# DEVELOPMENT OF A RAPID BIOASSESSMENT 

# PROTOCOL FOR SAMPLING FISHES 

## IN LARGE PRAIRIE RIVERS

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# DEVELOPMENT OF A RAPID BIOASSESSMENT PROTOCOL FOR SAMPLING FISHES IN LARGE PRAIRIE RIVERS 

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#### Abstract

We used seining and hoop netting to collect fishes at 15 sites in five large prairie rivers in Oklahoma to (1) determine the minimum amount of effort needed to detect the majority of fish species at a sample site and (2) examine the selectivity of fish species detected by the two gear types. Analysis of similarities of the fishes collected in six different habitat types identified two distinct habitat types based on fish species composition: shallow/backwater (SBW) habitat (depth $\leq 0.75 \mathrm{~m}$ ) and deep/non-wadeable (DNW) habitat (depth $>0.75 \mathrm{~m}$ ). We estimated that between 6 and $10(\bar{x}=8)$ SBW habitats andbetween 1 and 6 $(\bar{x}=4)$ DNW habitats at each sample site were needed to detect the majority of fish species during a sampling event. Minimum sampling distance needed to encounter the minimum number of habitats ranged from 400 m to 1600 m and averaged 887 m . Gear evaluation showed seining captured more species per unit effort than hoop netting (3.6 and 1.4 respectively); however, hoop netting captured significantly larger fish ( $527 \mathrm{~mm} ; P<0.001$ ) than seining ( 42 mm ). Based on these collections, we present recommendations for sampling fish assemblages in large prairie rivers in the southern Great Plains to aid in the rapid bioassessment and monitoring of fish assemblages in large prairie rivers.


## Introduction

Rapid bioassessment protocols to assess the biotic integrity of riverine environments have become an important component of water resources management. Methods for rapidly sampling biological species assemblages in small wadeable streams and rivers have been widely investigated and successfully implemented by biomonitoring agencies across the U.S. (Barbour et al. 1999). However, there has been relatively little research into the development of methods for rapidly assessing large non-wadeable rivers, particularly large prairie rivers. The paucity of large river sampling protocols can be attributable to the difficulty in characterizing fish assemblages in these environments and the lack of relatively unimpaired reaches for estimating reference conditions (Lyons et al. 2001). Fish sampling protocols that have been developed for large rivers are generally designed for rivers in the midwestern and western U.S. (OhioEPA 1989; Emery et al. 2003; Yoder and Kulik 2003) or target specific species (USGS 1998). Standardized fish sampling protocols for large prairie rivers in the interior Great Plains of the U.S. have focused primarily on the assessment and monitoring of targeted fish populations (i.e., paddlefish Polyodon spathula and pallid sturgeon Scaphirhynchus albus) in impounded sections of the Missouri River (USGS 1998) and therefore are not necessarily applicable to other large prairie rivers within the Great Plains.

The Great Plains Ecoregion in the U.S. (Omernik 1987) is dissected by three major river systems that flow from west to east and drain into the Mississippi River: the Missouri, Arkansas, and Red rivers and their tributaries.

Because of the surficial geology and solution processes within the soil and groundwater of the Great Plains, prairie rivers vary widely in their physical and chemical properties. These rivers are typically silt-laden, turbid, and alkaline with large fluctuations in discharge ranging from flash flooding during rain events to temporary intermittent flow during drought periods (Matthews 1988; Poff and Ward 1989; Dodds et al. 2004). However, there are distinct differences between prairie rivers in the northern and southern Great Plains. Prairie rivers of the northern Great Plains generally have more consistent flow and cobble substrates, whereas prairie rivers of the southern Great plains are characterized by irregular flow, distinct wet and dry seasons, ad smaller-sized substrates (Matthews 1988). Furthermore, southern Great Plains rivers have much greater concentrations of dissolved ions (i.e. higher conductivity) than northern rivers because southern rivers flow over Permian salt deposits from which chloride ions are leached to the surface by numerous springs and small creeks (Matthews 1988).

Prairie rivers with headwaters in the Rocky Mountains, such as the Platte and Arkansas rivers, form wide, shallow channels as they cross the Great Plains. Habitats for prairie river fishes that occur within these channels are influenced by geomorphology and hydrology. As these rivers course through this region, they cut away erosible banks and re-deposit fine sand and coarse alluvium uniformly throughout their river channels (Cross and Moss 1987). Therefore, many prairie rivers, particularly those of the southern Great Plains, have a fairly uniform sand or clay substratum (Matthews 1988). Cross and Moss (1987)recognized three
distinct habitat types in these rivers: channels of fluctuating flow and shifting sand beds, clear brooks and marshes sustained by springs and seeps, and residual pools (i.e., backwaters and side channels) that are dependent on water level. Hydrologic fluctuations in these rivers can be large resulting in harsh environmental conditions for fishes (Matthews 1998), causing some species to seek out pools where hydraulic stress is low (Statzner et al. 1988).

Prairie rivers pose unique challenges for sampling fishes because of their physical and chemical characteristics. Most fish sampling in prairie rivers has used some combination of seining, hoop netting, gillnetting, and electrofishing depending on the conductivity and turbidity of the water (Peters et al. 1989; Barfoot and White 1999; Milewski et al. 2001) Milewski et al. (2001)found that seining was most effective at sampling minnows, whereas trapnets and hoop nets were most effective at sampling larger bodiedbenthic fishes in prairie rivers of eastern South Dakota. They also found that gillnets often collected few or no fishes and electrofishing proved ineffectiveat sites with high turbidity.

Because of the high turbidity and conductivity of prairie rivers in the southern Great Plains, electrofishing is not feasible. Therefore, most fish sampling in this region has used a variety of seining and gillnetting methods. Pigg (1988), Pigg et al. (1992), Pigg et al. (1997) and Pigg et al. (1999) used standardized methods of seining for sampling fishes in the large prairie rivers of Oklahoma. Their methods included approximately 20 seine hauls of 10 m length, covering the same amount of surface area (approximately 200 m ) and the same segment of the shoreline at each site. In addition to seining, Pigg (1988) and

Pigg et al. (1992) placed a gillnet across the site for two hours. Ostrand and Wilde (2002) used a standardized method of seining to sample fishes in the Brazos River, a large prairie river in central Texas. They established15 to 20 transects, set 25 m apart, and made a total of 25 seine hauls of 5 m length at each sample site. In contrast to standardized sampling, Taylor et al. (1993) and Taylor et al. (1996) used judgmental sampling to collect fishes in several prairie rivers of the upper Red River basin. Their method included seining for 45 minutes to 1 hour at each site, and they attempted to sample all available habitats. Although several methods have been used to sample fishes in prairie rivers of the southern Great Plains, there is no single protocol that is both standardized and statistically based for sampling fish assemblages in the large prairie rivers of this region.

Determining the most efficient sampling method with the appropriate gear types, habitats to sample, and number of samples can be challenging. Statistically-based methods for developing a sampling design ensure that the data collected are appropriate for subsequentstatistical analysis and interpretation in water quality assessments (Chapman 1996). Moreover, use of a rapid, statistically valid sampling protocol allows for sampling of multiple sites in a single field season and the comparison of biological data among impaired and reference sites (OhioEPA 1989; Barbour et al. 1999). A rapid bioassessment protocol is needed to monitor the integrity of fish assemblages in large prairie rivers.

We developed a standard protocol for seining and hoop netting fish assemblages in five large prairie rivers in Oklahoma. Our objectives were to (1) determine the minimum number of samples needed to detect the majority of fish species in different habitats at a sample site, and (2) examine catchability of fish species by the two gear types. We present recommendations for the rapid bioassessment of fish assemblages in the large prairie rivers of the southern Great Plains.

## Study Area

All major rivers of the interior plains of the U.S. flow through the Great Plains ecoregion (Omernik 1987). Within this ecoregion, large prairie rivers of the northern Great Plains generally flow through the West-Central Semi-Arid Prairies and Temperate Prairies ecoregions and the prairie rivers of the southern Great Plains flow through the South-Central Semi-Arid Prairies ecoregion (Omernik 1987; Figure 1) These southern Great Plains rivers typically have shallow, braided channels with sandy substrate much of which is underlain by Permian salt beds and extensive areas of overlying gypsum (Cross and Moss 1987).

For this study, we defined large prairie rivers in Oklahoma based onthe following physico-chemical characteristics: (1) high average conductivity (>1,000 $\mu \mathrm{S} / \mathrm{cm}$ ), (2) predominately sand substrate, and (3) over half of the river channel is non-wadeable (i.e. > 1.5 m deep) and/or cannot be safely sampled using
wadeable sampling procedures (Barbour et al. 1999; OWRB 1999)during normal flow periods.

Physico-chemical characteristics were compiled from U.S. Geological Survey (USGS) gaging station data near 15 sites on the five large prairie rivers we sampled in Oklahoma (Table 1). Physico-chemical parameters varied among sites: contributing drainage area ranged from $12,398.3 \mathrm{~km}^{2}$ on the Washita River near Alex, Oklahoma, to $195,474.2 \mathrm{~km}^{2}$ on the Arkansas River near Coweta, Oklahoma, mean discharge ranged from $10.5 \mathrm{~m}^{3} / \mathrm{s}$ on the N. Canadian River near El Reno, Oklahoma, to $207.1 \mathrm{~m}^{3} / \mathrm{s}$ on the Arkansas River near Ralston, Oklahoma, specific conductance ranged from $950.4 \mu \mathrm{~S} / \mathrm{cm}$ on the Washita River near Durwood, Oklahoma, to $13,102.7 \mu \mathrm{~S} / \mathrm{cm}$ on the Cimarron River near Dover, Oklahoma, and turbidity ranged from 13.7 JTU on the Washita River near Alex, Oklahoma, to 262.5 JTU on the N. Canadian River near Wetumka, Oklahoma (Table 1).

## Methods

We measured habitat and collected fish at 15 different sample sites on five large prairie rivers in western and central Oklahoma, including three sites on each the Arkansas, Cimarron, North Canadian, Washita, and Red rivers (Figure 2). Fish were collected during periods of normal to low flow from May to October in 2003 and 2004. At each site we established 11 transects spaced 100 m apart and measured water depth and stream flow along each transect. Depth was measured using a $1.5 \mathrm{~m} \times 12.7 \mathrm{~mm}$ PVC pipe with cm increments. Current
velocity was estimated from the movement of the water about the pipe, which was calibrated with a Marsh-McBirney, Inc, Flo-Mate portable flow meter. In areas with a fast current velocity ( $>0.20 \mathrm{~m} / \mathrm{s}$ ), the water formed a " V " pattern about the pipe whereas in areas with a slow current velocity ( $<0.20 \mathrm{~m} / \mathrm{s}$ ), the water formed a "U" pattern about the pipe (Gorman and Karr 1978). Six different habitat types were identified based on these measurements and mapped along each transect. These habitat types were: shallow slow (SS), shallow fast (SF), deep slow (DS), deep fast (DF), non-wadeable (NW), and backwater (BW); Table 2).

We were unable to use electrofishing because of high conductivity (> 1000 $\mu \mathrm{S} / \mathrm{cm}$ ) at each sample site (Table 1), therefore, we used a seine ( $6.1 \mathrm{~m} \times 1.2 \mathrm{~m}$ $x 4.8 \mathrm{~mm}$ mesh) in shallow-water habitats (depth $\leq 0.75 \mathrm{~m}$ ) and large ( 0.9 m x $3.7 \mathrm{~m} \times 50.8 \mathrm{~mm}$ mesh) and small ( $0.6 \mathrm{~m} \times 2.4 \mathrm{~m} \times 25.4 \mathrm{~mm}$ mesh) hoop nets in deep-water habitats (depth $>0.75 \mathrm{~m}$ ). Sampling at each site consisted of approximately 20 seine hauls and 12 hoop net sets. We seined twice in four randomly selected SS and SF habitats and all BW habitats encountered throughout each 1000 m reach. Seine hauls were made parallel to the shoreline, with the current, for a distance of 10 m . When possible we used the river bank to trap the fish, and we attempted to seine near discrete microhabitats (e.g. cobble, bedrock) and structures (e.g. woody debris, boulder, undercut bank). We set 6 large hoop nets and 6 small hoop nets near in-stream structure and vegetation in DS, DF, and NW habitats. All nets were unbaited and allowed to fish over night and were emptied after approximately 12 hours.

To test for redundancy in fish species collected in the six habitat types, we compared the similarity in fish assemblage composition among habitat types (analysis of similarities; ANOSIM; PRIMER 5.0, Plymouth Marine Laboratory, Plymouth, UK). The ANOSIM test compares species (e.g., fish) composition between two groups (e.g., habitats) and generates an $R$-value, which is a measure of similarity ranging from 0 to 1 with 0 being completely similar and 1 being completely dissimilar (Clarke and Gorley 2001). The ANOSIM test is based on a non-parametric permutation procedure applied to a Bray-Curtis similarity matrix (Clarke and Warwick 2001). Fish assemblages were compared between all six habitat types with the groupings either being well separated ( $R>$ 0.75 ), overlapping but clearly different ( $R>0.5$ ), or barely separable ( $R<0.25$ ) based on thresholds described by Clarke and Gorley (2001).

To determine amount of sampling effort needed to detect a majority of the fish species present at each site, we created species-accumulation curves showing the cumulative increase in number of species with each added habitat. We used number of species captured and number of habitats sampled (averaged over all sample locations) to calculate the mean number of habitats needed to detect a majority of fish species in each habitat type at each sample site. We used regression analysis (SAS Institute Inc., Cary, NC) to test relationships between mean number of habitats needed and wetted width, contributing drainage area, and mean discharge at each sample site. We used analysis of variance (ANOVA) to test for differences in number of habitats needed to detect a majority of fish species among rivers. We also calculated number of samples
required to obtain a statistically valid sample size based on species-per-uniteffort (SPUE) estimates for the seine and two hoop net sizes. The equation for calculating number of samples $(N)$ needed is:

$$
N=\frac{t^{2} s^{2}}{(\bar{x} L)^{2}}
$$

where $t^{2}$ is the Student's statistic for the selected confidence level $(P=0.05), s^{2}$ is the variance of the sample, $\bar{x}$ is the sample mean, and $L$ is the allowable error of the mean (Snedecor and Cochran 1967; Fisher 1987). For our study, $t^{2}$ was 4, and we set $L$ at $0.1,0.25$, and 0.5 for comparative purposes. As a general rule, precision of sample size estimates within $10 \%(0.1)$ of the population mean are used in research studies, $25 \%$ (0.25) of the mean in management studies, 50\% (0.5) of the mean in preliminary surveys (Robson and Regier 1964; Wilde and Fisher 1996).

