N(I), AVGN(I), AVGTLII), AVGW(I), B(I), AVGB(I), Y(I), \\
\hline & & \(C(I), G P(I), N P(I), I=I, K)\) \\
\hline \multirow[t]{3}{*}{\(\operatorname{ccs} 2\)} & \multirow[t]{3}{*}{1100} & FORMAT (1HO, 17X, '0', 3X,F10.0,20X,F6.4,2F10.2,20X, 2F 10.2,/\%, \\
\hline & & 117X, 12, 2X, 2F10.0,2X,F5.0,4X,F5.0,1X,3F10.2,F10.0,2F10.2, \\
\hline & & /1) \\
\hline \multirow[t]{2}{*}{ccs 3
01194} & & WRITE (LP, 111J) K, TOTN, TOTAVN,TOTB, TOTAVB, TOTY, TOTC, TOTGP, TOTNP \\
\hline & 1110 & FORMAT (1H0,15X, 1-1,12.F11.0.F10.0,17X,3F10.2,F10.0.2F10.2,//, \\
\hline \multirow[t]{3}{*}{} & \multicolumn{2}{|l|}{\(\left.1130{ }^{\circ}=1\right)\) ) 1} \\
\hline & c & \\
\hline & C GENERATE AG & NERATE AGE DISTRIBUTIUN, AGE-SPECIFIC LENGTHS AND WEIGHTS FUR SIMYR+1 \\
\hline CCs. & \multicolumn{2}{|r|}{DC \(70 \mathrm{I}=1, \mathrm{~K}\)} \\
\hline 0096 & \multicolumn{2}{|r|}{LSTYRN(I) = N(I)} \\
\hline CCS 7 & \multicolumn{2}{|r|}{LSTYRLII) = AVGTLII)} \\
\hline CCs 8 & \multicolumn{2}{|r|}{LSTYRW(I) = AVGW(I)} \\
\hline c099 & \multirow[t]{2}{*}{} & 70 CONTINUE \\
\hline C100 & & DO \(80 \mathrm{I}=1 . \mathrm{K}\) \\
\hline C101 & \multicolumn{2}{|r|}{N(I+1) \(=\) LSTYRN(I)*EXP(-Z(I))} \\
\hline 0102 & \multicolumn{2}{|r|}{AVGTL(I+1) = LSTYRL(I)*EXP(GTL(I))} \\
\hline 0103 & \multicolumn{2}{|r|}{\(\operatorname{AVGW}(1+1)=\operatorname{LSTYRW}(1) * \operatorname{EXP}(\mathrm{GW}(1))\)} \\
\hline 0104 & \multicolumn{2}{|r|}{80 CONTINUE.} \\
\hline 0105 & \multicolumn{2}{|r|}{N(1) = TOTESG*YOYSRV} \\
\hline \multicolumn{3}{|l|}{C106 GOTO?} \\
\hline & \multicolumn{2}{|l|}{c} \\
\hline & \multicolumn{2}{|l|}{c WरITE 'ENO OF SIMULATION'} \\
\hline & \multicolumn{2}{|l|}{c} \\
\hline C107 & \multicolumn{2}{|l|}{SO WRITE (LP,1120)} \\
\hline \multirow[t]{2}{*}{0108} & 1120 &  \\
\hline & \multicolumn{2}{|r|}{1130 ('*リ)} \\
\hline C1C9 & \multicolumn{2}{|l|}{READ (IN, 1130\()\) MORE} \\
\hline 0110 & \multicolumn{2}{|l|}{1130 FORMAT (II)} \\
\hline 0111 & \multicolumn{2}{|r|}{IF (MORE - EQ. 1 ) GC TO 1} \\
\hline 0112 & \multicolumn{2}{|l|}{} \\
\hline 0113 & \multicolumn{2}{|r|}{END} \\
\hline
\end{tabular}


APPENDIX F

MODEL III SAMPLE OUTPUT



APPENDIX G

INPUT DATA FOR NOMINAL SIMULATION OF MODEL III

00000000011111111112222222222333333333344444444445555555555666666666677777777778 1234567890123456789012345678901234557890123 '55 5739512345678901234557890123 ' 567390
CARE 0001 0002 0003 0004 0005
0006
0007 0008 3009
0010
0011
0012196
00131969
00141970
00151971
00161972
\(0917 \quad 1973\)
09181974
\(019 \quad 1975\)
00201976
00211977
03220000
0023
810
155
3
200
.5092
.4531
.4381
.5000
0.000
0.000
1968
1969
1970
1971
1972
1973
1974
1975
1976
1977
0000
0

\author{
VITA \({ }^{2}\) \\ Donald John Orth \\ Candidate for the Degree of \\ Master of Science
}

\section*{Thesis: DEVELOPMENT OF A COMPUTER SIMULATION MODEL OF LARGEMOUTH BASS POPULATION DYNAMICS}

Major Field: Zoology
Biographical:
Personal Data: Born on December 5, 1954 in Berwyn, Illinois, the son of William John and Marie Joanne Orth; married Martha Lee Pennington, May 30, 1976.

Education: Graduated from St. Rita of Cascia High School, Chicago, Illinois, June 1972; received the Bachelor of Science degree in Environmental Biology from Eastern Illinois University, Charleston, Illinois, August, 1975; received the Master of Science degree in Zoology, with a specialty in Fisheries, from Oklahoma State University, Stillwater, Oklahoma, December, 1977.

Professional Experience: Technical Assistant with the Illinois Natural History Survey, Urbana, Illinois, May through August, 1975; Graduate Research Assistant with the Oklahoma Cooperative Fishery Research Unit, Oklahoma State University, Stillwater, Oklahoma, January, 1976 to present.

Professional Affiliations: American Fisheries Society; Ecological Society of America; American Association for the Advancement of Science; American Society of Limnology and Oceanography; Oklahoma Academy of Science; Beta Beta Beta Biological Society; Phi Sigma Society.```


[^0]:    $1_{\text {Water }}$ levels were obtained courtesy of the Hydraulics Research Laboratory, U.S. Department of Agriculture.

[^1]:    ... a deliberate melding of biological, economic, social and political values designed to produce the maximum benefit to society from stocks that are sought for human use, taking into account the effect of harvesting on dependent or associated species (p. 85).

[^2]:    ${ }^{\mathrm{a}} \mathrm{F}$ ratio (Mean square due to regression divided by mean square due to residual variation) used to test the null hypothesis $H_{o}$ : $b=0$.
    $\mathrm{b}_{\text {Probability }}$ of a greater value of F .

[^3]:    ${ }^{\mathrm{a}}$ Year-class failure

[^4]:    ${ }^{a}$ Year-class failure