We determined minimum sampling distance needed at each site to detect a majority of fish species from each habitat type based on sampling effort. To determine this distance, we classified habitat types in a 10 km section of the river near each of the 15 sample sites by photointerpreting black and white digital orthophoto quad (DOQ) maps using ArcView 3.3 GIS software (Whited et al. 2002). Habitat types were distinguished on the photos by their relative darkness in the wetted portion of the river channel. Light-colored areas (i.e., pixel color ranging from white to gray) were identified as shallow water (includes SS and SF habitats) and dark-colored areas (i.e., pixel color ranging from dark gray to black) as deep water (includes DS, DF, and NW habitats). Backwater habitats were
identified as water in the river channel separated from the main channel on three or more sides (Figure 3). We divided each 10-km-river-section into 100 m transects and calculated average wetted width and percentage of each habitat type for the entire section (Figure 3) and used this to estimate the minimum sampling distance needed for each sample site. We used a regression analysis to examine the relationship between the average percentages of habitat types and average wetted width at each sample site.

Species-per-unit-effort for seining was measured as number of species caught per seine haul (approximately 10 m ) and for hoop nets as number of species caught per net night (approximately 12 hours, set over night). We used an ANOVA, on $\log _{10}(X+1)$ transformed data, to test for differences in mean length and weight of fish between gear types and net sizes. We used ANOSIM to test for similarities in fish assemblage composition detected by the different gear types and net sizes.

## Results

## Habitat Classification

The six initial habitat types showed distinct similarities in fish assemblage composition (Global $R=0.452$ ). We found that $S$ S and SF habitats were the most similar ( $R=0.026$ ), followed by DS and DF $(R=0.080)$, DS and NW $R=$ 0.087), DF and NW ( $R=0.118$ ), and SS and BW habitats $(R=0.174)$. Fish species composition differed slightly between SF and BW habitats but not enough to definitively separate these two habitats $(R=0.283)$. The other habitat
combinations were either well separated or slightly overlapping ( $R>0.700$; Table 3). Based on these results, we reduced the six initial habitat types to two, shallow/backwater (SBW) and deep/non-wadeable (DNW). Shallow/backwater habitat was characterized by having a depth $\leq 0.75 \mathrm{~m}$ regardless of current velocity and DNW habitat was characterized by having a depth $>0.75 \mathrm{~m}$ regardless of current velocity. Based on fish species composition, SBW and DNW habitats were well separated with only moderate overlap ( $R=0.719$ ).

## Sampling Effort

We sampled an average of $11.3(S D=1.80)$ SBW habitats at each site and collected an average of $13.6(S D=3.36)$ species from each. We sampled an average of $5.3(\mathrm{SD}=1.54)$ DNW habitats at each sample site and collected an average of $5.2(S D=2.43)$ species from each. We detected a majority of fish species after sampling 6 to 10 SBW habitats $(\bar{x}=8)$ and 1 to 6 DNW habitats $(\bar{x}=4)$ at each site (Table 4). We found no significant relationship between the mean sampling effort and mean wetted width $\left(R^{2}=0.1773 ; P=0.118\right)$, contributing drainage area ( $R^{2}=0.228 ; P=0.072$ ), or mean discharge ( $R^{2}=$ 0.225; $P=0.074$ ) for each sample site. There was no significant difference in mean sampling effort among rivers $(P=0.296)$. We also found that BW habitats, on average, contained approximately $63 \%$ of the species detected in SBW habitat.

Number of samples required to estimate mean sample SPUE within 10\%, $25 \%$, and $50 \%$ of the mean population SPUE varied by gear type (Table 5). A
minimum of 5 and a maximum of 126 seine hauls are needed to estimate SPUE within $50 \%$ and $10 \%$ of the mean population respectively. A minimum of 13 and a maximum of 319 large hoop nets are needed and a minimum of 13 and a maximum of 322 small hoop nets are needed to estimate SPUE within $50 \%$ and $10 \%$ of the mean population respectively.

## Sample Distance

Average wetted width of the river at each sample site was 147.8 m (SD = 123.01; Table 6). On average, DNW habitat comprised 45.6\% (SD = 9.7), SW comprised 43.2\% (SD = 8.9), andBW comprised 11.2\% ( $\mathrm{SD}=4.9$ ) of each sample site (Table 6).

For each sample site we calculated the minimum sample distance as the distance needed to encounter at least eight SBW habitats and four DNW habitats, based on the average effort needed to detect a majority of fish species at each sample site. Minimum sampling distance ranged from 400 m to 1600 m and averaged $886.7 \mathrm{~m}(\mathrm{SD}=344.1$; Table 7) over all the sample sites. Based on these results, within a 900 m reach of a large prairie river we would expect to encounter 9 to 15 SW habitats $(\bar{x}=8.0$; SD $=3.02)$, 0 to 6 BW habitats $(\bar{x}=2.5$; SD = 1.81), and 5 to 22 DW habitats ( $\bar{x}=10.0 ; \mathrm{SD}=4.83$; Table 7 ).

Regression analysis between mean wetted width and average percentages of SW, DW, and BW habitats at each sample site indicated that $66.3 \%$ of the variation in wetted width could be explained by the occurrence of BW habitat ( $P<0.001$ ), whereas SW and DW habitat did not explain a significant
portion of the variation ( $P=0.320$ and $P=0.415$ respectively). Based on these results, occurrences of BW habitat in the river channel should increase as wetted width increases.

## Gear Evaluation

We collected 45,586 individual fish representing 13fami lies and 46 species from the 5 sample sites. We had a combined effort of 334 seine hauls and collected 44 speciesepresenting 13 families with a SPUE of $3.6(S D=2.04)$ species per seine haul (Table 8). Fish collected with the seine had a mean length of $42.4 \mathrm{~mm}(\mathrm{SD}=37.13)$ and a mean weight of $3.7 \mathrm{~g}(\mathrm{SD}=50.32)$. With the large hoop nets, we had a combined effort of 51 net nights and collected 11 species representing 7 families with a SPUE of $1.4(S D=1.8)$ species per net night (Table 8). Fish collected with the large hoop net had a mean length of $527.2 \mathrm{~mm}(\mathrm{SD}=199.15)$ and a mean weight of $1904.5 \mathrm{~g}(\mathrm{SD}=1963.08)$. With the small hoop nets, we had a combined effort of 79 net nights and collected 16 species representing 8 families with a SPUE of1.2 (SD = 1.0) species per net night (Table 8). Fish collected with the small hoop net had a mean length of $357.3 \mathrm{~mm}(\mathrm{SD}=200.75)$ and a mean weight of $718.8 \mathrm{~g}(\mathrm{SD}=917.38)$.

Significantly more species were collected per unit effort by seining than were captured by both the large hoop net ( $P<0.001$ ) and small hoop net ( $P<$ 0.001), when averaged over all sites. However, there was no significant difference in the SPUE between the two sizes of hoop nets ( $P=0.565$ ). Large
hoop nets captured significantly larger fish than both the seine ( $P<0.001$ ) and the small hoop nets ( $P<0.001$ ).

Analysis of similarities of fish species captured by each gear type showed distinct differences between the fish assemblage collected by seining and those collected with both sizes of hoop nets $(R=0.783)$; however, we found no difference between the fish assemblages captured with the large and small hoop nets $(R=0.022)$. Seining captured species from all 13 families collected in this study. The majority of species collected by seining were minnows, including red shiner Cyprinella lutrensis, emerald shiner Notropis atherinoides, western mosquitofish Gambusia affinis, and bullhead minnow Pimephales vigilax, which made up over $90 \%$ of the total catch (Table 8). In contrast, the majority of fish species captured by hoop nets were catfish. Channel catfish Ictalurus punctatus and flathead catfish Pylodictis olivaris represented over 50\% of the combined catch in large and small hoop nets whereas smallmouth buffalo Ictiobus bubalus and river carpsucker Carpiodes carpio represented $47 \%$ and $12 \%$ respectively (Table 8). Although there was no difference in the fishassemblages captured by large and small hoop nets $(R=0.022)$, small hoop nets captured more sunfish species $(N=4)$ than large hoop nets $(N=1$; Table 8$)$.

## Sampling Recommendations

The following sampling recommendations and protocol are intended for sample sites that meet the criteria for large prairie rivers in the southern Great Plains region and are based on sampling conducted at 15 sites in five large
prairie rivers in Oklahoma. Sampling in these rivers should proceed through three stages: (1) verification of sample site access and physical characteristics using maps and DOQs prior to leaving for the field, (2) habitat mapping and sample selection in the field, and (3) fish sampling (Figure 4).

Stage 1: site verification--Prior to going into the field, map sources (paper and digital topographic maps, and aerial photos including DOQs) should be consulted to determine site access and to assess river and riparian habitat features. Interpretation of habitat types on DOQs or other aerial photos will provide an initial estimate of the channel dimensions and proportion of shallow water, deep water, and backwater habitats, which can be used to allocate sampling effort (i.e., sample distance).

Stage 2: sample selection and allocation--Once in the field, map the sample reach by walking (or boating) transects perpendicular to the shoreline and spaced 100 m apart. Identify and map SBW and DNW habitat along each transect and number each habitat. After habitat types are mapped, randomly select eight shallow water habitats. Include (when present) at least three BW habitats in with the randomly selected SBW habitats. Deep non-wadeable habitats should be selected based on the presence of in-stream structure and vegetation.

Stage 3: fish sampling--A minimum of two hoop nets should be set in each selected DNW habitat throughout the sample reach, one large ( $\sim 0.91 \mathrm{~m}$ diameter, 50.8 mm mesh) and one small ( $\sim 0.61 \mathrm{~m}$ diameter, 25.4 mm mesh). Placement of hoop nets may be at the discretion of the biologist, but for best results they should be placed near in-stream structure (e.g. woody debris, boulders, and undercut banks) and vegetation. The opening of the net should be facing the center point of the structure and should extend parallel (or underneath if possible) to the structure. If possible, secure the front of the net to the structure and tie the cod end to a $\ddagger$ post (or log if available). The opening of the net should always be facing downstream. Nets should be fished overnight and for a minimum of 12 hours.

After hoop nets are set, seine eight randomly selected SBW habitats throughout the reach using a $6.1 \times 1.2 \mathrm{~m}$ seine with 4.8 mm mesh attached to 1.5 $m$ PVC brails ( $\sim 50.8 \mathrm{~mm}$ diameter). If present in the sample reach, include at least three BW habitats among the eight SBW habitats. Make two seine hauls at each selected habitat with each seine haul covering approximately 10 m . Use the river bank to trap the fish when possible and always seine with the current. Try to seine near unique microhabitat (e.g. cobble, bedrock) and structure (e.g. woody debris, boulder, undercut bank) but be careful not to hang up the seine. If an area is not safe to seine, do not proceed (e.g. water too fast, sharp metal debris, unstable substrate, barbed wire).

## Discussion

Prairie streams in the Great Plains ecoregion have become increasingly degraded over the past century because of extensive agricultural activities, urbanization, and alteration of natural hydrologyas a result of ground water withdrawal and construction of impoundments (Cross and Moss 1987; Cashner and Matthews 1988; Echelle et al. 1995; Matthews 1998; Dodds et al. 2004). Because of marked changes in the physical, chemical, and hydrological characteristics of some of these prairie rivers (Pigg 1988; Pigg et al. 1992; Pigg et al. 1997), several species that were once widespread have declined in their distribution and abundance. Species such as the Arkansas River shiner Notropis girardi and the Topeka shiner Notropis topeka have beerreduced to critically low numbers and are presently listed as federally endangered species (Echelle et al. 1995; Pigg et al. 1999; Dodds et al. 2004). As more species in prairie rivers become imperiled, maintaining existing natural prairie streams and restoring impaired streams will become increasingly difficult (Dodds et al. 2004). It is, therefore, imperative to assess and monitor impacts of human disturbances on the fauna of prairie rivers in this region.

Rapid bioassessment protocols for sampling fishes vary widely among biomonitoring agencies throughout the U.S. depending on the physico-chemical conditions of theriver being sampled and the context in which the protocols are being employed. Standardized sampling protocols for fishes in the Ohio River are based solely on the use of electrofishing (OhioEPA 1989; Emery et al. 2003; Yoder and Kulik 2003), whereas standard operating procedures for sampling
fishes in the Missouri River are based on the combined use of seining, gill netting, trammel netting, mini-fyke netting, benthic trawling, and electrofishing (USGS 1998). These studies illustrate that in most cases sampling methodologies are tailored for a specific area or speciesand are not necessarily applicable for areas outside their region. Perhaps the most widespread and well know rapid bioassessment protocols for sampling fishes were those developed by the U.S. EPA for use in wadeable (Barbour et al. 1999) and non-wadeable (Lazorchak et al. 2000) streams and rivers. Although these are widely applicable protocols, they rely solely on the use of electrofishing, which is not feasible in the highly conductive rivers of the southern Great Plains region.

Fish species in prairie rivers have been shown to be strongly associated with depth, current velocity, substrate type and the occurrence of in-stream structure and vegetation (O'Shea et al. 1990; Taylor et al. 1993; Ostrand and Wilde 2002). Because habitats act as filters determining the kind of fish that are present in a system (Poff 1997; Keddy and Weiher 1999) and are easily identified, it is appropriate to base fish sampling protocols on the occurrence of habitat types in a stream. We found distinct differences in fish species composition based on depth and flow in five large prairie rivers in Oklahoma. However, some of these differences may be an artifact of our sampling method. Nearly all SBW habitats were sampled using a seine and all DNW habitats were sampled using hoop nets. Seines and hoop nets each have their own sampling biases. Seines typically select for smaller bodied, slower moving fishes (Hayes et al. 1996; Bayley and Herendeen 2000), whereas hoop nets select for larger
bodied, more elusive fishes (Holland and Peters 1992; Hubert 1996). Between these two methods, however, we are confident that all but the rarest fish species were detected in our samples based on historical collection records (Jester et al. 1992; Pigg et al. 1992; Miller and Robison 2004). For instance, a comparison of our fish collections with historical fish collections from the North Canadian (Pigg et al. 1992; Pigg et al. 1997) and Cimarron (Pigg 1988) rivers showed that our method of seining and hoop netting detected a majority (>50\%) of all the fish species collected at similar sites, and over $90 \%$ of the common speciesas described by Pigg et al. (1992).

Even though there are distinct physico-chemical differences among the rivers we sampled, there was no relationship between these properties and estimates of required sampling effort. We found that to capture a majority of species during a relatively rapid sampling event (24 hours) at a site on large prairie rivers in Oklahoma, we had to seine 6 to $10(\bar{x}=8)$ SBW habitats and set hoop nets in 1 to $6(\bar{x}=4)$ DNW habitats. We based our minimum sampling effort on the recommendation by Lyons (1992) that the most efficient sampling effort for estimating species richness is the point where the cumulative species richness first levels off. Because conclusions about community dynamics are reliant on a representative sample, a majority of the species should be accounted for in a sample reach. Lyons (1992) and Angermeier and Smogor (1995) suggest that number of fish species typically exhibits a cumulative pattern of diminishing increase with increasing effort and found that proportions of species increased asymptotically and became less variable with greater sampling effort.

Similarly, our estimates of sample size requirements (Table 5) showed that a larger number of samples are needed depending on the goals of the study. For research studies, it may be necessary to sample extensively in both shallow and deep water habitat to get within $10 \%$ of the true species richness at a site; however, fewer samples are needed for preliminary surveys (Robson and Regier 1964; Wilde and Fisher 1996) such as rapid bioassessment protocols (Barbour et al. 1999).

There was a strong positive relationship between river size (i.e., wetted width) and occurrence of BW habitat. Because of the hydrologic variability in prairie rivers, BW habitats form fairly quickly and tend to strand many different species in a relatively small area. Fish move into river margins (e.g., oxbows, isolated backwaters) during times of high flow and are often stranded as the water recedes (Starrett 1951). Althoughour analysis of species -habitat similarities did not differentiate fish assemblages in BW habitats from those inthe other shallow water habitats (SS and SF), collecting fish in these habitats increases sampling efficiency because they are easy to locate in the river channel, are typically easy to sample with a seineand have a high speci es richness.

A common approach for determining the length of river to sample is to base distance on a multiple of the mean wetted width of the river section (Lyons 1992; Angermeier and Smoger 1995; Barbour et al. 1999). However, this approach will result in over-sampling in large prairie rivers because they typically have a wetted width > 100 m . Our method of determining sampling distance was
based on the average occurrence of habitat in the river channel by interpreting habitat types on aerial photographs using GIS. More importantly, since aerial photos (e.g. DOQs) have become more readily accessible, this approach would allow someone to customize sampling distance based on occurrence of habitat at particular sample site of interest without being bound by a channel morphology relationship.

We found that large and small hoop nets differ in size of fish and number of species collected. Overall, small hoop nets collected more species and smaller sized fish. This is a problem only if the management objective requires a representative sample from all size classes. We recommend using a combination of large and small hoop nets to ensure that all size classes have an equal opportunity to be detected. Our findings corroborate those by Holland and Peters (1992) who investigated the difference in detectability and fish size among hoop nets of three mesh sizes in the lower Platte River. They foundhat smaller mesh hoop nets ( 25 mm ) detected $82 \%$ of the fish species compared to only $18 \%$ in the larger mesh hoop nets (32 mm and 38 mm ). They also found that the larger-mesh hoop nets detected bigger fish on average ( 316 mm ) than smaller mesh hoop nets (266 mm). Hoop nets also are effective for capturing species such as channel catfish (Vokoun and Rabeni 2001; Vokoun and Rabeni 2002) and paddlefish (Dieterman et al. 2000) in prairie rivers and are a useful tool for assessing population characteristics when different mesh sizes are used concurrently.

We were able to detect a majority of the fish species present in five large prairie rivers in Oklahoma with our sampling approach during a single sampling event. However, our sampling recommendations may not be suitable for all management objectives. Because of the high variability in catch at each sample site, we determined that an extensive amount of effort is needed to obtain a statistically valid estimate of fish species richness within $25 \%$ or $10 \%$ of the true species richness. Although our sampling protocol enabled us to detect the majority of fish species present at each site, caution must be taken to ensure that rare species are not missed. Rare species are critical to the bioassessment of aquatic systems and not detecting these species can negatively influence the ability of community-based metrics to detect ecological changes (Cao et al. 1998). Additional species will likely be added if sampling is conducted in other seasons. In addition, se of a Bayesian approach to detect unsampled species known to occur at a site (Bayley and Peterson 2001) could be used to improve our sampling protocol. This approach estimates species-specific detection probabilities based on previous knowledge of species occurrence at a sample site and the catchability of each species based on sampling method (e.g., seining). With only slight adjustments, however, we feel that our recommendations can be applied to large prairie rivers throughout the southern Great Plains, as well as similar rivers throughout the northern Great Plains, to aid in the rapid bioassessment, and monitoring of prairie river fish assemblages.

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## References

Angermeier, P. L., and R. A. Smoger. 1995. Estimating number of species and relative abundances in stream-fish communities: effects of sampling effort and discontinuous spatial distributions. Canadian Journal of Fisheries and Aquatic Sciences 52:936-949.

Barbour, M. T., J. Gerritsen, B. D. Snyder, and J. B. Stribling. 1999. Rapid bioassessment protocols for use in wadeable streams and rivers. U.S. Environmental Protection Agency, EPA 841-B-99-002, Washington, DC.

Barfoot, C. A., and R. G. White. 1999. Fish assemblage and habitat relationships in a small northern Great Plains stream. Prairie Naturalist 31:87-107.

Bayley, P. B., and R. A. Herendeen. 2000. The efficiency of a seine net. Transactions of the American Fisheries Society 129:901-923.

Bayley, P. B., and J. T. Peterson. 2001. An approach to estimate probability of presence and richness of fish species. Transactions of the American Fisheries Society 130:620-633.

Cao, Y., D. D. Williams, and N. E. Williams. 1998. How important are rare species in aquatic community ecology and bioassessment. Limnology and Oceanography 43:1403-1409.

Cashner, R. C., and W. J. Matthews. 1988. Changes in the known Oklahoma fish fauna from 1973 to 1988. Proceedings of the Oklahoma Academy of Science 68:1-7.

Chapman, D. 1996. Water quality assessments: a guide to the use of biota, sediments and water in environmental monitoring, Second edition. Chapman \& Hall, London.

Clarke, K. R., and R. N. Gorley. 2001. PRIMER v5: user manual/tutorial. PRIMER-E Ltd, Plymouth, United Kingdom.

Clarke, K. R., and R. M. Warwick. 2001. Change in marine communities: an approach to statistical analysis and interpretation, 2nd edition. PRIMER-E Ltd, Plymouth, United Kingdom.

Cross, F. B., and R. E. Moss. 1987. Historic changes in fish communities and aquatic habitats in plains streams of Kansas. Pages 155-167 in D. C.

Heins, editor. Community and evolutionary ecology of North American stream fishes. University of Oklahoma Press, Norman, Oklahoma.

Dieterman, D. J., M. S. Baird, and D. L. Galat. 2000. Mortality of paddlefish in hoop nets in the lower Missouri River, Missouri. North American Journal of Fisheries Management 20:226-230.

Dodds, W. K., K. B. Gido, M. R. Whiles, K. M. Fritz, and W. J. Matthews. 2004. Life of the edge: the ecology of Great Plains prairie streams. Bioscience 54:205-216.

Echelle, A. A., G. R. Luttrell, R. D. Larson, A. V. Zale, W. L. Fisher, and D. M. Leslie, Jr. 1995. Decline of native prairie fishes. Pages 303-305 in M. J. Mac, editor. Our living resources: a report to the nation on the distribution, abundance, and health of U.S. plants, animals, and ecosystems. U.S. Department of the Interior, National Biological Service, Washington, D.C.

Emery, E. B., T. P. Simon, F. H. McCormick, P. L. Angermeier, J. E. Deshon, C. O. Yoder, R. E. Sanders, W. D. Pearson, G. D. Hickman, R. J. Reash, and J. A. Thomas. 2003. Development of a multimetric index for assessing the biological condition of the Ohio River. Transactions of the American Fisheries Society 132:791-808.

Fisher, W. L. 1987. Benthic fish sampler for use in riffle habitats. Transactions of the American Fisheries Society 116:768-772.

Gorman, O. T., and J. R. Karr. 1978. Habitat structure and stream fish communities. Ecology 59:507-515.

Hayes, D. B., C. P. Ferreri, and W. W. Taylor. 1996. Active capture techniques. Pages 193-218 in D. W. Willis, editor. Fisheries techniques. American Fisheries Society, Bethesda, Maryland.

Holland, R. S., and E. J. Peters. 1992. Differential catch by hoop nets of three mesh sizes in the lower Platte River. North American Journal of Fisheries Management 12:237-243.

Hubert, W. A. 1996. Passive capture techniques. Pages 157-182 in D. W. Willis, editor. Fisheries techniques, second edition. American Fisheries Society, Bethesda, Maryland.

Jester, D. B., A. A. Echelle, W. J. Matthews, J. Pigg, C. M. Scott, and K. D. Collins. 1992. The fishes of Oklahoma, their gross habitats, and their tolerance of degradation in water quality and habitat. Proceedings of the Oklahoma Academy of Science 72:7-19.

Keddy, P., and E. Weiher. 1999. Introduction: the scope and goals of research on assembly rules. Pages 1-20 in P. Keddy, editor. Ecological assembly rules: perspectives, advances, retreats. Cambridge University Press, Cambridge.

Lazorchak, J. M., B. H. Hill, D. K. Averill, D. V. Peck, and D. J. Klemm. 2000. Environmental monitoring and assessment program-surface waters: field operations and methods for measuring the ecological condition of nonwadeable rivers and streams. U.S. Environmental Protection Agency, EPA/620/R-00/007, Washington D.C.

Lyons, J. 1992. The length of stream to sample with a towed electrofishing unit when fish species richness is estimated. North American Journal of Fisheries Management 12:198-203.

Lyons, J., R. R. Piette, and K. W. Niermeyer. 2001. Development, validation, and application of a fish-based index of biotic integrity for Wisconsin's large warmwater rivers. Transactions of the American Fisheries Society 130:1077-1094.

Matthews, W. J. 1988. North American prairie streams as systems for ecological study. Journal of the North American Benthological Society 7:387-409.

Matthews, W. J. 1998. Patterns in freshwater fish ecology. Kluwer Academic Publishers, Norwell, Massachusetts.

Milewski, C. L., C. R. Berry, Jr., and D. J. Dieterman. 2001. Use of the index of biological integrity in eastern South Dakota rivers. Prairie Naturalist 33:135-151.

Miller, R. J., and H. W. Robison. 2004. The fishes of Oklahoma. University of Oklahoma Press, Norman.

OhioEPA. 1989. Biological criteria for the protection of aquatic life. Vol. III. Standardized biological field sampling and laboratory methods for assessing fish and macroinvertebrate communities. Division of Water Quality Monitoring and Assessment, Surface Water Section, Columbus, Ohio. Available: www.epa.state.oh.us/dsw/bioassess/BioCriteriaProtAqLife.html. (May 2005).

Omernik, J. M. 1987. Ecoregions of the conterminous United States. American Association of Geography 77:118-125.

O'Shea, D. T., W. A. Hubert, and S. H. Anderson. 1990. Assemblages of small fish in three habitat types along the Platte River, Nebraska. Prairie Naturalist 22:145-154.

Ostrand, K. G., and G. R. Wilde. 2002. Seasonal and spatial variation in a prairie stream-fish assemblage. Ecology of Freshwater Fish 11:137-149.

OWRB (Oklahoma Water Resources Board). 1999. Standard operating procedures for stream assessments and biological collections related to biological criteria in Oklahoma. Oklahoma Water Resources Board, Technical Report 99-3, Oklahoma City, Oklahoma.

Peters, E. J., R. S. Holland, M. A. Callam, and D. L. Bunnell. 1989. Platte River suitability criteria: habitat utilization, preference and suitability index criteria for fish and aquatic invertebrates in the lower Platte River. Nebraska Game and Parks Commission, Project F-78-R, Lincoln, Nebraska.

Pigg, J. 1988. Aquatic habitats and fish distribution in a large Oklahoma river, the Cimarron, from 1976 to 1986. Proceedings of the Oklahoma Academy of Science 68:9-31.

Pigg, J., M. S. Coleman, and J. Duncan. 1992. An ecological investigation of the ichthyofauna of the North Canadian River in Oklahoma: 1976-1989. Proceedings of the Oklahoma Academy of Science 72:21-32.

Pigg, J., M. S. Coleman, and R. Gibbs. 1997. Temporal and spatial distribution of cyprinid fishes between 1921 and 1995 in the North Canadian River drainage, Oklahoma. Proceedings of the Oklahoma Academy of Science 77:43-92.

Pigg, J., R. Gibbs, and K. K. Cunningham. 1999. Decreasing abundance of the Arkansas River shiner in the South Canadian River, Oklahoma. Proceedings of the Oklahoma Academy of Science 79:7-12.

Poff, N. L. 1997. Landscape filters and species traits: towards mechanistic understanding and prediction in stream ecology. Journal of the North American Benthological Society 16:391-409.

Poff, N. L., and J. V. Ward. 1989. Implications of streamflow variability and predictability for lotic community structure: a regional analysis of streamflow patterns. Canadian Journal of Fisheries and Aquatic Sciences 46:1805-1817.

Robson, D. S., and H. A. Regier. 1964. Sample size in Petersen mark-recapture experiments. Transactions of the American Fisheries Society 93:215-226.

Snedecor, G. W., and W. G. Cochran. 1967. Statistical methods, 6th edition. Iowa State University Press, Ames, IA.

Starrett, W. C. 1951. Some factors affecting the abundance of minnows in the Des Moines River, Iowa. Ecology 32:13-27.

Statzner, B., J. A. Gore, and V. H. Resh. 1988. Hydraulic stream ecology: observed patterns and potential applications. Journal of the North American Benthological Society 7:307-360.

Taylor, C. M., M. R. Winston, and W. J. Matthews. 1993. Fish speciesenvironment and abundance relationships in a Great Plains river system. Ecography 16:16-23.

Taylor, C. M., M. R. Winston, and W. J. Matthews. 1996. Temporal variation in tributary and mainstem fish assemblages in a Great Plains stream system. Copeia 1996:280-289.

USGS (U.S. Geological Survey). 1998. Standard operating procedures to evaluate population structure and habitat use of benthic fishes along the Missouri and lower Yellowstone Rivers. U.S. Geological Survey Biological Resources Division, Available: www.cerc.usgs.gov/pubs/benfish/SOP_index.htm. (May 2005).

Vokoun, J. C., and C. F. Rabeni. 2001. A standardized sampling protocol for channel catfish in prairie streams. North American Journal of Fisheries Management 21:188-197.

Vokoun, J. C., and C. F. Rabeni. 2002. Distribution of channel catfish life stages in a prairie river basin. Prairie Naturalist 34:47-59.

Whited, D., J. A. Stanford, and J. S. Kimball. 2002. Application of airborne multispectral digital imagery to quantify riverine habitats at different base flows. River Research and Applications 18:583-594.

Wilde, G. R., and W. L. Fisher. 1996. Reservoir fisheries sampling and experimental design. American Fisheries Society Symposium 16:397-409.

Yoder, C. O., and B. H. Kulik. 2003. The development and application of multimetric indices for the assessment of impacts to fish assemblages in
large rivers: a review of current science and applications. Canadian Water Resources Journal 28:301-311.

Table 1--Physical and chemical characteristics of 15 sites on five large non-wadeable rivers in central Oklahoma; averaged from long term USGS gaging station data. Standard Deviation is in parentheses. Site names were based on city names near the location where fish were collected.


Table 2--Description of the habitat types used when mapping the sample reach.

| Habitat Type | Code | Depth (m) | Velocity <br> $(\mathrm{m} / \mathrm{s})$ | Channel Location |
| :--- | :---: | :---: | :---: | :---: |
| Shallow Slow | SS | $<0.75$ | $<0.20$ | Typically found along the <br> bank or around mid- <br> channel islands. |
| Shallow Fast | SF | $<0.75$ | $>0.20$ | Typically found mid- <br> channel away from <br> obstructions. |
| Deep Slow | DS | $0.76-1.50$ | $<0.20$ | Lateral pools typically <br> surrounded by woody <br> debris (or other structure). <br> Lateral or mid-channel <br> pools typically free of <br> obstructions. |
| Neep Fast | DF | $0.76-1.50$ | $>0.20$ | Any |
| Nonpically found in the <br> wadeable | NW | $>1.50$ | Velocity | Thalweg and surrounded by <br> very little structure. |
| Backwater | BW | Any Depth | $<0.01$ | Still water either mostly or <br> totally separated from the <br> main channel. |

Table 3--Analysis of similarity of fish collections from six habitat types at 15 sites on five large prairie rivers in Oklahoma. Habitats sampled were: shallow fast (SF), shallow slow (SS), deep fast (DF), deep slow (DS), non-wadeable (NW), and backwater (BW). An R-statistic less than 0.5 indicates the two habitat types have a similar species composition and a significance level less than 5\% ( $P<$ 0.05 ) indicates a significant comparison. Based on fish species composition, habitat groups are either well separated ( $R>0.75$ ), overlapping but clearly different ( $R>0.50$ ), or barely separable ( $R<0.25$ ).

| Groups | R Statistic | Significance Level |
| :--- | :---: | :---: |
| SS, SF | 0.026 | 0.039 |
| DS, DF | 0.080 | 0.060 |
| DS, NW | 0.087 | 0.095 |
| DF, NW | 0.118 | 0.054 |
| SS, BW | 0.174 | 0.001 |
| SF, BW | 0.283 | 0.001 |
| BW, DS | 0.706 | 0.001 |
| BW, DF | 0.709 | 0.001 |
| SF, DS | 0.736 | 0.001 |
| SF, DF | 0.745 | 0.001 |
| BW, NW | 0.757 | 0.001 |
| SF, NW | 0.782 | 0.001 |
| SS, DS | 0.798 | 0.001 |
| SS, DF | 0.801 | 0.001 |
| SS, NW | 0.831 | 0.001 |

Table 4--Results from species accumulation curves for shallow/backwater (SBW) and deep/non-wadeable (DNW) habitats from 15 sample sites located on five large prairie rivers in Oklahoma.

| River/Site | Habitats Needed To Attain Maximum Species Richness |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SBW |  |  |  | DNW |  |  |  | Grand Total |
|  | SS | SF | BW | Total | DS | DF | NW | Total |  |
| ARKANSAS R. |  |  |  |  |  |  |  |  |  |
| NEW | 1 | 4 | 2 | 7 | 2 | 0 | 1 | 3 | 10 |
| RAL | 3 | 2 | 1 | 6 | 0 | 1 | 0 | 1 | 7 |
| COW | 3 | 3 | 2 | 8 | 2 | 0 | 0 | 2 | 10 |
| Cimarron R. |  |  |  |  |  |  |  |  |  |
| DOV | 3 | 2 | 3 | 8 | 1 | 3 | 0 | 4 | 12 |
| GUT | 3 | 3 | 2 | 8 | 1 | 0 | 1 | 2 | 10 |
| COY | 2 | 3 | 3 | 8 | 2 | 2 | 2 | 6 | 14 |
| N. Canadian R. |  |  |  |  |  |  |  |  |  |
| ELR | 3 | 1 | 2 | 6 | 2 | 1 | 0 | 3 | 9 |
| HAR | 3 | 3 | 2 | 8 | 2 | 3 | 0 | 5 | 13 |
| WET | 3 | 3 | 3 | 9 | 3 | 1 | 1 | 5 | 14 |
| Washita R. |  |  |  |  |  |  |  |  |  |
| ALE | 2 | 2 | 2 | 6 | 2 | 1 | 1 | 4 | 10 |
| PAU | 3 | 2 | 3 | 8 | 2 | 2 | 0 | 4 | 12 |
| DUR | 4 | 1 | 4 | 9 | 1 | 1 | 3 | 5 | 14 |
| Red R. |  |  |  |  |  |  |  |  |  |
| WAU | 4 | 3 | 3 | 10 | 0 | 2 | 0 | 2 | 12 |
| THA | 4 | 4 | 1 | 9 | 0 | 4 | 0 | 4 | 13 |
| ART | 3 | 3 | 3 | 9 | 1 | 0 | 1 | 2 | 11 |
| Mean ( $\pm$ SD) | $\begin{array}{r} 3 \\ (0.8) \\ \hline \end{array}$ | $\begin{array}{r} 3 \\ (0.9) \\ \hline \end{array}$ | $\begin{array}{r} 2 \\ (0.8) \\ \hline \end{array}$ | $\begin{array}{r} 8 \\ (1.2) \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ (0.9) \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ (1.2) \\ \hline \end{array}$ | $\begin{array}{r} 1 \\ (0.9) \\ \hline \end{array}$ | $\begin{array}{r} 4 \\ (1.5) \\ \hline \end{array}$ | $\begin{array}{r} 11 \\ (2.1) \\ \hline \end{array}$ |

Table 5--Species-per-unit-effort (SPUE) for fish species collected in seines, and/or large and small hoop nets from 15 sample sites on five large prairie rivers in Oklahoma. SPUE estimates were used to predict the number of samples required to obtain species richness estimates within $10 \%, 25 \%$, and $50 \%$ of the true population means ( $\mathrm{P}<0.05$ ). $\mathrm{N}=$ number of samples collected, and SD = standard deviation.

| Gear Type |  | SPUE |  |  | Samples required to detect difference of: |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  |  | Mean | SD |  | $10 \%$ | $25 \%$ | $50 \%$ |
| Seine | 334 | 3.61 | 2.03 |  | 126 | 20 | 5 |
| Small Hoop Net | 79 | 1.16 | 1.04 |  | 322 | 51 | 13 |
| Large Hoop Net | 51 | 1.4 | 1.25 |  | 319 | 51 | 13 |

Table 6--Average wetted width and percentages of shallow water (SW), backwater (BW), and deep water (DW) habitat derived from historical digital orthophoto quad maps for 15 sample sites on five large prairie rivers in Oklahoma.

| River/Site | Wetted Width (m) | Percentage of Habitat |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | SW | BW | DNW |
| ARKANSAS R. |  |  |  |  |
| NEW | 134.7 (53.4) | 48.3 | 10.0 | 41.7 |
| RAL | 340.9 (91.2) | 49.4 | 11.9 | 38.8 |
| COW | 466.4 (109.3) | 40.1 | 17.1 | 42.8 |
| Cimarron R. |  |  |  |  |
| DOV | 131.0 (66.5) | 52.2 | 14.2 | 33.6 |
| GUT | 131.3 (66.5) | 40.1 | 17.1 | 42.8 |
| COY | 127.3 (58.9) | 47.9 | 10.1 | 42.0 |
| N. Canadian R. |  |  |  |  |
| ELR | 21.1 (4.6) | 58.4 | 6.9 | 34.7 |
| HAR | 29.2 (6.7) | 41.8 | 4.6 | 53.6 |
| WET | 66.7 (23.9) | 28.0 | 8.7 | 63.3 |
| Washita R. |  |  |  |  |
| ALE | 44.0 (14.7) | 47.0 | 7.5 | 45.5 |
| PAU | 68.2 (18.5) | 45.3 | 6.8 | 47.9 |
| DUR | 74.1 (21.9) | 46.4 | 7.5 | 46.1 |
| Red R. |  |  |  |  |
| WAU | 245.1 (129.0) | 45.4 | 22.7 | 31.9 |
| THA | 132.5 (27.0) | 32.2 | 9.7 | 58.1 |
| ART | 203.8 (46.3) | 25.5 | 13.3 | 61.2 |
| Mean ( $\pm$ SD) | 147.8 (123.0) | 43.2 (8.9) | 11.2 (4.9) | 45.6 (9.7) |

Table 7--Minimum sampling distance needed at each site to encounter a minimum of eight shallow water (SW) habitats, or combination of eight SW and backwater (BW) habitats, and four deep water (DW) habitats. Based on data collected from 15 sample sites on five large prairie rivers in Oklahoma.

| River/Site | Minimum Distance (m) | Number of Habitats Encountered |  |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | SW | BW | DW |
| ARKANSAS R. |  |  |  |  |
| NEW | 700 | 6 | 3 | 9 |
| RAL | 1000 | 15 | 5 | 6 |
| cow | 500 | 5 | 4 | 5 |
| Cimarron R. |  |  |  |  |
| DOV | 700 | 7 | 3 | 8 |
| GUT | 700 | 10 | 0 | 10 |
| COY | 700 | 6 | 3 | 11 |
| N. Canadian R. |  |  |  |  |
| ELR | 1600 | 14 | 1 | 5 |
| HAR | 1500 | 9 | 0 | 17 |
| WET | 1100 | 7 | 2 | 9 |
| WAShita R. |  |  |  |  |
| ALE | 800 | 8 | 1 | 11 |
| PAU | 600 | 6 | 3 | 7 |
| DUR | 800 | 6 | 3 | 10 |
| Red R. |  |  |  |  |
| WAU | 400 | 5 | 6 | 5 |
| THA | 1100 | 9 | 0 | 22 |
| ART | 1100 | 7 | 3 | 15 |
| Mean ( $\pm$ SD) | 886.7 (344.1) | 8.0 (3.0) | 2.5 (1.8) | 10.0 (4.8) |

Table 8--Comparison oftotal number of fish caught in two gear types and two hoop net sizes. Based on data collected from 15 sample sites on five large prairie rivers in Oklahoma.

| Family/Species | Seine | Hoop Net |  |
| :---: | :---: | :---: | :---: |
|  |  | Large | Small |
| ATHERINOPSIDAE |  |  |  |
| Menidia beryllina | 1134 | 0 | 0 |
| Catostomidae |  |  |  |
| Carpiodes carpio | 335 | 19 | 16 |
| Ictiobus bubalus | 106 | 29 | 7 |
| Ictiobus cyprinellus | 8 | 1 | 0 |
| Moxostoma erythrurum | 1 | 0 | 0 |
| Minytrema melanops | 1 | 0 | 0 |
| Centrarchidae |  |  |  |
| Lepomis cyanellus | 28 | 0 | 0 |
| Lepomis gulosus | 1 | 0 | 0 |
| Lepomis humilis | 6 | 0 | 1 |
| Lepomis macrochirus | 366 | 0 | 9 |
| Lepomis megalotis | 0 | 0 | 22 |
| Micropterus punctulatus | 20 | 1 | 0 |
| Pomoxis annularis | 23 | 0 | 2 |
| Clupeidae |  |  |  |
| Dorosoma cepedianum | 85 | 0 | 10 |
| Dorosoma petenense | 37 | 0 | 0 |
| Hiodon alosoides | 1 | 0 | 0 |
| Cyprinidae |  |  |  |
| Campostoma anomalum | 2 | 0 | 0 |
| Cyprinella lutrensis | 28442 | 0 | 0 |
| Cyprinus carpio | 2 | 28 | 13 |
| Hybognathus placitus | 6002 | 0 | 0 |
| Macrhybopsis australis | 40 | 0 | 0 |
| Macrhybopsis hyostoma | 211 | 0 | 0 |
| Notropis atherinoides | 4126 | 0 | 0 |
| Notropis bairdi | 589 | 0 | 0 |
| Notropis blennius | 7 | 0 | 0 |
| Notropis buchanani | 54 | 0 | 0 |
| Notropis stramineus | 96 | 0 | 0 |
| Phenacobius mirabilis | 22 | 0 | 0 |
| Pimephales vigilax | 1686 | 0 | 0 |
| Cyprinodontidae |  |  |  |
| Cyprinodon rubrofluviatilis | 87 | 0 | 0 |
| FUNDULIDAE |  |  |  |
| Fundulus zebrinus | 109 | 0 | 0 |
| ICtaluridae |  |  |  |
| Ictalurus furcatus | 2 | 2 | 2 |
| Ictalurus punctatus | 23 | 13 | 63 |
| Pylodictis olivaris | 0 | 21 | 34 |
| LEPISOSTEIDAE |  |  |  |
| Lepisosteus oculatus | 6 | 0 | 1 |
| Lepisosteus osseus | 11 | 15 | 2 |


| Lepisosteus platostomus | 15 | 0 | 6 |
| :---: | :---: | :---: | :---: |
| Moronidae |  |  |  |
| Morone chrysops | 4 | 1 | 6 |
| Morone saxatilis | 8 | 0 | 0 |
| Percidae |  |  |  |
| Ammocrypta clara | 1 | 0 | 0 |
| Percina phoxocephala | 2 | 0 | 0 |
| Percina sciera | 5 | 0 | 0 |
| Percina shumardi | 1 | 0 | 0 |
| Sander vitreus | 3 | 0 | 0 |
| Poecilildae |  |  |  |
| Gambusia affinis | 3575 | 0 | 0 |
| Sciamidae |  |  |  |
| Aplodinotus grunniens | 8 | 6 | 12 |
| Mean number of species ( $\pm$ SD) | 3.6 (2.04) | 1.4 (1.25) | 1.2 (1.04) |
| Mean length (mm) ( $\pm$ SD) | 42.4 (37.13) | 527.2 (199.15) | 357.3 (200.75) |
| Mean weight (g) ( $\pm$ SD) | 3.7 (50.32) | 1904.5 (1963.08) | 718.9 (917.38) |

Figure 1: Relationship between large prairie rivers in the interior plains of the U.S. and Omernik's Level II ecoregions (Omernik 1987).

Figure 2: Map of samæe sites (see Table 1 for site abbreviations).

Figure 3: A 10 km section of the Cimarron River near Coyle, Oklahoma. Habitat was delineated from digital orthophoto quad maps using ArcView GIS software. The 10 km section was divided into transects placed every 100 m and habitat was mapped and counted along each transect. This process was repeated for each sample site (see Figure 3) so average wetted width and percentages of shallow water (SW), backwater (BW), and deep water (DW) habitat could be calculated.

Figure 4: Flow chart depicting the multistage method for sampling fishes in large prairie river.





APPENDIXES

Appendix A--Total abundance of fish species collected in shallow slow (SS), shallow fast (SF), deep slow (DS), deep fast (DF), non-wadeable (NW), and backwater (BW) habitat at 15 sites on five large prairie rivers in Oklahoma.

| River/Site | Species | SS | SF | DS | DF | NW | BW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| ARKANSAS RIVER |  |  |  |  |  |  |  |  |
| NEW | Carpiodes carpio | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Cyprinella lutrensis | 215 | 254 | 0 | 0 | 0 | 969 | 1438 |
|  | Cyprinus carpio | 0 | 0 | 4 | 0 | 1 | 0 | 5 |
|  | Fundulus zebrinus | 5 | 0 | 0 | 0 | 0 | 2 | 7 |
|  | Gambusia affinis | 36 | 3 | 0 | 0 | 0 | 86 | 125 |
|  | Ictalurus punctatus | 0 | 6 | 3 | 1 | 0 | 0 | 10 |
|  | Ictiobus bubalus | 8 | 0 | 4 | 0 | 0 | 24 | 36 |
|  | Lepisosteus oculatus | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Lepisosteus osseus | 0 | 0 | 12 | 0 | 0 | 0 | 12 |
|  | Lepisosteus platostomus | 0 | 0 | 1 | 0 | 1 | 0 | 2 |
|  | Lepomis macrochirus | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Menidia beryllina | 469 | 112 | 0 | 0 | 0 | 479 | 1060 |
|  | Micropterus puntulatus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Minytrema melanops | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Morone chrysops | 0 | 0 |  | 0 | 0 | 0 | 1 |
|  | Morone saxatilis | 0 | 4 | 0 | 0 | 0 | 0 | 4 |
|  | Notropis atherinoides | 34 | 144 | 0 | 0 | 0 | 740 | 918 |
|  | Notropis bairdi | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Notropis stramineus | 0 | 0 | 0 | 0 | 0 | 7 | 7 |
|  | Percina phoxocephala | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Pimephales vigilax | 13 | 21 | 0 | 0 | 0 | 33 | 67 |
|  | Polydictis olivaris | 0 | 0 | 0 | 2 | 0 | 0 | 2 |
| RAL | Cyprinella lutrensis | 8 | 1 | 0 | 0 | 0 | 149 | 158 |
|  | Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 6 | 6 |
|  | Gambusia affinis | 2 | 0 | 0 | 0 | 0 | 1437 | 1439 |
|  | Ictalurus punctatus | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Ictiobus bubalus | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 16 | 16 |
|  | Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 349 | 349 |
|  | Macrhybopsis hyostoma | 0 | 7 | 0 | 0 | 0 | 0 | 7 |
|  | Menidia beryllina | 4 | 0 | 0 | 0 | 0 | 8 | 12 |
|  | Micropterus puntulatus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Notropis atherinoides | 506 | 284 | 0 | 0 | 0 | 59 | 849 |
|  | Phenacobius mirabilis | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Pimephales vigilax | 0 | 0 | 0 | 0 | 0 | 13 | 13 |
| cow | Cyprinus carpio | 0 | 0 | 4 | 0 | 0 | 0 | 4 |
|  | Ictalurus punctatus | 0 | 0 | 4 | 0 | 0 | 0 | 4 |
|  | Lepomis macrochirus | 0 | 0 | 7 | 0 | 0 | 0 | 7 |
|  | Lepomis megalotis | 0 | 0 | 2 | 0 | 0 | 0 | 2 |
|  | Menidia beryllina | 24 | 1 | 0 | 0 | 0 | 15 | 40 |
|  | Notropis atherinoides | 140 | 95 | 0 | 0 | 0 | 106 | 341 |
|  | Pimephales vigilax | 8 | 1 | 0 | 0 | 0 | 47 | 56 |

Appendix--A (continued)

| River/Site | Species | SS | SF | DS | DF | NW | BW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Carpiodes carpio | 16 | 0 | 0 | 0 | 0 | 0 | 16 |
|  | Cyprinella lutrensis | 883 | 171 | 0 | 0 | 0 | 551 | 1605 |
|  | Gambusia affinis | 0 | 0 | 0 | 0 | 0 | 36 | 36 |
|  | Ictiobus bubalus | 8 | 0 | 2 | 0 | 0 | 0 | 10 |
|  | Macrhybopsis hyostoma | 83 | 97 | 0 | 0 | 0 | 0 | 180 |
|  | Micropterus puntulatus | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Morone saxatilis | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Percina shumardi | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Phenacobius mirabilis | 8 | 3 | 0 | 0 | 0 | 0 | 11 |
| Cimarron River |  |  |  |  |  |  |  |  |
| DOV | Aplodinotus grunniens | 0 | 0 | 1 | 1 | 0 | 0 | 2 |
|  | Carpiodes carpio | 3 | 0 | 0 | 0 | 0 | 1 | 4 |
|  | Cyprinella lutrensis | 6 | 16 | 0 | 0 | 0 | 129 | 151 |
|  | Cyprinodon rubrofluviatilis | 0 | 0 | 0 | 0 | 0 | 8 | 8 |
|  | Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 8 | 8 |
|  | Fundulus zebrinus | 5 | 0 | 0 | 0 | 0 | 1 | 6 |
|  | Gambusia affinis | 4 | 0 | 0 | 0 | 0 | 162 | 166 |
|  | Hybognathus placitus | 12 | 15 | 0 | 0 | 0 | 0 | 27 |
|  | Ictalurus punctatus | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Micropterus puntulatus | 0 | 0 | 0 | 0 | 0 | 12 | 12 |
|  | Notropis atherinoides | 327 | 247 | 0 | 0 | 0 | 270 | 844 |
|  | Phenacobius mirabilis | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Polydictis olivaris | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| GUT | Carpiodes carpio | 11 | 6 | 1 | 0 | 8 | 110 | 136 |
|  | Cyprinella lutrensis | 45 | 10 | 0 | 0 | 0 | 16 | 71 |
|  | Cyprinodon rubrofluviatilis | 0 | 0 | 0 | 0 | 0 | 8 | 8 |
|  | Cyprinus carpio | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Dorosoma cepedianum | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Fundulus zebrinus | 0 | 0 | 0 | 0 | 0 | 3 | 3 |
|  | Gambusia affinis | 2 | 0 | 0 | 0 | 0 | 40 | 42 |
|  | Hybognathus placitus | 322 | 258 | 0 | 0 | 0 | 63 | 643 |
|  | Lepisosteus oculatus | 0 | 0 | 0 | 0 | 0 | 5 | 5 |
|  | Lepomis humilis | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Morone chrysops | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Notropis atherinoides | 77 | 59 | 0 | 0 | 0 | 28 | 164 |
|  | Pomoxis annularis | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
| COY |  | 0 | 0 | 3 | 2 | 0 | 0 | 5 |
|  | Cyprinus carpio | 0 | 0 | 0 | 0 | 3 | 0 | 3 |
|  | Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 26 | 26 |
|  | Fundulus zebrinus | 0 | 1 | 0 | 0 | 0 | 80 | 81 |
|  | Hybognathus placitus | 140 | 246 | 0 | 0 | 0 | 1078 | 1464 |
|  | Ictalurus punctatus | 0 | 0 | 7 | 1 | 0 | 0 | 8 |
|  | Lepisosteus osseus | 0 | 0 | 0 | 1 | 0 | 1 | 2 |
|  | Lepomis macrochirus | 0 | 0 | 1 | 0 | 0 | 4 | 5 |
|  | Notropis atherinoides | 284 | 173 | 0 | 0 | 0 | 27 | 484 |
|  | Pimephales vigilax | 0 | 0 | 0 | 0 | 0 | 3 | 3 |
|  | Polydictis olivaris | 0 | 0 | 1 | 1 | 2 | 0 | 4 |

Appendix--A (continued)

| River/Site | Species | SS | SF | DS | DF | NW | BW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Carpiodes carpio | 2 | 1 | 2 | 1 | 0 | 37 | 43 |
|  | Cyprinella lutrensis | 2 | 36 | 0 | 0 | 0 | 5 | 43 |
|  | Cyprinodon rubrofluviatilis | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Gambusia affinis | 0 | 0 | 0 | 0 | 0 | 98 | 98 |
|  | Ictiobus bubalus | 0 | 0 | 0 | 0 | 3 | 0 | 3 |
|  | Macrhybopsis hyostoma | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Morone chrysops | 0 | 0 | 0 | 4 | 0 | 0 | 4 |
|  | Notropis bairdi | 60 | 45 | 0 | 0 | 0 | 48 | 153 |
| N. Canadian River |  |  |  |  |  |  |  |  |
| ELR | Cyprinella lutrensis | 1691 | 971 | 0 | 0 | 0 | 11 | 2673 |
|  | Cyprinus carpio | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Fundulus zebrinus | 2 | 5 | 0 | 0 | 0 | 0 | 7 |
|  | Gambusia affinis | 0 | 0 | 0 | 0 | 0 | 4 | 4 |
|  | Ictalurus punctatus | 0 | 0 | 3 | 16 | 0 | 0 | 19 |
|  | Lepomis cyanellus | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
|  | Lepomis humilis | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Lepomis megalotis | 0 | 0 | 11 | 8 | 0 | 0 | 19 |
|  | Notropis bairdi | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Notropis blennius | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Notropis stramineus | 36 | 35 | 0 | 0 | 0 | 0 | 71 |
|  | Phenacobius mirabilis | 2 | 2 | 0 | 0 | 0 | 1 | 5 |
|  | Pimephales vigilax | 9 | 1 | 0 | 0 | 0 | 10 | 20 |
|  | Polydictis olivaris | 0 | 0 | 1 | 7 | 0 | 0 | 8 |
|  | Sander vitreus | 0 | 0 | 0 | 0 | 0 | 3 | 3 |
| HAR |  | 0 | 0 | 2 | 6 | 0 | 0 | 8 |
|  | Cyprinella lutrensis | 439 | 182 | 0 | 0 | 0 | 101 | 722 |
|  | Cyprinus carpio | 0 | 0 | 1 | 6 | 0 | 1 | 8 |
|  | Dorosoma pretense | 0 | 1 | 0 | 0 | 0 | 23 | 24 |
|  | Gambusia affinis | 1 | 0 | 0 | 0 | 0 | 94 | 95 |
|  | Hybognathus placitus | 0 | 0 | 0 | 0 | 0 | 3 | 3 |
|  | Ictalurus furcatus | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Ictalurus punctatus | 0 | 0 | 3 | 1 | 0 | 0 | 4 |
|  | Ictiobus bubalus | 0 | 1 | 3 | 9 | 0 | 0 | 13 |
|  | Ictiobus cyprinellus | 0 | 0 | 0 | 0 | 0 | 5 | 5 |
|  | Lepisosteus osseus | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Lepisosteus platostomus | 0 | 0 | 1 | 0 | 0 | 2 | 3 |
|  | Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Lepomis megalotis | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Morone chrysops | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Notropis buchanani | 0 | 0 | 0 | 0 | 0 | 14 | 14 |
|  | Notropis stramineus | 4 | 6 | 0 | 0 | 0 | 0 | 10 |
|  | Pimephales vigilax | 2 | 15 | 0 | 0 | 0 | 36 | 53 |
|  | Polydictis olivaris | 0 | 0 | 2 | 9 | 0 | 0 | 11 |
|  | Pomoxis annularis | 0 | 0 | 0 | 0 | 0 | 4 | 4 |
| WET | Cyprinus carpio | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Ictalurus punctatus | 0 | 1 | 4 | 0 | 1 | 0 | 6 |
|  | Notropis atherinoides | 1 | 3 | 0 | 0 | 0 | 1 | 5 |
|  | Pimephales vigilax | 456 | 15 | 0 | 0 | 0 | 211 | 682 |

Appendix--A (continued)

| River/Site | Species | SS | SF | DS | DF | NW | BW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Aplodinotus grunniens | 4 | 0 | 0 | 0 | 0 | 1 | 5 |
|  | Carpiodes carpio | 5 | 0 | 0 | 0 | 0 | 3 | 8 |
|  | Cyprinella lutrensis | 3785 | 432 | 0 | 0 | 0 | 8618 | 12835 |
|  | Gambusia affinis | 9 | 0 | 0 | 0 | 0 | 53 | 62 |
|  | Ictiobus bubalus | 0 | 0 | 2 | 0 | 1 | 0 | 3 |
|  | Lepisosteus platostomus | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Lepomis humilis | 2 | 0 | 0 | 0 | 0 | 1 | 3 |
|  | Polydictis olivaris | 0 | 0 | 0 | 3 | 0 | 0 | 3 |
|  | Pomoxis annularis | 1 | 0 | 1 | 0 | 0 | 5 | 7 |
| Washita River |  |  |  |  |  |  |  |  |
| ALE | Carpiodes carpio | 6 | 9 | 0 | 0 | 0 | 115 | 130 |
|  | Cyprinella lutrensis | 1895 | 1217 | 0 | 0 | 0 | 720 | 3832 |
|  | Gambusia affinis | 2 | 0 | 0 | 0 | 0 | 34 | 36 |
|  | Ictalurus furcatus | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
|  | Ictalurus punctatus | 0 | 0 | 9 | 1 | 1 | 0 | 11 |
|  | Ictiobus bubalus | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Lepisosteus osseus | 0 | 0 | 0 | 0 | 0 | 7 | 7 |
|  | Lepisosteus platostomus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Lepomis humilis | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
|  | Macrhybopsis hyostoma | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Notropis stramineus | 5 | 1 | 0 | 0 | 0 | 0 | 6 |
|  | Pimephales vigilax | 234 | 26 | 0 | 0 | 0 | 27 | 287 |
|  | Polydictis olivaris | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| PAU | Carpiodes carpio | 4 | 0 | 0 | 2 | 0 | 0 | 6 |
|  | Cyprinella lutrensis | 1677 | 530 | 0 | 0 | 0 | 1197 | 3404 |
|  | Fundulus zebrinus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Gambusia affinis | 0 | 0 | 0 | 0 | 0 | 116 | 116 |
|  | Ictalurus furcatus | 0 | 0 | 1 | 1 | 0 | 0 | 2 |
|  | Ictalurus punctatus | 2 | 3 | 6 | 5 | 0 | 2 | 18 |
|  | Ictiobus bubalus | 11 | 0 | 1 | 1 | 0 | 25 | 38 |
|  | Lepisosteus osseus | 0 | 0 | 0 | 2 | 0 | 0 | 2 |
|  | Lepisosteus platostomus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Lepomis cyanellus | 1 | 0 | 0 | 0 | 0 | 4 | 5 |
|  | Macrhybopsis hyostoma | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Notropis stramineus | 1 | 1 | 0 | 0 | 0 | 0 | 2 |
|  | Percina sciera | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Pimephales vigilax | 248 | 7 | 0 | 0 | 0 | 127 | 382 |
|  | Polydictis olivaris | 0 | 0 | 1 | 5 | 0 | 0 | 6 |
| DUR | Aplodinotus grunniens | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Campostoma anomalum | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Cyprinus carpio | 0 | 0 | 0 | 1 | 1 | 0 | 2 |
|  | Ictalurus punctatus | 0 | 0 | 1 | 0 | 1 | 5 | 7 |
|  | Ictiobus cyprinellus | 1 | 0 | 0 | 0 | 1 | 2 | 4 |
|  | Notropis buchanani | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Percina sciera | 3 | 0 | 0 | 0 | 0 | 0 | 3 |
|  | Pimephales vigilax | 56 | 0 | 0 | 0 | 0 | 22 | 78 |
|  | Polydictis olivaris | 0 | 0 | 0 | 0 | 1 | 1 | 2 |

Appendix--A (continued)

| River/Site | Species | SS | SF | DS | DF | NW | BW | Total |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Cyprinella lutrensis | 366 | 25 | 0 | 0 | 0 | 252 | 643 |
|  | Dorosoma cepedianum | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Gambusia affinis | 9 | 0 | 0 | 0 | 0 | 347 | 356 |
|  | Hiodon alosoides | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Ictiobus bubalus | 17 | 0 | 2 | 0 | 1 | 9 | 29 |
|  | Lepisosteus platostomus | 0 | 0 | 0 | 0 | 0 | 8 | 8 |
|  | Macrhybopsis hyostoma | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Menidia beryllina | 7 | 0 | 0 | 0 | 0 | 0 | 7 |
|  | Micropterus puntulatus | 3 | 0 | 0 | 0 | 0 | 2 | 5 |
|  | Morone saxatilis | 3 | 0 | 0 | 0 | 0 | 0 | 3 |
|  | Moxostoma erythrurum | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Red River |  |  |  |  |  |  |  |  |
| WAU | Aplodinotus grunniens | 0 | 0 | 0 | 1 | 0 | 1 | 2 |
|  | Cyprinella lutrensis | 88 | 280 | 0 | 0 | 0 | 81 | 449 |
|  | Cyprinodon rubrofluviatilis | 0 | 0 | 0 | 0 | 0 | 17 | 17 |
|  | Cyprinus carpio | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Fundulus zebrinus | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Gambusia affinis | 3 | 2 | 0 | 0 | 0 | 150 | 155 |
|  | Hybognathus placitus | 48 | 229 | 0 | 0 | 0 | 2788 | 3065 |
|  | Ictalurus punctatus | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Ictiobus bubalus | 1 | 0 | 0 | 1 | 0 | 0 | 2 |
|  | Lepisosteus osseus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Lepisosteus platostomus | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Lepomis gulosus | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
|  | Macrhybopsis australis | 9 | 24 | 0 | 0 | 0 | 3 | 36 |
|  | Macrhybopsis hyostoma | 5 | 5 | 0 | 0 | 0 | 0 | 10 |
|  | Menidia beryllina | 1 | 1 | 0 | 0 | 0 | 3 | 5 |
|  | Morone chrysops | 0 | 0 | 0 | 0 | 0 | 2 | 2 |
|  | Notropis atherinoides | 13 | 43 | 0 | 0 | 0 | 6 | 62 |
|  | Notropis bairdi | 13 | 27 | 0 | 0 | 0 | 26 | 66 |
|  | Notropis buchanani | 0 | 3 | 0 | 0 | 0 | 17 | 20 |
|  | Pimephales vigilax | 1 | 2 | 0 | 0 | 0 | 30 | 33 |
|  | Polydictis olivaris | 0 | 0 | 0 | 3 | 0 | 0 | 3 |
|  | Pomoxis annularis | 0 | 0 | 0 | 0 | 0 | 4 | 4 |
| THA | Aplodinotus grunniens | 0 | 0 | 0 | , | 0 | 0 | 1 |
|  | Cyprinus carpio | 0 | 0 | 0 | , | 0 | 0 | 1 |
|  | Dorosoma cepedianum | 2 | 2 | 0 | 0 | 0 | 0 | 4 |
|  | Hybognathus placitus | 136 | 11 | 0 | 0 | 0 | 0 | 147 |
|  | Ictalurus punctatus | 0 | 1 | 0 | , | 0 | 0 | 2 |
|  | Lepomis cyanellus | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Macrhybopsis australis | 2 | 2 | 0 | 0 | 0 | 0 | 4 |
|  | Menidia beryllina | 2 | 0 | 0 | 0 | 0 | 0 | 2 |
|  | Notropis atherinoides | 127 | 37 | 0 | 0 | 0 | 0 | 164 |
|  | Pimephales vigilax | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
|  | Polydictis olivaris | 0 | 0 | 0 | 8 | 0 | 0 | 8 |

## Appendix--A (continued)

| River/Site | Species | SF | DS | DF | NW | BW | Total |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Cyprinella lutrensis | 158 | 11 | 0 | 0 | 0 | 0 | 169 |
|  | Dorosoma pretense | 13 | 0 | 0 | 0 | 0 | 0 | 13 |
|  | Gambusia affinis | 112 | 1 | 0 | 0 | 0 | 4 | 117 |
|  | Ictiobus bubalus | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
|  | Lepisosteus platostomus | 0 | 1 | 0 | 1 | 0 | 0 | 2 |
|  | Macrhybopsis hyostoma | 0 | 6 | 0 | 0 | 0 | 0 | 6 |
|  | Notropis bairdi | 17 | 1 | 0 | 0 | 0 | 0 | 18 |
|  | Notropis blennius | 3 | 0 | 0 | 0 | 0 | 0 | 3 |
|  | Pomoxis annularis | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
|  |  |  |  |  |  |  |  |  |
|  | Ammocrypta clara | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
|  | Aplodinotus grunniens | 0 | 0 | 0 | 0 | 1 | 1 | 2 |
|  | Carpiodes carpio | 0 | 1 | 1 | 0 | 2 | 0 | 4 |
|  | Cyprinella lutrensis | 113 | 65 | 0 | 0 | 0 | 53 | 231 |
|  | Cyprinodon rubrofluviatilis | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Dorosoma cepedianum | 17 | 9 | 0 | 0 | 0 | 5 | 31 |
|  | Hybognathus placitus | 0 | 6 | 0 | 0 | 0 | 0 | 6 |
|  | Ictalurus furcatus | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Ictalurus punctatus | 1 | 1 | 0 | 0 | 0 | 0 | 2 |
|  | Ictiobus bubalus | 0 | 1 | 1 | 0 | 0 | 2 | 4 |
|  | Lepisosteus osseus | 0 | 0 | 0 | 0 | 1 | 2 | 3 |
|  | Lepisosteus platostomus | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Lepomis cyanellus | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Macrhybopsis hyostoma | 0 | 2 | 0 | 0 | 0 | 0 | 2 |
|  | Menidia beryllina | 3 | 4 | 0 | 0 | 0 | 1 | 8 |
|  | Micropterus puntulatus | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
|  | Notropis atherinoides | 62 | 48 | 0 | 0 | 0 | 10 | 120 |
|  | Notropis bairdi | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
|  | Notropis blennius | 0 | 3 | 0 | 0 | 0 | 0 | 3 |
|  | Notropis buchanani | 11 | 8 | 0 | 0 | 0 | 0 | 19 |
|  | Pimephales vigilax | 6 | 4 | 0 | 0 | 0 | 0 | 10 |
|  | Polydictis olivaris | 0 | 0 | 0 | 0 | 2 | 0 | 2 |
| Pomoxis annularis | 5 | 0 | 0 | 0 | 0 | 0 | 5 |  |
| Total | 15776 | 6667 | 126 | 121 | 33 | 22863 | 45586 |  |

Appendix B--Abundance and catch-per-unit-effort (CPUE) of fish species collected by seining and hoop netting at 15 sample sites on five large prairie rivers in Oklahoma. ND = no data.

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| ARKANSAS River |  |  |  |  |  |  |  |  |  |
| NEW | Carpiodes carpio | 0 | 0.000 | 0 | 0.000 | 1 | 0.200 | 1 | 0.200 |
|  | Cyprinella lutrensis | 1438 | 65.364 | 0 | 0.000 | 0 | 0.000 | 1438 | 65.364 |
|  | Cyprinus carpio | 0 | 0.000 | 3 | 0.600 | 2 | 0.400 | 5 | 1.000 |
|  | Fundulus zebrinus | 7 | 0.318 | 0 | 0.000 | 0 | 0.000 | 7 | 0.318 |
|  | Gambusia affinis | 125 | 5.682 | 0 | 0.000 | 0 | 0.000 | 125 | 5.682 |
|  | Ictalurus punctatus | 6 | 0.273 | 1 | 0.200 | 3 | 0.600 | 10 | 1.073 |
|  | Ictiobus bubalus | 32 | 1.455 | 1 | 0.200 | 3 | 0.600 | 36 | 2.255 |
|  | Lepisosteus oculatus | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Lepisosteus osseus | 0 | 0.000 | 12 | 2.400 | 0 | 0.000 | 12 | 2.400 |
|  | Lepisosteus platostomus | 0 | 0.000 | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 |
|  | Lepomis macrochirus | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Menidia beryllina | 1060 | 48.182 | 0 | 0.000 | 0 | 0.000 | 1060 | 48.182 |
|  | Micropterus puntulatus | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Minytrema melanops | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Morone chrysops | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | Morone saxatilis | 4 | 0.182 | 0 | 0.000 | 0 | 0.000 | 4 | 0.182 |
|  | Notropis atherinoides | 918 | 41.727 | 0 | 0.000 | 0 | 0.000 | 918 | 41.727 |
|  | Notropis bairdi | 2 | 0.091 | 0 | 0.000 | 0 | 0.000 | 2 | 0.091 |
|  | Notropis stramineus | 7 | 0.318 | 0 | 0.000 | 0 | 0.000 | 7 | 0.318 |
|  | Percina phoxocephala | 2 | 0.091 | 0 | 0.000 | 0 | 0.000 | 2 | 0.091 |
|  | Pimephales vigilax | 67 | 3.045 | 0 | 0.000 | 0 | 0.000 | 67 | 3.045 |
|  | Polydictis olivaris | 0 | 0.000 | 1 | 0.200 | 1 | 0.200 | 2 | 0.400 |

Appendix B--(continued)

9

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| RAL | Cyprinella lutrensis | 158 | 7.182 | 0 | 0.000 | 0 | 0.000 | 158 | 7.182 |
|  | Dorosoma cepedianum | 6 | 0.273 | 0 | 0.000 | 0 | 0.000 | 6 | 0.273 |
|  | Gambusia affinis | 1439 | 65.409 | 0 | 0.000 | 0 | 0.000 | 1439 | 65.409 |
|  | Ictalurus punctatus | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Ictiobus bubalus | 0 | 0.000 | 1 | 0.250 | 0 | 0.000 | 1 | 0.250 |
|  | Lepomis cyanellus | 16 | 0.727 | 0 | 0.000 | 0 | 0.000 | 16 | 0.727 |
|  | Lepomis macrochirus | 349 | 15.864 | 0 | 0.000 | 0 | 0.000 | 349 | 15.864 |
|  | Macrhybopsis hyostoma | 7 | 0.318 | 0 | 0.000 | 0 | 0.000 | 7 | 0.318 |
|  | Menidia beryllina | 12 | 0.545 | 0 | 0.000 | 0 | 0.000 | 12 | 0.545 |
|  | Micropterus puntulatus | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Notropis atherinoides | 849 | 38.591 | 0 | 0.000 | 0 | 0.000 | 849 | 38.591 |
|  | Phenacobius mirabilis | 1 | 0.045 | 0 | 0.000 | 0 | 0.000 | 1 | 0.045 |
|  | Pimephales vigilax | 13 | 0.591 | 0 | 0.000 | 0 | 0.000 | 13 | 0.591 |
| cow | Carpiodes carpio | 16 | 0.800 | 0 | 0.000 | 0 | 0.000 | 16 | 0.800 |
|  | Cyprinella lutrensis | 1605 | 80.250 | 0 | 0.000 | 0 | 0.000 | 1605 | 80.250 |
|  | Cyprinus carpio | 0 | 0.000 | 1 | 0.200 | 3 | 0.600 | 4 | 0.800 |
|  | Gambusia affinis | 36 | 1.800 | 0 | 0.000 | 0 | 0.000 | 36 | 1.800 |
|  | Ictalurus punctatus | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 | 4 | 0.800 |
|  | Ictiobus bubalus | 8 | 0.400 | 2 | 0.400 | 0 | 0.000 | 10 | 0.800 |
|  | Lepomis macrochirus | 0 | 0.000 | 0 | 0.000 | 7 | 1.400 | 7 | 1.400 |
|  | Lepomis megalotis | 0 | 0.000 | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 |
|  | Macrhybopsis hyostoma | 180 | 9.000 | 0 | 0.000 | 0 | 0.000 | 180 | 9.000 |
|  | Menidia beryllina | 40 | 2.000 | 0 | 0.000 | 0 | 0.000 | 40 | 2.000 |
|  | Micropterus puntulatus | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | Morone saxatilis | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Notropis atherinoides | 341 | 17.050 | 0 | 0.000 | 0 | 0.000 | 341 | 17.050 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
|  | Percina shumardi | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Phenacobius mirabilis | 11 | 0.550 | 0 | 0.000 | 0 | 0.000 | 11 | 0.550 |
|  | Pimephales vigilax | 56 | 2.800 | 0 | 0.000 | 0 | 0.000 | 56 | 2.800 |
| Cimarron R. |  |  |  |  |  |  |  |  |  |
| DOV | Aplodinotus grunniens | 0 | 0.000 | ND | ND | 2 | 0.333 | 2 | 0.333 |
|  | Carpiodes carpio | 4 | 0.235 | ND | ND | 0 | 0.000 | 4 | 0.235 |
|  | Cyprinella lutrensis | 151 | 8.882 | ND | ND | 0 | 0.000 | 151 | 8.882 |
|  | C. rubrofluviatilis | 8 | 0.471 | ND | ND | 0 | 0.000 | 8 | 0.471 |
|  | Dorosoma cepedianum | 8 | 0.471 | ND | ND | 0 | 0.000 | 8 | 0.471 |
|  | Fundulus zebrinus | 6 | 0.353 | ND | ND | 0 | 0.000 | 6 | 0.353 |
|  | Gambusia affinis | 166 | 9.765 | ND | ND | 0 | 0.000 | 166 | 9.765 |
|  | Hybognathus placitus | 27 | 1.588 | ND | ND | 0 | 0.000 | 27 | 1.588 |
|  | Ictalurus punctatus | 0 | 0.000 | ND | ND | 1 | 0.167 | 1 | 0.167 |
|  | Lepomis cyanellus | 2 | 0.118 | ND | ND | 0 | 0.000 | 2 | 0.118 |
|  | Micropterus puntulatus | 12 | 0.706 | ND | ND | 0 | 0.000 | 12 | 0.706 |
|  | Notropis atherinoides | 844 | 49.647 | ND | ND | 0 | 0.000 | 844 | 49.647 |
|  | Phenacobius mirabilis | 1 | 0.059 | ND | ND | 0 | 0.000 | 1 | 0.059 |
|  | Polydictis olivaris | 0 | 0.000 | ND | ND | 1 | 0.167 | 1 | 0.167 |
| GUT | Carpiodes carpio | 127 | 6.048 | ND | ND | 9 | 1.800 | 136 | 7.848 |
|  | Cyprinella lutrensis | 71 | 3.381 | ND | ND | 0 | 0.000 | 71 | 3.381 |
|  | C. rubrofluviatilis | 8 | 0.381 | ND | ND | 0 | 0.000 | 8 | 0.381 |
|  | Cyprinus carpio | 1 | 0.048 | ND | ND | 0 | 0.000 | 1 | 0.048 |
|  | Dorosoma cepedianum | 2 | 0.095 | ND | ND | 0 | 0.000 | 2 | 0.095 |
|  | Fundulus zebrinus | 3 | 0.143 | ND | ND | 0 | 0.000 | 3 | 0.143 |
|  | Gambusia affinis | 42 | 2.000 | ND | ND | 0 | 0.000 | 42 | 2.000 |
|  | Hybognathus placitus | 643 | 30.619 | ND | ND | 0 | 0.000 | 643 | 30.619 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| COY | Lepisosteus oculatus | 5 | 0.238 | ND | ND | 0 | 0.000 | 5 | 0.238 |
|  | Lepomis humilis | 1 | 0.048 | ND | ND | 0 | 0.000 | 1 | 0.048 |
|  | Morone chrysops | 1 | 0.048 | ND | ND | 0 | 0.000 | 1 | 0.048 |
|  | Notropis atherinoides | 164 | 7.810 | ND | ND | 0 | 0.000 | 164 | 7.810 |
|  | Pomoxis annularis | 2 | 0.095 | ND | ND | 0 | 0.000 | 2 | 0.095 |
|  | Aplodinotus grunniens | 0 | 0.000 | 3 | 0.500 | 2 | 0.333 | 5 | 0.833 |
|  | Carpiodes carpio | 40 | 1.739 | 2 | 0.333 | 1 | 0.167 | 43 | 2.239 |
|  | Cyprinella lutrensis | 43 | 1.870 | 0 | 0.000 | 0 | 0.000 | 43 | 1.870 |
|  | C. rubrofluviatilis | 2 | 0.087 | 0 | 0.000 | 0 | 0.000 | 2 | 0.087 |
|  | Cyprinus carpio | 0 | 0.000 | 3 | 0.500 | 0 | 0.000 | 3 | 0.500 |
|  | Dorosoma cepedianum | 26 | 1.130 | 0 | 0.000 | 0 | 0.000 | 26 | 1.130 |
|  | Fundulus zebrinus | 81 | 3.522 | 0 | 0.000 | 0 | 0.000 | 81 | 3.522 |
|  | Gambusia affinis | 98 | 4.261 | 0 | 0.000 | 0 | 0.000 | 98 | 4.261 |
|  | Hybognathus placitus | 1464 | 63.652 | 0 | 0.000 | 0 | 0.000 | 1464 | 63.652 |
|  | Macrhybopsis hyostoma | 1 | 0.043 | 0 | 0.000 | 0 | 0.000 | 1 | 0.043 |
|  | Ictalurus punctatus | 0 | 0.000 | 0 | 0.000 | 8 | 1.333 | 8 | 1.333 |
|  | Ictiobus bubalus | 0 | 0.000 | 3 | 0.500 | 0 | 0.000 | 3 | 0.500 |
|  | Lepisosteus osseus | 1 | 0.043 | 1 | 0.167 | 0 | 0.000 | 2 | 0.210 |
|  | Lepomis macrochirus | 4 | 0.174 | 0 | 0.000 | 1 | 0.167 | 5 | 0.341 |
|  | Morone chrysops | 0 | 0.000 | 0 | 0.000 | 4 | 0.667 | 4 | 0.667 |
|  | Notropis atherinoides | 484 | 21.043 | 0 | 0.000 | 0 | 0.000 | 484 | 21.043 |
|  | Notropis bairdi | 153 | 6.652 | 0 | 0.000 | 0 | 0.000 | 153 | 6.652 |
|  | Pimephales vigilax | 3 | 0.130 | 0 | 0.000 | 0 | 0.000 | 3 | 0.130 |
|  | Polydictis olivaris | 0 | 0.000 | 1 | 0.167 | 3 | 0.500 | 4 | 0.667 |
| N. Canadian River |  |  |  |  |  |  |  |  |  |
| ELR | Cyprinella lutrensis | 2673 | 121.500 | ND | ND | 0 | 0.000 | 2673 | 121.500 |
|  | Cyprinus carpio | 0 | 0.000 | ND | ND | 1 | 0.167 | 1 | 0.167 |
|  | Fundulus zebrinus | 7 | 0.318 | ND | ND | 0 | 0.000 | 7 | 0.318 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
|  | Gambusia affinis | 4 | 0.182 | ND | ND | 0 | 0.000 | 4 | 0.182 |
|  | Ictalurus punctatus | 0 | 0.000 | ND | ND | 19 | 3.167 | 19 | 3.167 |
|  | Lepomis cyanellus | 2 | 0.091 | ND | ND | 0 | 0.000 | 2 | 0.091 |
|  | Lepomis humilis | 2 | 0.091 | ND | ND | 0 | 0.000 | 2 | 0.091 |
|  | Lepomis megalotis | 0 | 0.000 | ND | ND | 19 | 3.167 | 19 | 3.167 |
|  | Notropis bairdi | 1 | 0.045 | ND | ND | 0 | 0.000 | 1 | 0.045 |
|  | Notropis blennius | 1 | 0.045 | ND | ND | 0 | 0.000 | 1 | 0.045 |
|  | Notropis stramineus | 71 | 3.227 | ND | ND | 0 | 0.000 | 71 | 3.227 |
|  | Phenacobius mirabilis | 5 | 0.227 | ND | ND | 0 | 0.000 | 5 | 0.227 |
|  | Pimephales vigilax | 20 | 0.909 | ND | ND | 0 | 0.000 | 20 | 0.909 |
|  | Polydictis olivaris | 0 | 0.000 | ND | ND | 8 | 1.333 | 8 | 1.333 |
|  | Sander vitreus | 3 | 0.136 | ND | ND | 0 | 0.000 | 3 | 0.136 |
| HAR | Carpiodes carpio | 0 | 0.000 | 8 | 1.600 | 0 | 0.000 | 8 | 1.600 |
|  | Cyprinella lutrensis | 722 | 36.100 | 0 | 0.000 | 0 | 0.000 | 722 | 36.100 |
|  | Cyprinus carpio | 1 | 0.050 | 7 | 1.400 | 0 | 0.000 | 8 | 1.450 |
|  | Dorosoma pretense | 24 | 1.200 | 0 | 0.000 | 0 | 0.000 | 24 | 1.200 |
|  | Gambusia affinis | 95 | 4.750 | 0 | 0.000 | 0 | 0.000 | 95 | 4.750 |
|  | Hybognathus placitus | 3 | 0.150 | 0 | 0.000 | 0 | 0.000 | 3 | 0.150 |
|  | Ictalurus furcatus | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | Ictalurus punctatus | 0 | 0.000 | 3 | 0.600 | 1 | 0.167 | 4 | 0.767 |
|  | Ictiobus bubalus | 1 | 0.050 | 11 | 2.200 | 1 | 0.167 | 13 | 2.417 |
|  | Ictiobus cyprinellus | 5 | 0.250 | 0 | 0.000 | 0 | 0.000 | 5 | 0.250 |
|  | Lepisosteus osseus | 0 | 0.000 | 0 | 0.000 | 1 | 0.167 | 1 | 0.167 |
|  | L. platostomus | 2 | 0.100 | 0 | 0.000 | 1 | 0.167 | 3 | 0.267 |
|  | Lepomis macrochirus | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | Lepomis megalotis | 0 | 0.000 | 0 | 0.000 | 1 | 0.167 | 1 | 0.167 |
|  | Morone chrysops | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Notropis buchanani | 14 | 0.700 | 0 | 0.000 | 0 | 0.000 | 14 | 0.700 |
|  | Notropis stramineus | 10 | 0.500 | 0 | 0.000 | 0 | 0.000 | 10 | 0.500 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| WET | Pimephales vigilax | 53 | 2.650 | 0 | 0.000 | 0 | 0.000 | 53 | 2.650 |
|  | Polydictis olivaris | 0 | 0.000 | 9 | 1.800 | 2 | 0.333 | 11 | 2.133 |
|  | Pomoxis annularis | 4 | 0.200 | 0 | 0.000 | 0 | 0.000 | 4 | 0.200 |
|  | Aplodinotus grunniens | 5 | 0.238 | 0 | 0.000 | 0 | 0.000 | 5 | 0.238 |
|  | Carpiodes carpio | 8 | 0.381 | 0 | 0.000 | 0 | 0.000 | 8 | 0.381 |
|  | Cyprinella lutrensis | 12835 | 611.190 | 0 | 0.000 | 0 | 0.000 | 12835 | 611.190 |
|  | Cyprinus carpio | 0 | 0.000 | 0 | 0.000 | 1 | 0.167 | 1 | 0.167 |
|  | Gambusia affinis | 62 | 2.952 | 0 | 0.000 | 0 | 0.000 | 62 | 2.952 |
|  | Ictalurus punctatus | 1 | 0.048 | 1 | 0.250 | 4 | 0.667 | 6 | 0.964 |
|  | Ictiobus bubalus | 0 | 0.000 | 2 | 0.500 | 1 | 0.167 | 3 | 0.667 |
|  | L. platostomus | 0 | 0.000 | 0 | 0.000 | 1 | 0.167 | 1 | 0.167 |
|  | Lepomis cyanellus | 1 | 0.048 | 0 | 0.000 | 0 | 0.000 | 1 | 0.048 |
|  | Lepomis humilis | 3 | 0.143 | 0 | 0.000 | 0 | 0.000 | 3 | 0.143 |
|  | Notropis atherinoides | 5 | 0.238 | 0 | 0.000 | 0 | 0.000 | 5 | 0.238 |
|  | Pimephales vigilax | 682 | 32.476 | 0 | 0.000 | 0 | 0.000 | 682 | 32.476 |
|  | Polydictis olivaris | 0 | 0.000 | 0 | 0.000 | 3 | 0.500 | 3 | 0.500 |
|  | Pomoxis annularis | 6 | 0.286 | 0 | 0.000 | 1 | 0.167 | 7 | 0.452 |
| Washita River |  |  |  |  |  |  |  |  |  |
| ALE | Carpiodes carpio | 130 | 5.909 | ND | ND | 0 | 0.000 | 130 | 5.909 |
|  | Cyprinella lutrensis | 3832 | 174.182 | ND | ND | 0 | 0.000 | 3832 | 174.182 |
|  | Gambusia affinis | 36 | 1.636 | ND | ND | 0 | 0.000 | 36 | 1.636 |
|  | Ictalurus furcatus | 0 | 0.000 | ND | ND | 1 | 0.167 | 1 | 0.167 |
|  | Ictalurus punctatus | 0 | 0.000 | ND | ND | 11 | 1.833 | 11 | 1.833 |
|  | Ictiobus bubalus | 1 | 0.045 | ND | ND | 0 | 0.000 | 1 | 0.045 |
|  | Lepisosteus osseus | 7 | 0.318 | ND | ND | 0 | 0.000 | 7 | 0.318 |
|  | L. platostomus | 1 | 0.045 | ND | ND | 0 | 0.000 | 1 | 0.045 |
|  | Lepomis humilis | 0 | 0.000 | ND | ND | 1 | 0.167 | 1 | 0.167 |
|  | Macrhybopsis hyostoma | 2 | 0.091 | ND | ND | 0 | 0.000 | 2 | 0.091 |
|  | Notropis stramineus | 6 | 0.273 | ND | ND | 0 | 0.000 | 6 | 0.273 |

Appendix B--(continued)


Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
|  | Ictiobus cyprinellus | 3 | 0.150 | 1 | 0.167 | 0 | 0.000 | 4 | 0.317 |
|  | L. platostomus | 8 | 0.400 | 0 | 0.000 | 0 | 0.000 | 8 | 0.400 |
|  | Macrhybopsis hyostoma | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Menidia beryllina | 7 | 0.350 | 0 | 0.000 | 0 | 0.000 | 7 | 0.350 |
|  | Micropterus puntulatus | 5 | 0.250 | 0 | 0.000 | 0 | 0.000 | 5 | 0.250 |
|  | Morone saxatilis | 3 | 0.150 | 0 | 0.000 | 0 | 0.000 | 3 | 0.150 |
|  | Moxostoma erythrurum | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Notropis buchanani | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Percina sciera | 3 | 0.150 | 0 | 0.000 | 0 | 0.000 | 3 | 0.150 |
|  | Pimephales vigilax | 78 | 3.900 | 0 | 0.000 | 0 | 0.000 | 78 | 3.900 |
|  | Polydictis olivaris | 0 | 0.000 | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 |
| Red River |  |  |  |  |  |  |  |  |  |
| WAU | Aplodinotus grunniens | 1 | 0.040 | 1 | 0.333 | 0 | 0.000 | 2 | 0.373 |
|  | Cyprinella lutrensis | 449 | 17.960 | 0 | 0.000 | 0 | 0.000 | 449 | 17.960 |
|  | C. rubrofluviatilis | 17 | 0.680 | 0 | 0.000 | 0 | 0.000 | 17 | 0.680 |
|  | Cyprinus carpio | 0 | 0.000 | 0 | 0.000 | 1 | 0.250 | 1 | 0.250 |
|  | Dorosoma cepedianum | 2 | 0.080 | 0 | 0.000 | 0 | 0.000 | 2 | 0.080 |
|  | Fundulus zebrinus | 2 | 0.080 | 0 | 0.000 | 0 | 0.000 | 2 | 0.080 |
|  | Gambusia affinis | 155 | 6.200 | 0 | 0.000 | 0 | 0.000 | 155 | 6.200 |
|  | Hybognathus placitus | 3065 | 122.600 | 0 | 0.000 | 0 | 0.000 | 3065 | 122.600 |
|  | Ictalurus punctatus | 0 | 0.000 | 0 | 0.000 | 1 | 0.250 | 1 | 0.250 |
|  | Ictiobus bubalus | 1 | 0.040 | 1 | 0.333 | 0 | 0.000 | 2 | 0.373 |
|  | Lepisosteus osseus | 1 | 0.040 | 0 | 0.000 | 0 | 0.000 | 1 | 0.040 |
|  | L. platostomus | 1 | 0.040 | 0 | 0.000 | 1 | 0.250 | 2 | 0.290 |
|  | Lepomis gulosus | 1 | 0.040 | 0 | 0.000 | 0 | 0.000 | 1 | 0.040 |
|  | Macrhybopsis australis | 36 | 1.440 | 0 | 0.000 | 0 | 0.000 | 36 | 1.440 |
|  | Macrhybopsis hyostoma | 10 | 0.400 | 0 | 0.000 | 0 | 0.000 | 10 | 0.400 |
|  | Menidia beryllina | 5 | 0.200 | 0 | 0.000 | 0 | 0.000 | 5 | 0.200 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| THA | Morone chrysops | 2 | 0.080 | 0 | 0.000 | 0 | 0.000 | 2 | 0.080 |
|  | Notropis atherinoides | 62 | 2.480 | 0 | 0.000 | 0 | 0.000 | 62 | 2.480 |
|  | Notropis bairdi | 66 | 2.640 | 0 | 0.000 | 0 | 0.000 | 66 | 2.640 |
|  | Notropis buchanani | 20 | 0.800 | 0 | 0.000 | 0 | 0.000 | 20 | 0.800 |
|  | Pimephales vigilax | 33 | 1.320 | 0 | 0.000 | 0 | 0.000 | 33 | 1.320 |
|  | Polydictis olivaris | 0 | 0.000 | 0 | 0.000 | 3 | 0.750 | 3 | 0.750 |
|  | Pomoxis annularis | 4 | 0.160 | 0 | 0.000 | 0 | 0.000 | 4 | 0.160 |
|  | Aplodinotus grunniens | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | Cyprinella lutrensis | 169 | 10.563 | 0 | 0.000 | 0 | 0.000 | 169 | 10.563 |
|  | Cyprinus carpio | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | Dorosoma cepedianum | 4 | 0.250 | 0 | 0.000 | 0 | 0.000 | 4 | 0.250 |
|  | Dorosoma pretense | 13 | 0.813 | 0 | 0.000 | 0 | 0.000 | 13 | 0.813 |
|  | Gambusia affinis | 117 | 7.313 | 0 | 0.000 | 0 | 0.000 | 117 | 7.313 |
|  | Hybognathus placitus | 147 | 9.188 | 0 | 0.000 | 0 | 0.000 | 147 | 9.188 |
|  | Ictalurus punctatus | 1 | 0.063 | 0 | 0.000 | 1 | 0.250 | 2 | 0.313 |
|  | Ictiobus bubalus | 0 | 0.000 | 1 | 0.200 | 0 | 0.000 | 1 | 0.200 |
|  | L. platostomus | 1 | 0.063 | 0 | 0.000 | 1 | 0.250 | 2 | 0.313 |
|  | Lepomis cyanellus | 1 | 0.063 | 0 | 0.000 | 0 | 0.000 | 1 | 0.063 |
|  | M. australis | 4 | 0.250 | 0 | 0.000 | 0 | 0.000 | 4 | 0.250 |
|  | M. hyostoma | 6 | 0.375 | 0 | 0.000 | 0 | 0.000 | 6 | 0.375 |
|  | Menidia beryllina | 2 | 0.125 | 0 | 0.000 | 0 | 0.000 | 2 | 0.125 |
|  | Notropis atherinoides | 164 | 10.250 | 0 | 0.000 | 0 | 0.000 | 164 | 10.250 |
|  | Notropis bairdi | 18 | 1.125 | 0 | 0.000 | 0 | 0.000 | 18 | 1.125 |
|  | Notropis blennius | 3 | 0.188 | 0 | 0.000 | 0 | 0.000 | 3 | 0.188 |
|  | Pimephales vigilax | 2 | 0.125 | 0 | 0.000 | 0 | 0.000 | 2 | 0.125 |
|  | Polydictis olivaris | 0 | 0.000 | 5 | 1.000 | 3 | 0.750 | 8 | 1.750 |
|  | Pomoxis annularis | 2 | 0.125 | 0 | 0.000 | 0 | 0.000 | 2 | 0.125 |

Appendix B--(continued)

| River/Site | Species | Seine |  | Hoop Net |  |  |  | Total Abundance | Total CPUE |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | Large |  | Small |  |  |  |
|  |  | Abundance | CPUE | Abundance | CPUE | Abundance | CPUE |  |  |
| ART | Ammocrypta clara | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Aplodinotus grunniens | 0 | 0.000 | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 |
|  | Carpiodes carpio | 1 | 0.050 | 1 | 0.333 | 2 | 0.400 | 4 | 0.783 |
|  | Cyprinella lutrensis | 231 | 11.550 | 0 | 0.000 | 0 | 0.000 | 231 | 11.550 |
|  | C. rubrofluviatilis | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | D. cepedianum | 31 | 1.550 | 0 | 0.000 | 0 | 0.000 | 31 | 1.550 |
|  | Hybognathus placitus | 6 | 0.300 | 0 | 0.000 | 0 | 0.000 | 6 | 0.300 |
|  | Ictalurus furcatus | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | Ictalurus punctatus | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | Ictiobus bubalus | 1 | 0.050 | 2 | 0.667 | 1 | 0.200 | 4 | 0.917 |
|  | Lepisosteus osseus | 2 | 0.100 | 0 | 0.000 | 1 | 0.200 | 3 | 0.300 |
|  | L. platostomus | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Lepomis cyanellus | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | M. hyostoma | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | Menidia beryllina | 8 | 0.400 | 0 | 0.000 | 0 | 0.000 | 8 | 0.400 |
|  | M. puntulatus | 1 | 0.050 | 0 | 0.000 | 0 | 0.000 | 1 | 0.050 |
|  | Notropis atherinoides | 120 | 6.000 | 0 | 0.000 | 0 | 0.000 | 120 | 6.000 |
|  | Notropis bairdi | 2 | 0.100 | 0 | 0.000 | 0 | 0.000 | 2 | 0.100 |
|  | Notropis blennius | 3 | 0.150 | 0 | 0.000 | 0 | 0.000 | 3 | 0.150 |
|  | Notropis buchanani | 19 | 0.950 | 0 | 0.000 | 0 | 0.000 | 19 | 0.950 |
|  | Pimephales vigilax | 10 | 0.500 | 0 | 0.000 | 0 | 0.000 | 10 | 0.500 |
|  | Polydictis olivaris | 0 | 0.000 | 0 | 0.000 | 2 | 0.400 | 2 | 0.400 |
|  | Pomoxis annularis | 5 | 0.250 | 0 | 0.000 | 0 | 0.000 | 5 | 0.250 |
| Total |  | 47291 | 2255.486 | 136 | 24.256 | 206 | 33.294 | 47633 | 2313.036 |

VITA
Nicholas John Utrup
Candidate for the Degree of
Master of Science

## Thesis: DEVELOPMENT OF A RAPID BIOASSESSMENT PROTOCOL FOR SAMPLING FISHES IN LARGE PRAIRIE RIVERS

Major Field: Wildlife and Fisheries Ecology
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Personal Data: Born in Lima, Ohio, October 21, 1978, son of Gerald T. Utrup and Nina M. Nuveman.

Education: Graduated from Ottawa-Glandorf High School, Ottawa, Ohio in May 1997; received Bachelor of Science degree in Zoology from The Ohio State University in June 2002; completed the requirements for the Master of Science degree at Oklahoma State University in July 2005.

Experience: Assistant Fish Biologist Intern, Ohio Environmental Protection Agency, Division of Surface Water, Ecological Assessment Unit, June 1999-September 1999; Research Technician, The Ohio State University, Aquatic Ecology Laboratory, September 1999- June 2002; Teaching Assistant and Research Technician, The Ohio State University, Ohio Sea Grant, F.T. Stone Laboratory, June 2001August 2001; Graduate Research Assistant, Oklahoma Cooperative Fish and Wildlife Research Unit, July 2002-August 2004; Graduate Teaching Assistant, Oklahoma State University, Department of Zoology, August 2004-May 2005; Fishery Biologist, U.S. Fish and Wildlife Service, Columbia Fishery Resources Office, May 2005present.

Professional Memberships: American Fisheries Society; Oklahoma Chapter of the American Fisheries Society; Oklahoma Academy of Science.

## Major Field: Wildlife and Fisheries Ecology

Scope and Method of Study: Assessment of the biotic integrity of large rivers depends on efficient and statistically-sound procedures for sampling. Large prairie rivers in the Great Plains ecoregion pose a unique challenge for sampling because their physical and chemical properties limit the types of gear, such as electrofishing, that are traditionally used to sample fish for bioassessments. In this study, I used seining and hoop netting to collect fishes at 15 sites in five large prairie rivers in Oklahoma to (1) determine the minimum amount of effort needed to detect the majority of fish species at a sample site and (2) examine the selectivity of fish species detected by the two gear types. Based on my findings, I present recommendations for sampling fish assemblages in large prairie rivers.

Findings and Conclusions: Analysis of similarities of the fishes collected in six different habitat types identified two distinct habitat types based on fish species composition: shallow/backwater (SBW) habitat (depth $\leq 0.75 \mathrm{~m}$ ) and deep/non-wadeable (DNW) habitat (depth $>0.75 \mathrm{~m}$ ). I estimated that eight SBW habitats and four DNW habitats at each sample site were needed to detect a majority of fish species during a sampling event. Minimum sampling distance needed to encounter the required number of habitats ranged from 400 m to 1600 m and averaged 887 m . Gear evaluation showed seining captured more species-per-unit-effort than hoop netting (3.6 and 1.4 respectively); however, hoop nettingcaptured significantly larger fish (527 $\mathrm{mm} ; P<0.001$ ) than seining ( 42 mm ). Based on these results I recommend the following protocol for sampling fish assemblages in large prairie rivers: (1) use aerial photos to determine site accessibility and sampling distance prior to leaving for the field, (2)sample a minimum of eight randomly selecte d SBW habitats using a seine, and (3)set a minimum of two hoop nets per DNW habitat. With some modifications these recommendations can be applied to similar rivers throughout the Great Plains to aid in the rapid bioassessment and monitoring of fish assemblages in large prairie rivers.

